

Freeport-McMoRan Chino Mines Company

# Smelter/Tailing Soils Investigation Unit Feasibility Study

**Smelter Tailing Soils Investigation Unit  
Chino Mine Investigation Area, Grant County, New Mexico**

February 2025

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## Acronyms and Abbreviations

AOC	Administrative Order on Consent
ARAR	Applicable or Relevant and Appropriate Requirement
bgs	Below Ground Surface
BMP	Best Management Practices
CCP	Closure/Closeout Plan
CERCLA	Comprehensive Environmental Response, Compensation, and Liability Act
CFR	Code of Federal Regulations
CGCS	Comprehensive Groundwater Characterization Study
Chino	Freeport-McMoRan Chino Mines Company
COC	Constituent of Concern
COPC	Constituent of Potential Concern
csm	Conceptual Site Model
DEL	De minimis effects level
DOC	dissolved organic carbon
EC50	effects concentration at 50%
EPC	exposure point concentration
ERA	Ecological Risk Assessment
FS	Feasibility Study
FMRS	Freeport-McMoRan Reclamation Services
HHRA	Human Health Risk Assessment
HQ	hazard quotient
HSIU	Hurley Soils Investigation Unit
IA	Investigation Area
IRA	Interim Remedial Action
mg/kg	milligrams per kilogram
mm	millimeters
NCP	National Contingency Plan
NMAC	New Mexico Administrative Code
NMDOT	New Mexico Department of Transportation

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NMED	New Mexico Environmental Department
NMWQCC	New Mexico Water Quality Control Commission
NRCS	National Resource Conservation Survey
OAT	Observed Apparent Trend
O&M	Operation and Maintenance
pCu	Cupric ion activity (pCu <sup>2+</sup> )
PEL	Probable Effects Level
pH	Hydrogen ion (standard units)
PRG	Preliminary Remediation Goals
RAC	Remedial Action Criteria
RAO	Remedial Action Objective
RI	Remedial Investigation
ROD	Record of Decision
ROW	Right-of-Way
SGFB	Small ground feeding bird
Site	Smelter/Tailing Soils Investigation Unit at the Chino Mine Investigation Area, Grant County, New Mexico
SSC	Site-Specific Criteria
STSIU	Smelter/Tailing Soils Investigation Unit
SWQB	Surface Water Quality Bureau
TBC	To Be Considered
TP	Tailing pond
UAA	Use Attainability Analysis
UCL	Upper Confidence Limit
USEPA	United States Environmental Protection Agency
WBT	wind-blown tailings
WER	Water Effect Ratio
WQS	Water Quality Standard
XRF	x-ray fluorescence

# 1. Introduction

The Feasibility Study (FS) was prepared for Freeport-McMoRan Chino Mines Company (Chino) to develop and evaluate potential remedial alternatives for the Smelter Tailing Soils Investigation Unit (STSIU) at the Chino Mine Investigation Area, Grant County, New Mexico (site). This FS has been developed in accordance with the requirements in the Administrative Order on Consent (AOC) following the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA) guidance. The AOC, effective December 24, 1994, distinguishes between historical mineral processing activities and current operations at Chino.

## 1.1 Background

The STSIU is one of the Investigation Units identified within the Investigation Area (IA) of the AOC. The IA includes all areas in which environmental media may have been affected by historical operations at mining and processing facilities. The STSIU is located approximately 12 miles southeast of Silver City, and includes historical smelting facilities, mineral processing facilities, tailing ponds, and surrounding areas (Figure 1-1). The STSIU is located to the east of the town of Hurley, New Mexico which contained the Hurley Smelter, and has previously been defined as all areas containing and proximal to Chino's former copper smelter and ancillary facilities, including the tailings disposal facility (SRK 2008a).

In accordance with the AOC Scope of Work, a Remedial Investigation (RI) for the STSIU was conducted to generate the data necessary to evaluate the potential effects to human health and the environment from historically affected media in the STSIU. Data have been collected in the STSIU starting in 1995 and continuing to present day to determine potential impacts to soil, sediment, and surface water from historical mineral processing activities. The human health risk assessment (HHRA) and ecological risk assessment (ERA) have shown that areas of the STSIU have elevated metals concentrations and depressed pH in soil and surface water, as described in Sections 3.1.1 and 4.1.1. The FS Proposal for the STSIU presented activities necessary to evaluate remedial alternatives that comply with New Mexico Environment Department (NMED) Pre-FS remedial action criteria (RAC) (Arcadis 2011a).

NMED issued Pre-FS RAC for the STSIU (NMED 2010a, 2011a) including:

### **Soil**

- 27 milligrams per kilogram (mg/kg) arsenic (0-1")
- 100,000 mg/kg iron in soil (0-1")
- 5,000 mg/kg copper (0-1")
- 1,600 mg/kg copper (0-6")
- Cupric ion activity ( $\text{pCu}^{2+}$ ) (hereafter referred to as "pCu")  $\geq 5$  where copper  $> 327$  mg/kg. Note: Chino interprets this to actually mean NMED selected areas for remedial decisions in the STSIU where cupric ion activity ( $\text{pCu}^{2+}$ )  $< 5$  where copper is  $> 327$  mg/kg.

### **Surface Water**

- Water quality criteria contained in New Mexico Administrative Code (NMAC) §20.6.4.

These Pre-FS RAC values were developed based on the evaluations conducted in the RI, HHRA, and ERA. The FS and Record of Decision (ROD) will be completed consistent with the National Contingency Plan (NCP). Pre-FS RAC are consistent with the use of preliminary remediation goals (PRG) by United States Environmental Protection Agency (USEPA) in the NCP; therefore, new information can be used to refine the Pre-FS RAC and selection of alternatives (§300.430(e)(2)(i) NCP). Final remediation goals will be determined in the ROD. Further details about the Pre-FS RAC are presented in Section 2.4.

## 1.2 Objectives

The primary objectives of this FS, as initially presented in the FS Proposal (Arcadis 2011a), are to identify potential remedial areas and remedial technologies to address contaminated soil, sediment, and surface water in the STSIU. To achieve this, remedial action objectives (RAOs) were developed to define the basis for remediation, including numerical Pre-FS RAC as discussed in the previous section. Remedial technologies described in this document were assessed using the CERCLA FS criteria (Section 4.3, USEPA 1988) to determine their potential to meet the project RAOs.

As part of the FS Proposal, a comprehensive literature review was completed, and potential remedial technologies were identified for application at the STSIU (Arcadis 2011a). Remedial technology alternatives evaluated in this report are unchanged from the FS Proposal and analysed individually and in comparison with each other using the following criteria: overall protection of human and ecological receptors, compliance with applicable or relevant and appropriate requirements (ARARs); long-term effectiveness and permanence; reduction in toxicity, short-term effectiveness, implementability; and cost. The FS process followed to conduct the evaluation of alternatives presented in the FS includes the following steps:

- Summarize RAOs and Pre-FS RAC that address the key risk drivers and potential routes of exposure;
- Identify areas where potential remedial action(s) may be necessary to address RAOs and Pre-FS RAC;
- Identify and screen potential remedial technologies;
- Develop remedial alternatives; and
- Evaluate the remedial alternatives considering the FS criteria.

The above steps will be used to guide the selection of the preferred remedial alternatives.

## 1.3 Summary of Related Current Activities

Between the start of the AOC process in April 1995 and July 2022, there have been a number of site characterization or remediation activities in or near the STSIU at Chino. These include the following:

- AOC Interim Removal Action Work Plan and Completion Report for Interim Remedial Action, Hurley Soils Investigation Unit, (Golder 2006, 2008)
- Phase 1 Remedial Investigation Proposal, Hanover and Whitewater Creeks Investigation Unit (Golder 1999a)
- Comprehensive Groundwater Characterization Study (CGCS) (Golder 1999b)
- Phase II Revised Remedial Investigation Report, Hurley Soils Investigation Unit (Golder 2000a)

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- Phase 1 Revised Remedial Investigation Report, Hanover/Whitewater Creek Investigation Unit (Golder 2000b)
- Remedial Investigation Proposal, Smelter Tailings Soils Investigation Unit (SRK 2004)
- Sitewide Baseline Ecological Risk Assessment (NewFields 2006a)
- Slag Pile Characterization (Golder 2007)
- Remedial Investigation, Smelter Tailings Soils Investigation Unit (SRK 2008a,b)
- Smelter Tailings Soils Investigation Unit Ecological Risk Assessment (NewFields 2008)
- Smelter Tailings Soils Investigation Unit Human Health Risk Assessment (Gradient 2008)
- Smelter/Tailing Soils Investigation Unit Interim Removal Action (IRA) Completion Report (“Golf Course IRA” Arcadis 2009)
- STSIU Terrestrial Invertebrate Bioaccumulation and Bioavailability Study (Arcadis 2010a)
- Groundwater Quality Pre-FS RAC for Drainage Sediments Study (Arcadis 2011b)
- Chino Mines Tailing Pond 1 & 2 Construction Quality Assurance Report (Golder 2013a)
- Tailing Pond B/C Construction Quality Assurance Report (Golder 2013b)
- Supplemental Completion Report, Interim Removal Action, Smelter/Tailing Soils Investigation Unit (“Railroad IRA”, Golder 2013c).
- Development of Site-Specific Copper Criteria Interim Report (Arcadis 2013a,b).
- Construction Quality Assurance Report, Lake One Reclamation (EMC2 2014)
- Supplemental Completion Report, Razorback Ridge Area, Interim Remedial Action, Smelter/Tailing Soils Investigation Unit (Golder 2015)
- Year 5 Monitoring Report for STSIU Amendment Study Plots (Arcadis 2017) (Appendix A)
- Year 5 Report on pH Monitoring to Evaluate the Effect of the White Rain on STSIU (Arcadis 2023) (Appendix B)
- B-Ranch Interim Remedial Action Completion Report Smelter/Tailing Soils Investigation Unit (Arcadis 2021)
- Assessment Report for Apache Tejo Wash (Golder 2024)
- Phytotoxicity and Vegetation Community Study (Arcadis 2018, Appendix C)
- Reference Area Technical Memorandum (Arcadis 2024, Attachment A in Appendix D)

A brief overview of the above reports for 2011 and prior years was included in the approved FS Proposal (Arcadis 2011a). Several of the above reports have been updated or are new since the submittal of the FS Proposal in 2011. These reports are summarized below including reports that followed the FS Proposal.

***Interim Removal Action for STSIU - Golf Course and Railroad***

The *Interim Removal Action (IRA) for the STSIU* (Arcadis 2006, 2009) and *Supplemental IRA for the STSIU* (Golder 2013c) addressed elevated copper in surface soil to the north and west of Hurley, New Mexico. A former

golf course existed to the north of the Town of Hurley. The North and South Golf Course designation refers to this area; the division between the north and south is the access road to the Chino facility entrance in Hurley. Areas to the east and west of Highway 180 on the west side of the Golf Course were included (Arcadis 2009) as well as acres alongside the railroad (Golder 2013c)<sup>1</sup>.

The objective of the IRA was to remove areas where soil copper concentrations were greater than 5,000 mg/kg (lateral delineation). Chino applied the NMED-approved residential RAC for the Hurley Soils Investigation Unit (HSIU) as a conservative measure in 2008. STSIU Pre-FS RAC was defined by NMED in 2009. Within the area exceeding 5,000 mg/kg copper, the excavation was completed vertically until the soil did not exceed 2,700 mg/kg copper or bedrock was encountered and, as a result, the spatially weighted 95 percent upper confidence limit (95UCL) copper concentration in the 0-6 inch depth interval is estimated to be 1,314 mg/kg. Therefore, while this remedy was implemented prior to NMED issuing final pre-FS RAC, the soil removal meets the bird pre-FS RAC of 1,600 mg/kg.

Freeport-McMoRan Reclamation Services (FMRS) performed the IRA excavation activities in partnership with Arcadis, who conducted the field engineering services, which took place from January 2008 to August 2008. A supplemental IRA excavation of the remaining 32 acres between the town of Hurley and the golf course sites was conducted from August 2012 to December 2012 by TIPE Construction with oversight by Golder. Excavation depths during the IRA activities averaged three inches below ground surface (bgs) for the total removal area. Based on confirmation sampling and refinement of the removal areas, approximately 170 acres of the proposed 190 acres in 2008 and approximately 32 acres in 2012 were ultimately remediated for a total of 202 acres. Final excavation volumes were determined by contractor load counts. The total volume removed during these IRAs are estimated at 68,112 cubic yards (Arcadis 2009) and 22,125 cubic yards (Golder 2013c).

Following the completion of the IRAs in 2008 and 2012, five years of monitoring was completed for both sites to document the successful revegetation of the IRA footprint (Arcadis 2014c, Golder 2018). The results of the five-year monitoring effort demonstrated the successful development of an early-seral stage mixed grama herbaceous alliance across the Golf Course sites. The canopy cover levels exceed the reference area guidance for canopy cover, and the data indicated that native species colonized the site with minimal cover from non-native species. Further, the vegetation limited erosion and controlled dust.

#### ***Interim Removal Action for STSIU – Razorback Ridge and B-Ranch***

An IRA was completed east of Lake One in areas adjacent to the Whitewater Creek Diversion Channel and James Canyon in 2013 and 2014 in an area referred to as Razorback Ridge (Golder 2015). The objective of this IRA was to remove soils in areas with copper concentrations higher than 5,000 mg/kg, delineated laterally, as per the STSIU pre-FS RAC. Within the areas with higher than or equal to 5,000 mg/kg, soils were removed vertically to a depth of 1 to 2 feet as this surface soil volume was utilized as borrow fill material for reclamation. The final excavation area for the Razorback Ridge Area (which includes the East Removal Borrow Area) addressed in 2013 and 2014 was 94 acres. An additional 29 acres within the IRA Area were remediated through operational construction excavation and borrow activities prior to 2002. Soils were excavated to a depth of approximately 12 inches down to 40 feet during excavation for borrow material, exceeding the target remediation depths. See next

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<sup>1</sup>The Highway 180 right-of-way (ROW) extends 50 feet on either side of the roadway and is considered by the New Mexico Department of Transportation (NMDOT) to be under its management. Any impacted soils within the Highway ROW will be addressed by NMDOT as part of their planned expansion of Highway 180 from a 2-lane roadway to a 4-lane roadway for the entire corridor between Bayard and Deming (NMDOT 2021). Chino expects to coordinate with NMDOT prior to planned construction and anticipates proposing an option for disposal of copper contaminated soils in compliance with existing regulatory requirements.

bullet section below. In all areas of removal, the Pre-FS RAC for the STSIU of 5,000 mg/kg for copper was achieved (Golder 2015). Approximately 113,000 cubic yards of soil (73 acres) from Razorback Ridge and 34,000 cubic yards of soil (21 acres) from the East Removal Borrow Area were removed between 2013 and 2014 and used for borrow material, specific to the IRA requirements.

An additional IRA was completed in 2019 for the B-Ranch area located adjacent to the reclaimed Slag Pile, the Golf Course IRA sites, as well as the Town of Hurley. Approximately 16,000 cubic yards of soil (22 acres) from B-Ranch site were removed. In all areas of removal, the Pre-FS RAC for the STSIU of 5,000 mg/kg for copper was achieved (Arcadis 2021).

### ***Lake One and Tailing Pond Reclamation – Borrow Fill***

As part of the Closure/Closeout Plan (CCP) reclamation activities, Chino removed borrow fill, including wind-blown tailings (WBTs), from the east and west sides of the historic tailings dams in 2013 (Golder 2013a, 2013b, and 2013d) and 2014 (EMC2 2014). The extent of the borrow removal of more than 520 acres of land included the Razorback Ridge Borrow Area (87 acres), the South James Canyon Borrow Area (90 acres), Borrow Area A (71 acres), Borrow Area B (41 acres), Borrow Area C (50 acres), and Borrow D (153 acres). There was an additional 574 acres of WBT utilized as grading fill material where only the surface one foot depth was borrowed for Tailings Pond 1, 4, and 6. Borrow Areas A, B, and C, as well as Razorback Ridge, were used for fill and cover material for Tailings Pond 1, 2, B, C, 4, 6E and 6W and for Lake One. The top 2 feet of surface areas on the Razorback Ridge Areas and South James Canyon Borrow Area were removed and used as grading fill (i.e., below the base of the cover) within Lake One. Once placed, the WBT fill material also received a minimum 3-foot thickness of cover material along with the rest of the Lake One reclamation area. These borrow areas are addressed under the CCP.

### ***Site Specific Copper Criteria Study***

Arcadis conducted a copper water effect ratio (WER) study for the STSIU surface waters in 2011. As described in the WER work plan (Arcadis 2011c), the purpose of the WER study was to evaluate the site-specific toxicity of copper in STSIU surface waters and to determine whether the hardness-specific copper criteria accurately reflect site-specific copper bioavailability and toxicity. A total of 24 STSIU water samples were collected in the WER study, including 18 toxicity test samples and an additional 6 analytical chemistry samples.

Results from the WER study were described in the *Development of Site-Specific Copper Criteria Interim Report* submitted to NMED in March 2013 (Arcadis 2013a). The primary objectives of this Criteria Adjustment Interim report were to report all data collected as well as any deviations from the work plan, evaluate all collected data with quality control criteria described in the WER guidance (USEPA 1994, 2001), and determine if the collected data are sufficient to develop copper site-specific criteria (SSC) that can be applied to STSIU surface waters. Broadly, this study demonstrated that hardness-adjusted copper criteria are overly protective for surface waters of STSIU drainages, and the methods listed in Section 20.6.4.10, part D of NMAC can be successfully applied to surface water in STSIU drainages. The Interim WER Report also established that the toxicity and chemistry data collected were acceptable for deriving WERs when these data are compared to USEPA (1994 and 2001) WER acceptability criteria. Preliminary WERs were calculated according to USEPA guidance (USEPA 1994, 2001) and presented in that report. A site-specific copper WER model was subsequently developed to derive adjusted copper criteria in STSIU surface waters in the *Revised Site-Specific Copper Toxicity Model Report* submitted to NMED in October 2013 (Arcadis 2013b). The site-specific criteria for STSIU surface waters were adopted by NMED and are contained in NMAC §20.6.4.808 - 809.



### ***STSIU Amendment Study Plots***

Chino implemented the STSIU Amendment Study in 2008 based upon an NMED-approved work plan to explore the possible remedial options to determine if lime and organic matter amendments with or without tilling could be an effective and feasible remedial action to address the elevated copper concentrations and depressed pH in surface soils within the STSIU (Arcadis 2008a). The study tested the longevity of pH stabilization, copper sequestration ability, vegetative re-colonization, and constructability (Arcadis 2017, see Appendix A).

In June of 2008, Chino applied lime and organic matter to three test plots, and one test plot was used as a control with no amendments or tilling. A control plot with no treatment was also established adjacent to each of these four plots. During the application of lime and organic matter, two of the test plots on relatively level ground were also tilled. Chino conducted monitoring of the copper, pH, and vegetation cover, diversity, and richness before and after amendment application. The results from the monitoring events were presented in the *Year 5 Monitoring Report for Smelter/Tailings Soils Investigation Unit Amendment Study Plots* (Arcadis 2017). In summary, an unexpected white rain (discussed in next section) increased pH and pCu significantly and reduced copper uptake into plants of all the plots, which was sustained over the 5-year monitoring period. As a result, the treatments added to the soil in the plots provided minimal if any additional benefit in further reducing copper impacts to the plant community because the white rain already significantly improved the soil chemistry. The white rain also increased plant species richness on the untreated plots but had a minimal, if any, positive effect on total plant cover (decreased cover in some plots). The vegetation community of the steeper plot was degraded by the two amendments, with a loss of rangeland grasses and desirable shrubs. The tilling and mixing of lime amendments into the soil tended to increase pH but degraded the plant community on the fair rangeland plot and improved the plant community on the poor rangeland plot; however, the recovery of degraded conditions could take decades. Liming and tilling are only recommended in relatively flat, poor rangeland rocky areas where phytotoxicity from copper can be demonstrated. Tilling alone has been shown to be effective and should be tried first at an 8-inch depth and deeper to evaluate the proper depth but only on slopes < 13% and where the rangeland condition is poor. Organic matter application in the form of manure was not recommended.

### ***STSIU pH Monitoring***

A significant upward shift in pH was observed in the STSIU following a “white rain” precipitation event on January 7, 2008 (Arcadis 2008b). To quantify the permanence of this trend, Chino submitted, and NMED approved, the Soil pH Monitoring Work Plan in 2010 (Arcadis 2010b). The Work Plan outlined eighteen locations (increased to 22 in 2011) to monitor soil pH over 5 years<sup>2</sup>, with annual reports provided to NMED after each sampling event and a final, Year 5 report summarizing the results of the pH monitoring effort (Arcadis 2023, see Appendix B).

The results of the 5-year study showed that the 2008 white rain benefitted the STSIU soils by increasing the pH and pCu of the acidic soils, making copper less bioavailable likely due to the increase in copper adsorption by secondary soil minerals, such as iron hydroxide, at higher pH values (Arcadis 2023). The increase in pCu led to a decrease in the uptake of copper into living organisms. In locations showing an improvement (increase) in soil pH, the tissue copper concentrations decreased after the white rain by more than 60% for the plants and by up to 40% for the insects. The ultimate result appears to be reduced toxicity to wildlife due to reduction in copper in their diet. Thus, the white rain improved wildlife and rangeland habitat by decreasing copper in food sources (and improved vegetation richness, but not cover as discussed in Appendix A). Based on Mining and Minerals Division guidelines and mineralogical analysis, the potential of STSIU soils to generate acid is consistently low in most areas. Based

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<sup>2</sup> Soil excavations from reclamation borrow activities over the monitoring period removed five of the sample locations, so five additional locations were added in 2012. Overall, 17 locations were monitored all five years.

on results reported in the *Year 5 Report on pH Monitoring to Evaluate the Effect of the White Rain on the Smelter/Tailings Soils Investigation Unit* (Arcadis 2023), persistence in the future cannot be predicted with certainty, and additional five-year monitoring is recommended as part of the STSIU FS to confirm the prediction that the pH increase should be sustained. Appendix B includes the revised report plus a brief response to NMED comments (see Appendix I of the pH monitoring report for response to comments).

### ***Phytotoxicity and Vegetation Community Study***

The site-wide and STSIU-specific ERAs stated that elevated concentrations of copper and other metals, combined with depressed soil pH, have led to a risk of phytotoxicity for some areas of the Site based on site-specific greenhouse phytotoxicity studies and plant community surveys in 1999 (NewFields 2006a, NewFields 2008, Schafer and Associates 1999a,b). For plants, these ERAs linked the toxic action of total copper in the shallow surface layer of the soil to cupric ion activity, quantified as pCu. Site conditions changed since the original studies, however, requiring additional studies; specifically, the two historical smelter stacks were shut down in 2002 and demolished in 2007. Reclamation of Lake One and older tailing dams was also completed. Additionally, a significant upward shift in soil pH was observed at STSIU following the “white rain” precipitation event on January 7, 2008. The change in soil pH due to the white rain event has lowered cupric ion activity of the soil (Arcadis 2023) and, more importantly, possibly changed the complex soil geochemistry (e.g., cation exchange capacity) in a manner that could shift the relationship between pCu and plant and community endpoints. In addition, the 1999 phytotoxicity and community studies did not explicitly account for site-specific seeds of native species, representative background, or potentially confounding physical and chemical factors across various soil categories.

New studies were proposed by Chino in 2013 to address these changes, approved by NMED in 2014. Data collection was completed in 2015, results submitted in April 2017 and re-submitted in August 2017 for NMED review. The report in Appendix C was updated in 2018 and addresses the last set of NMED comments provided in January 2018; the responses to NMED comments are attached at the end of the report. The native site species did not always perform better than non-native species in the greenhouse study, and their addition produced a minimal change in the results relative to 1999 results. However, dose-response models for each category of soil (bedrock, steep slope, flat granular, flat rocky) produced a different de minimus effects level (DEL) or probable effects level (PEL) for toxicity of copper than the 1999 data based on the plant community data. A DEL was defined as the “lowest soil pCu level... below which statistically demonstrable effects are unlikely”. The PEL was defined as “pCu levels at which statistically demonstrable effects are probable (e.g., pCu at which approximately 50% of the test population is affected).” The variable DEL and PEL values, depending on soil conditions, create uncertainty around the application of a pre-FS RAC for vegetation across all soil types, and on whether more harm than good will result if simple cleanup criteria are used across the entire Investigation Unit. This study’s information was useful for the FS because it defines areas on the site by the four soil categories of bedrock, steep slope > 13%, flat granular, and flat rocky soils that strongly influence plant responses to pCu. The community study was based on site soils that were not homogenized in the laboratory (maintaining pCu stratification) and are the most representative of field conditions.

### ***Draft Technical Memorandum on pCu Reference Area Visit and Analysis***

Based on NMED comments regarding the phytotoxicity report and an August 2018 meeting between Chino and NMED, Chino proposed sampling additional locations for “reference” or “background” to increase confidence in conclusions. Reference areas were identified in the field with NMED, and Chino sampled the community metric endpoints of cover, richness, and rangeland condition (via Observed Apparent Trend [OAT] score) as well as soil chemistry (pH, total copper, and sulfate) across all four soil categories identified in the phytotoxicity study (Arcadis

2018): flat granular, flat rocky, bedrock, and steeper slopes (>13%). Eight reference areas were sampled for soil and surveyed for three plant endpoints (OAT score, richness, cover) in October 2018, bringing the total reference area dataset to be used for comparing plant communities on site to reference areas off site to ten, given that two reference sites (wildlife reference north and STS-PT-2013-26) had been sampled previously. Three additional STSIU onsite locations were sampled or surveyed. Two of these were sampled for soil chemistry and/or plant community data to evaluate pCu at highly overgrazed areas on the site, and one (wildlife reference south plot) was sampled and surveyed along with the wildlife reference north location to calibrate changes<sup>3</sup> in plant cover due to the climatic differences over 3 different years of sampling (2011, 2014, 2018); this repeat survey ensured that on-site and off-site area data are comparable.

Like the two previously sampled reference locations, all new reference locations had sulfate concentrations and pH in the soil within the expected range for the geology and soils of the location, indicating that they do not have any smelter or windblown tailing impacts. These results support that the new reference locations adequately represent background conditions without mining impacts. The new reference area data are **not** incorporated in the phytotoxicity study report, however, because it has been through several review cycles with NMED. Instead, data from the new reference locations were evaluated in a Technical Memorandum (Tech Memo) and in this FS as to how they affect DELs and PELs for pCu impacts on the vegetation community. The Tech Memo (Arcadis 2024) is provided as Attachment A to Appendix D and the final analysis and methods used to identify DELs and PELs in that Tech Memo using the new reference data is discussed in Appendix D. DELs and PELs presented in the community section of the phytotoxicity report were recalculated with the new data. The recalculation is important because the phytotoxicity study did not have any reference areas representative of flat rocky, bedrock, or slopes > 13%. Previous community PEL and DEL were based on the flat granular reference area, underestimating the values for the other 3 soil categories. The average of wildlife habitat quality-based endpoints (plant richness and cover) was calculated separate from the rangeland condition endpoint (OAT score). The highest of the two PELs was selected as the final PEL protective of wildlife habitat and rangeland condition for each soil category. The final PEL ranged from 4.11 to 4.98, depending on the soil category. Flat rocky soils had the highest PEL.

### **Assessment Report for Apache Tejo Wash**

Chino documented completion of an NMED request for a “ground survey” of the entire Apache Tejo drainage system in 2021. The “ground survey” includes historical information, as well as reconnaissance results and property ownership information. The “ground survey” covers the entire wash from the Chino tailing impoundments to Whitewater Creek, including two tributaries to Apache Tejo Wash. Chino also characterized tailing and other materials along the wash and interpreted the data with respect to the potential for exposure and the potential for re-release to other media, including groundwater. The Assessment Report concluded that the potential for constituents to infiltrate to groundwater is low; the tailing in the channel and the Big Berm generally showed little to no potential to generate acid or leach metals (Golder 2024). The depth to groundwater is 100 feet along the wash, and the vadose zone is comprised of alluvium and Gila Conglomerate with neutralizing potential. Also, there is limited potential to re-release constituents to surface water because recent, cleaner sediment has covered much of the historical tailing along the wash with a couple limited exceptions. The data were screened

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<sup>3</sup>Wildlife Reference North and South locations were sampled each year of community sampling to calibrate cover. The wildlife reference south location was originally thought to be a reference area but had lower pCu once sampled and became a “de minimis” location along with three bedrock reference locations that had lower pCu than 5 (see phytotoxicity report).

against STSIU Pre-FS RAC for birds and vegetation and no further action was recommended. Apache Tejo Wash will be further considered under Discharge Plan DP-1340 Sitewide Abatement.

## **1.4 AOC Requirements**

The AOC between Chino and NMED became effective on December 24, 1994 and requires Chino to conduct the following work:

- Assess present STSIU condition in the investigation area associated with risks to public health and welfare of the environment;
- To the extent necessary to select a remedy, or remedies, evaluate alternative remedial technologies appropriate for the Investigation Unit in the investigation area; and
- Implement the selected remedy or remedies.

FS activities that were identified in the AOC Scope of Work and described in the FS Proposal (Arcadis 2011a) include, but are not limited to:

- Description of current situation;
- Treatability studies and identification and screening of potential applicable technologies;
- Development of remedial alternatives;
- Initial screening of remedial alternatives;
- Detailed evaluation of remedial alternatives;
- Description and justification of preferred alternative; and
- Production of the FS report.

This FS addresses the above bullets.

## **1.5 Organization of FS**

This FS was prepared to determine and fulfil the needed data requirements of the AOC identified FS activities. The FS is organized as follows:

- Section 1.0: Introduction
- Section 2.0: Regulatory Components of the FS
- Section 3.0: Description of Current Situation: Soil
- Section 4.0: Description of Current Situation: Surface Water
- Section 5.0: Identification and Screening of Potential Applicable Technologies
- Section 6.0: Assembly and Development of Remediation Alternatives
- Section 7.0: Analysis of Alternatives
- Section 8.0: Final Acres for Remedial Actions or Monitoring

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- Section 9.0: References

## 2. Regulatory Components of the FS

There are a number of regulatory components associated with an FS. Section 2.1 presents the specific FS tasks that are stated and required in the AOC associated with STSIU. Section 2.2 presents the ARARs associated with the STSIU. Sections 2.3 and 2.4 provide details on the RAOs and the Pre-FS RAC that have been developed for the STSIU. Finally, Section 2.5 describes the AOC STSIU study boundaries.

### 2.1 AOC FS Tasks

AOC FS Tasks were presented in the FS Proposal (Arcadis 2011a) and are updated herein based on additional work completed.:

#### ***Description of Current Situation***

Updates to the current situation are detailed in Sections 3 and 4.

#### ***Treatability Studies and Identification and Screening of Potentially Applicable Technologies***

Remedial technologies were identified in the approved STSIU FS Proposal (Arcadis 2011a). As described in the FS Proposal, several of the remedial technologies would have relied on the results of pilot treatability studies, including ferrihydrite as a soil amendment, chelating agents as a soil amendment, *ex-situ* and *in-situ* soil washing, phytoextraction, and electrokinetic remediation. However, none of those technologies were retained for inclusion in the remedial alternatives assembled in this FS Report (see Section 6) and, therefore, no treatability studies were implemented other than the completed amendment study was initiated before the FS Proposal was submitted and approved. The technologies that were identified to carry forward are incorporated into remedial alternatives herein.

#### ***Development of Remedial Alternatives***

Remedial alternatives were developed in the FS Proposal (Arcadis 2011a) and no additional alternatives were identified or developed through the implementation of that proposal.

#### ***Initial Screening of Remedial Alternatives***

The initial screening of remedial alternatives was completed in the FS Proposal (Arcadis 2011a). For this FS, the alternatives of using chelating agents (for *ex-situ* and *in-situ* soil washing and phytoextraction), an impermeable soil cover, phytoremediation, and electrokinetic remediation are screened out, as described further below in Section 5.

#### ***Evaluation of Remedial Alternatives***

Remedial alternatives were evaluated in the FS Proposal (Arcadis 2011a) and no changes to that evaluation have been identified through the implementation of that Proposal. The remedial alternatives identified in the FS Proposal that were carried forward into the FS include the following:

Soil

- No Action
- Monitoring
- Excavation and Reuse
- Excavation and Disposal

- Limestone and Organic Matter Soil Amendment
- Tilling
- Ferrihydrite Soil Amendment
- Soil Cover
- Phytostabilization

#### Surface Water

- No Action
- Monitoring
- Excavation (In-Drainage, Upland, or Stock Ponds)
- In-Stream Removal of Suspended Sediments
- Limestone Treatment
- Sediment and Erosion Control

#### ***Description and Justification of Preferred Alternative***

In the FS Report, Chino shall describe and justify its preferred alternative based on the criteria listed above.

## **2.2 Applicable or Relevant and Appropriate Requirements**

ARARs are those cleanup standards, standards of control, and other substantive environmental protection requirements, criteria, or limitations promulgated under federal or state law that specifically address the situation at a CERCLA site. The requirement is applicable if the jurisdictional prerequisites of the standard show a direct correspondence when objectively compared to conditions at the site. An applicable federal requirement is an ARAR. An applicable state requirement is an ARAR only if it is more stringent than the federal ARAR.

If the requirement is not legally applicable, then the requirement is evaluated to determine whether it is relevant and appropriate. Relevant and appropriate requirements are those cleanup standards, standards of control, and other substantive environmental protection requirements, criteria, or limitations promulgated under federal or state law that, while not applicable, address problems or situations similar to the circumstances of the proposed response action (relevant) and are well suited to the conditions (appropriate) of the site. A requirement must be determined to be both relevant and appropriate in order to be considered (TBC) an ARAR.

The criteria for determining relevance and appropriateness are listed in the Code of Federal Regulations (40 CFR), Section 300.400(g)(2), and include general comparisons between the following:

- The purpose of the requirement and the purpose of the action;
- The medium regulated or affected by the requirement and the medium contaminated or affected at the site;
- The substances regulated by the requirement and the response action contemplated at the site;
- Any variances, waivers, or exemptions of the requirement and their availability for the circumstances at the site;

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- The type of place regulated and the type of place affected by the release; and
- Any consideration of use or potential use of affected resources in the requirement and the use or potential use of the affected resources at the site.

According to the USEPA CERCLA guidance, a requirement may be “applicable” or “relevant and appropriate” but not both (USEPA 1988). Identification of ARARs must be done on a site-specific basis and involves a two-part analysis: first, a determination of whether a given requirement is applicable; and then, if it is not applicable, a determination of whether it is, nevertheless, both relevant and appropriate. When the analysis determines that a requirement is not applicable but is both relevant and appropriate, the requirement must be compiled with the same degree as if it were applicable.

ARARs are generally divided into three categories: chemical specific; location specific; and action specific in accordance with USEPA guidance (USEPA 1988):

- **Chemical Specific:** Chemical specific ARARs are generally health or risk based numerical values or methods applied to site-specific conditions that results in the establishment of a cleanup level. Many potential ARARs associated with particular response alternative (such as closure) can be characterized as action-specific but include numerical values or methods to establish them so they fit in two categories, chemical-specific and action-specific.
- **Location Specific:** Location specific ARARs are included for environmentally sensitive areas including riparian and other hydrologic resources, and biological and other natural resources are the resource categories relating to location-specific requirements potentially affected by the STSIU remedial actions.
- **Action Specific:** Action specific ARARs are included for the potential remedial actions that will be used in the STSIU.

This classification was developed to aid in the identification of ARARs. Some ARARs do not fall precisely into one group or another. ARARs are identified on a site-specific basis for remedial actions where CERCLA authority is the basis for cleanup.

For the determination of relevance and appropriateness, the pertinent criteria were examined to determine whether the requirements address problems or situations sufficiently similar to the circumstances of the release or response action contemplated, and whether the requirement is well suited to the site. A negative determination of relevance and appropriateness indicates that the requirement does not meet the pertinent criteria.

To qualify as a state ARAR under CERCLA, a state requirement must be:

- A state law or regulation;
- An environmental or facility law or regulation;
- Promulgated;
- Substantive;
- More stringent than federal requirements;
- Identified in a timely manner; and
- Consistently applied.



To constitute an ARAR, a requirement must be substantive. Therefore, in some cases only the substantive provisions of requirements identified as ARARs in this analysis are considered to be ARARs. Permits are considered to be procedural or administrative requirements though may contain substantive requirements that are ARARs which must be attained and/or qualify as “to be considered” (TBC) materials that may be used in determining the necessary level of cleanup for protection of human health or the environment.

Provisions of generally relevant federal and state statutes and regulations that were determined to be procedural or not environmental in nature, including permit requirements, are not considered ARARs. CERCLA Section 121(e)(1), (42 USC Section 9621(e)(1)), states that “No Federal, State, or local permit shall be required for the portion of any removal or remedial action conducted entirely on-site, where such remedial action is selected and carried out in compliance with this section.” Consistent with 40 CFR, the term “on-site” is defined for purposes of this ARARs discussion as “the areal extent of contamination and all suitable areas in very close proximity to the contamination necessary for implementations of the response action.

In addition to ARARs, non-promulgated advisories, proposed standards, criteria, guidance, or policy documents developed by the federal or state government, or other information referred to as TBC materials may also be used in conjunction with ARARs to achieve an acceptable level of risk at a site. Although not legally binding, TBCs may be used when determining protective cleanup levels or response actions where no ARARs exist, or where ARARs alone would not be sufficiently protective of human health and the environment. Because TBCs are not ARARs, their early identification is not mandatory.

The state permit conditions for the Chino Mine shall be considered TBC materials and considered in the FS for developing remedial alternatives.

Chino had the primary responsibility for identifying federal ARARs for the STSIU. Potential federal ARARs that have been identified for the remediation of STSIU were determined in the RI (SRK 2008a) and FS Proposal (Arcadis 2011a) and are presented in Tables 2-1, 2-2, and 2-3. Pursuant to the definition of the term “on-site” in 40 CFR Section 300.5, the area that is considered part of the remedial action is STSIU (see Section 2.5).

## **2.3 Remedial Action Objectives**

This section identifies the environmental media for the STSIU where potentially unacceptable risks were determined to exist through the risk assessments completed during the RI, as well as the constituents determined to be responsible for the potential for unacceptable risk. This section also presents the specific RAOs developed for the STSIU for each media of interest.

RAOs are medium-specific goals designed to protect human health and the environment. RAOs serve to focus an FS and provide context for the overall scope of potential cleanup activities at a site. Each RAO specifies: the contaminant of concern; the relevant exposure routes and receptors; and an acceptable contaminant concentration or range of concentrations for each exposure pathway.

The STSIU RI referenced the completed STSIU HHRA and ERA to determine if any constituents present in the environmental media should be considered Constituents of Potential Concern (COPCs). The sitewide and STSIU-specific ERAs (NewFields 2006a, 2008) considered sensitive representative receptors from a number of receptor classes including mammals, birds, plants, and invertebrates. NewFields evaluated direct contact for plants and invertebrates and incidental soil ingestion and food-chain transfer for birds and mammals. The RI also referenced the comprehensive HHRA performed by Gradient (2008) to determine if any chemicals present in environmental media at the site are responsible for potentially unacceptable risk to human receptors in the context of plans for

future site use. Accordingly, the human receptor classes evaluated in the HHRA included current resident, future resident, trespasser, construction worker, rancher, industrial worker, recreator swimmer, and trespasser swimmer. Specific pathways considered during the HHRA included direct dermal contact with surface soil, incidental ingestion of surface soil, inhalation of surface soil, dermal contact with surface water, and incidental ingestion of surface water. The risk assessments were implemented according to appropriate guidance and methodologies, which along with the detailed results from the assessments, were previously presented in the STSIU HHRA and ERA reports (Gradient 2008; NewFields 2006a, 2008).

NMED, after reviewing the HHRA and ERA, concluded that arsenic, copper, and iron are potential soil-based risk drivers for at least one human receptor evaluated, and that copper and pCu are potential soil-based risk drivers for at least one ecological receptor evaluated for the STSIU. The Gradient HHRA and NewFields ERA will be discussed in detail in Section 3.1.1.

Based on the findings from the STSIU RI Report, HHRA and ERAs (SRK 2008a,b, Gradient 2008, NewFields 2006a, 2008), the RAOs for the STSIU include:

- Protection of human receptors, as represented by current and future resident or industrial workers, from exposure to arsenic, copper and iron from ingestion of contaminated soil.
- Prevent the ingestion of copper by the small ground-feeding bird (SGFB) receptor at levels that result in unacceptable population-level risks.
- Toxicity to vegetation or other biological elements of habitat should be reduced to levels that allow for a self-sustaining ecosystem and prevent adverse impacts on local wildlife populations or subpopulations. In areas where habitat function is degraded due to toxicity of elevated copper concentrations and/or decreased pH from either smelter emissions or contamination released from tailings impoundments, remedial actions should focus on the restoration of wildlife habitat. Newfields (2006b) also indicated that “risk management actions to protect wildlife and wildlife habitat should recognize the impacts of other land-uses such as grazing”.
- Restore water quality to water quality objectives that are protective of beneficial uses within reasonable timeframe and maintain existing water quality that complies with water quality objectives. RAOs should reduce the likelihood of contact between surface water and soils/sediments that contain heavy metal contaminants at concentrations that could cause deleterious effects to aquatic receptor populations.

## 2.4 Pre-FS RAC

In a letter dated September 16, 2010 and then amended via a dispute resolution letter dated March 3, 2011, NMED provided Chino with a Pre-FS RAC for the STSIU (NMED 2010a, 2011a). Based upon the information documented in the ERA, HHRA, as well as the comments and input provided from all parties, NMED has determined the Pre-FS RAC values for soil and surface water as discussed below.

### **Surface Soil**

NMED has determined a Pre-FS RAC for metals that potentially pose human health and ecological risk based on the information documented in the ERA, HHRA, and probability analysis. NMED established HHRA Pre-FS RAC values for arsenic, copper, and iron as follows:

- Arsenic: NMED selected the cancer target risk with a Pre-FS RAC of 27 mg/kg. This value is supported by the probability analysis (Arcadis 2010c,d,e) and is consistent with the range of arsenic clean-up levels previously set in New Mexico by USEPA.

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- Copper: NMED selected the non-cancer risk human health Pre-FS RAC as previously determined for the HSIU of < 5,000 mg/kg copper at private, commercial, and public developed properties.
- Iron: NMED selected the non-cancer risk Pre-FS RAC based on the probability analysis (Arcadis 2010c,d,e) of 100,000 mg/kg.

NMED established ecological Pre-FS RAC values for copper as follows:

- To reduce soil toxicity to plants from copper to  $pCu \geq 5$  where copper is > 327 mg/kg.
- To reduce soil toxicity to SGFB of 1,600 mg/kg. The SGFB Pre-FS RAC is applicable to the 95UCL of the area-weighted average concentration of copper in surface soil (0-6") within exposure units in the STSIU. In addition, NMED required monitoring for copper concentrations in surface soil between 1,100 and 1,600 mg/kg.

Based on the final Pre-FS RAC issued in a letter dated March 2011, NMED stated:

*Since the FS and ROD will be completed consistent with the NCP, new information can be used to refine RACs and selection of alternatives. This is supported by the NCP in §300.430(e)(2)(i) which states "Establish remedial action objectives specifying contaminants and media of concern, potential exposure pathways, and remediation goals. Initially, preliminary remediation goals are developed based on readily available information, such as chemical-specific ARARs or other reliable information. Preliminary remediation goals should be modified, as necessary, as more information becomes available during the RI/FS. Final remediation goals will be determined when the remedy is selected. Remediation goals shall establish acceptable exposure levels that are protective of human health and the environment..." It must be noted that NMED's pre-FS RACs are equivalent to preliminary remediation goals referred to in the NCP."*

Thus, Pre-FS RAC are consistent with the use of PRG by USEPA in the NCP, and new information can be used to refine the Pre-FS RAC and selection of alternatives. Additional information is provided herein in Section 1.3 and Appendices A, B, C and D. As discussed in Section 1.3, based on discussions and field work conducted with NMED, soil sampling and vegetation community measurements were conducted in 2018 from new agreed-upon locations. Results were incorporated into calculations of DELs and PELs presented in the phytotoxicity study for four soil/topographic categories (see Section 3.1.5 in Appendix C, referred to as soil categories in this FS). The average PELs for pCu across the soil categories ranged from 4.0 to 4.98, with the highest values in the flat rocky soil category. Based on this new information, the PELs are used to identify acres for remedial alternative evaluation herein. Moreover, percent cover, richness and OAT are used to further identify poor rangeland as described in the approved FS Work Plan (Arcadis 2011a).

In cases where the above criteria overlap for a given area, the constituent which requires the largest remedial footprint will be considered the "risk driver" and the remedial technology will be selected to address this "driver". The selected remedial technology will be evaluated to confirm that all "non-risk driver" constituents are in compliance with their respective RACs. With the issuance of the Pre-FS RAC, constituents identified in the RI and risk assessment as COPCs are now considered constituents of concern or COCs. The FS and ROD will be completed consistent with the NCP. Final remediation goals will be documented in the ROD.

### **Surface Water**

NMED selected the Pre-FS RAC for surface water based upon the State of New Mexico Standards for Interstate and Intrastate Surface Waters, Part 20.6.4 NMAC for risk to aquatic life. The Pre-FS RAC for all constituents is Part 20.6.4 NMAC, including all approaches and tools listed in the Code which provide options for site-specific

application. New Mexico promulgated rulemaking to codify the application of site-specific criteria for copper and also to codify the use attainability of the drainages, see §20.6.4.808 – 809.

### **Ground Water**

NMED (2010a) discovered a data gap in the RI Report during the preparation of Pre-FS RAC letter. Drainage sediment sample results exceed the NMED Dilution Attenuation Factors (DAF) developed to protect ground water. Although this potential impact was included in the conceptual site model (CSM) as insignificant (along with sediment, Figure 3-1) and in the risk assessments, NMED requested Chino to further investigate this potential pathway. NMED requested that Chino collect additional samples from all the drainages that currently exceed the DAF 1 or exceed background concentrations for the following COPCs: arsenic, barium, cadmium, copper, iron, molybdenum, and selenium.

Chino submitted a work plan on October 20, 2010 and NMED approved the work plan on November 10, 2010. Chino implemented the work plan in December 2010 with NMED participation. The samples were analyzed using 1) Synthetic Precipitation Leaching Procedure and 2) Acid-base Accounting and the results were compared to Ground Water Quality Standards to determine if this potential pathway may impact groundwater and if potential remedial alternatives should be developed in the FS. The results showed that there is no risk associated with sediments leaching to groundwater; therefore, groundwater will not be evaluated as part of the STSIU FS. The results from the sampling effort were presented in the *Groundwater Quality Pre-Feasibility Study Remedial Action Criteria for Drainage Sediments* (Arcadis 2011b).

## **2.5 AOC Study Boundaries**

As specifically described in the AOC Scope of Work (NMED 1994), the STSIU includes all areas containing and proximal to Chino's copper smelter including the slag, all areas containing or which contained facilities ancillary to the primary smelter, and the soil adjacent or impacted by the tailings facility (excluding the HSIU). The investigation area is contained in section 31 of T18S, R12W and sections 4, 5, 6, 7, 8, 9, 16, 17, 18, 19, 20, 21, 28, 29, 30, 31, 32, and 33 of T19S, R12W. The above township-range sections, as shown in Figure 2-1, include the town of Hurley, Chino operational area, and all tailings ponds; however, the town of Hurley and the former Hurley Operations Area are excluded from the Investigation Unit because applicable remediation or closure of these areas are covered under separate regulatory activities. The western extent of the 1994 AOC area is Highway 180, the eastern extent is approximately a quarter mile east of the toe of Tailings Pond 7, and the southern extent is approximately 1 mile south of Tailings Pond 7.

The former Hurley Smelter was decommissioned and demolished in 2007. The Hurley Operations Area is bounded by the Town of Hurley to the west, Whitewater Creek to the northeast, Lake One to the east, and the tailings impoundments to the south. Current land uses adjacent to the former Hurley Operations Area are residential in the towns of Hurley and North Hurley, tailings disposal south of the operational area, and livestock grazing elsewhere. Chino owns the majority of land included in the STSIU and the majority of this land is currently leased for livestock grazing.

The above investigation area has been expanded over the last 15 years as sampling events have delineated the boundaries of the smelter impacted soils with greater accuracy. The current smelter investigation area now includes areas to the north, east, and west of the original Smelter Investigation Unit located in section 31 of T18S, R12W that have copper concentrations greater than the established background value of 327 mg/kg.

## 3. Description of Current Situation: Soil

The following sections describe the current understanding of the physical characteristics of the STSIU surface soil based on previous field investigations.

- Section 3.1 summarizes the conceptual site model,
- Section 3.2 addresses how the Pre-FS RAC may be applied across the relatively expansive area encompassed by the STSIU as exposure units, and
- Section 3.3 discusses the upland areas that require potential remedial action as a result of the evaluation.

### 3.1 Conceptual Site Model

The CSM for sources associated with the STSIU was originally presented in the RI Proposal for STSIU (SRK 2004) as well as in the risk assessments (Gradient 2008, NewFields 2006b, 2008). Figure 3-1 illustrates a CSM via pathway segments and mechanisms required to understand how potential contamination occurred, including the source, release and transport of mineral processing constituents. Primary sources listed in this CSM are historical mineral processing activities and tailing impoundments including historical and current facilities.

Soils within the STSIU were affected by historical stack and fugitive dust emissions from historical mineral processing activities associated with the former Hurley Operations area and the tailings area. The AOC Background Report (Chino 1995) identified the primary sources of contamination in the former Hurley Operations Area resulting from historical smelting, milling, concentrating, and handling of copper bearing material. For the tailings area of the STSIU, the tailings impoundments are considered the primary source of constituents of interest that have the potential to affect environmental media. Both historical and currently operational tailing impoundments may have affected surrounding media. Pathways such as historical run-off and fugitive dust emissions transported constituents in tailings to surrounding soil, drainages, and groundwater. Fugitive dust emissions from the tailings ponds occur when high winds mobilized tailings resulting in transport and subsequent deposition of constituents to the surrounding soil. For groundwater impacts, the state discharge permits, DP-484 and DP-1340, require addressing any groundwater impacts from the historical or current impoundments that infiltrated to groundwater under a sitewide abatement program (Golder 2016). Groundwater impacts are addressed under that program rather than in this FS under the AOC. Groundwater for the STSIU under that program is addressed through monitoring on a periodic basis to ensure it meets permit conditions. Sediment leaching to groundwater was evaluated via request by NMED (NMED 2010a), an approved work plan (Arcadis 2010f, NMED 2010b), a report (Arcadis 2011b), and an approval letter from NMED that sediment does not need to be evaluated in the FS (NMED 2011b).

Prevailing winds tend to be from the northwest (Chino 1995); therefore, surface soils in areas to the south and east of the former Hurley Operations Area and the tailings impoundments are likely to have been the most affected by dryfall from these aerial sources. Following airborne deposition onto soils, metals and other inorganic constituents may be further redistributed by a combination of physical (air and water erosion) and/or chemical (leaching) processes. Windblown tailing (WBT) deposition onto soils currently accumulates south and east of the active Tailing Pond 7 (TP 7) operations within its operational boundaries as well as onto acres investigated under the AOC but within the Mining Permit Boundary. Chino addresses this TP 7 affected area in the recently updated CCP, Section 6.3 (shown in Figure 12 and Plate 25 of the CCP, see Figures 3-2a and 3-2b) (Chino 2024) as

described below. The WBT areas shown on Figure 3-2a and 3-2b addressed in the CCP will not be considered for remedy in this FS due to on-going operations.

### **CCP - 6.3.6 Miscellaneous SMA Reclamation Areas**

“Wind-blown tailings were identified east and south of Tailing Pond 7 within the Tailing Investigation Unit under the Chino Administrative Order on Consent (AOC). The AOC investigation and feasibility study based on 2007 and 2011 sample data determined that there are no ecological or human health risks, nor are there communication of constituents to groundwater. However, the tailing deposition sources from the currently active Tailing Pond 7 operations, and those affected acres fall under the operational and closure permit boundaries. These deposition sources are not part of the AOC’s definitions for investigation. Thus, this CCP update includes removing an average of 6 inches of windblown tailing from the affected area, ripping the exposed substrate, and revegetating.” For the FS, areas east of TP7 are assessed but will not be considered for FS remedy due to proximity to current operations effects. Chino operations have implemented additional source controls to address the higher lifts of Tailing Pond 7 and decrease any wind-blown effects. Additionally, Chino is currently evaluating the construction of another tailing pond to the south of TP7 and in Whitewater Creek basin; however, the permit application process has not been initiated with NMED, and the expansion is not addressed in this FS.

Areas retained for remediation were evaluated for overlap with the windblown tailings polygon in Figure 3-2a and were removed from remediation if they fell within the polygon.

Figure 3-1c identifies the following potential release mechanisms associated with affected surface soil:

- Aerial dispersion of dust and re-suspension:  
Mobilization of affected surface soil to air by wind;
- Run-off: Transport of suspended or dissolved constituents in surface water run-off. Potentially affected sediments may be generated as a result of surface water erosion of surface soil;
- Infiltration: Infiltration of surface water through affected surface soil may potentially release constituents of interest from affected surface soil into groundwater; and

The potential soil exposure routes for current and future human receptors include:

- Inhalation of dust;
- Ingestion of contaminated soil (incidental) and food items from garden foods or animals grazing on the soil; and
- Direct contact (dermal).

The potential exposure routes for current and future ecological receptors include:

- Direct contact (root uptake); and
- Ingestion (contaminated biota and soil).

The above potential human health and ecological exposure routes were evaluated in the STSIU HHRA (Gradient 2008) and STSIU ERA (NewFields 2008). Since the completion and approval of the risk assessments, the STSIU has been naturally “limed” by the “white rain” event as described in Section 1.3. This natural event has shifted the

geochemistry of the soils and affected the bioavailability of contamination in the STSIU, as documented in the *Year 5 pH Monitoring Report for STSIU* (Arcadis 2023).

In addition to the white rain event, historic sources to STSIU were fully reclaimed, as described further below in Section 3.1.2.

### **3.1.1 Previous Investigations**

The upland soil in the STSIU was evaluated in the STSIU RI (SRK 2008a,b), the HHRA (Gradient 2008), and the ERA (NewFields 2008) prior to the development and issuing of the Pre-FS RAC. The following sections summarize the findings of these three reports.

#### ***Remedial Investigation***

The STSIU RI Proposal was completed by SRK in 2004 and the RI Report was submitted in 2006 and resubmitted in 2008 with supplemental data (SRK 2008a,b). Surface soil samples were taken at 165 locations in the STSIU during the RI sampling efforts (SRK 2008a,b).

The RI concluded that copper is the primary metal elevated in surface soils in the STSIU. In addition, arsenic, cadmium, and iron were also detected above the decision criteria, but their respective concentrations largely fall within the range of reference, and elevated concentrations of these metals generally fall within the footprint of copper concentrations. Consistent with conclusions reported in the AOC Background Report (Chino 1995) and the *Phase II Remedial Investigation Report for the Ecological Investigation Unit* (Arcadis 2001), copper concentrations in surface soil exhibit strong spatial characteristics with decreasing concentrations as distance in the direction of prevailing winds increases (SRK 2008a).

#### ***Human Health Risk Assessment***

In 2008, Gradient evaluated the risks to human health posed by constituent concentrations in the STSIU. Gradient calculated both cancer risks and non-cancer hazards for potential receptors on the site. The methodology used for this risk assessment was consistent with USEPA guidelines and used conservative, default assumptions, whenever site-specific data were not available. The receptors evaluated in the HHRA included current and future residents, adolescent recreators, adolescent trespassers, ranchers, construction workers, and industrial workers. The upland risk assessment evaluated exposures to soil, windblown dust in air, and locally produced food items.

Gradient indicated that site exposure to soils could result in unacceptable cancer risks for current and future residents in all five exposure areas of the STSIU that they delineated and evaluated. However, these cancer risks are largely driven by the consumption of locally grown foods. Gradient stated that the “consumption of locally grown food” pathway was evaluated using conservative assumptions and tends to overestimate risk. Gradient also found that 90% of the excess lifetime cancer risk is attributed to arsenic. All other receptors (recreators, trespassers, ranchers, construction workers, and smelter workers) did not have unacceptable cancer risk. Furthermore, all residential non-cancer risks in the STSIU were lower than residential reasonable maximum exposure non-cancer hazard calculated for the reference area. All other receptors (recreators, trespassers, ranchers, construction workers, and smelter workers) had acceptable non-cancer hazards.

Gradient evaluated copper risks separately using a probabilistic method, only for ingestion of soil. The most sensitive endpoint for copper toxicity is nausea; therefore, copper risks were based on estimating the annual number of nausea episodes that an individual might experience, at a given soil copper concentration. Overall, the industrial worker in the smelter area had the highest copper risk, with an estimated 65 nausea events per year.

Based on the results, NMED established human health RACs for three constituents, including arsenic, copper, and iron, as discussed in Section 2.4.

### ***Ecological Risk Assessment***

The Site Wide ERA (NewFields 2006a) was completed and used as the basis to streamline the Investigation Unit -specific ERA for the STSIU (NewFields 2008). These ERA reports evaluated the risks from soil to terrestrial receptors. The methodology used for this risk assessment was consistent with USEPA guidelines, and used conservative, default assumptions, whenever site-specific data were not available. The methodology and parameter selection are described in Technical Memo 1 – ERA Workplan and the sampling and analytical approach are described in Technical Memo 2 (Schafer and Associates 1999a, 1999b). The receptors evaluated in the terrestrial ERA included terrestrial vegetation, herbivorous birds, omnivorous birds, raptors, herbivorous mammals, omnivorous mammals, ruminants, and mammalian predators. The risk assessment evaluated exposures from direct contact, incidental soil ingestion, and ingestion of prey items.

It was concluded that the risks to plants and wildlife were primarily related to elevated copper concentrations and depressed pH in soil (NewFields 2008). The risks to the SGFB, the most sensitive receptor evaluated, appear to be elevated in the STSIU due to exposure to copper. The potential for risk, for both plants and the SGFB, is greatest in the areas immediately to the east of the smelter and the tailings impoundments and decreases with the increased distance to the east of those features.

Since 2008, when the NewFields STSIU ERA was approved by NMED, there has been further analysis of the risk to the SGFB. The following studies and updated exposure parameters have led to a decrease in predicted exposure levels for the SGFB.

- Ingestion Rate of Food and Soil: The ingestion rates used by NewFields in the ERA were updated to more recent allometric-based values in the literature (Nagy 2001).
- The dose equation was updated to model all potential dietary intakes in dry weight.
- Insect tissue concentrations: Additional insect tissue was collected in 2010 to supplement the tissue that was collected and analysed for in the ERA. The results of this sampling event are presented in the *Terrestrial Invertebrate Copper Bioaccumulation and Bioavailability Study* (Arcadis 2010a).

The reason the invertebrate study was conducted was to further refine the conceptual model of copper bioaccumulation into and bioavailability of invertebrates in the Site Wide ERA (NewFields 2006a). Chino developed and NMED approved the *Terrestrial Invertebrate Copper Bioaccumulation and Bioavailability Study for STSIU* (Arcadis 2010a), and subsequently field data were collected with NMED. The results of this investigation indicate that exposure levels from the invertebrate diet were lower than reported in the ERA for the SGFB. Specifically, incorporating the data from the terrestrial invertebrate bioavailability study changed the risk analysis such that the avian pre-FS RAC for the STSIU was increased from its initial proposed value of 626 mg/kg based on the sitewide ERA and STSIU ERA to 1,600 mg/kg for soil. This information was incorporated into the pre-FS RAC (Arcadis 2011d).

### ***Other Risk-Related Assessments***

As summarized in Section 1.3, the amendment study, pH monitoring report, and draft phytotoxicity report, were submitted subsequent to the approval of the FS Proposal, and all three are referenced herein as Appendices A, B, and C, respectively.



### 3.1.2 Previous Reclamation and Interim Actions

The Hurley Mill was closed in 1985 and the Smelter was closed and demolished in 2003 and 2007, respectively. The historical mineral processing operational area was fully reclaimed with 3 feet of borrow fill including slag, improving overall air quality, and eliminating the source of acid and copper to STSIU soils. Closure plans were implemented under DP-1340 for the reclamation of historical tailing impoundments (Tailings Pond 1, 2, B, C, 4, 6E and 6W) (Golder 2013a,b). The Town of Hurley, which is located in the HSIU, was remediated in 2006 and 2007. In addition, also under DP-1340, Condition 90, the Lake One Closure Plan was submitted to NMED in December 2012, approved by NMED, and implemented in 2013 (EMC<sup>2</sup> 2014). Borrow fill for reclamation was taken from within the STSIU boundaries, outside of the operational footprint as discussed in Section 1.3.

Additionally, interim actions were completed in 2008 and 2012, as summarized in the *STSIU Interim Removal Action Completion Report* (“Golf Course IRA”) and the *Supplemental STSIU Interim Removal Action Completion Report* (“Railroad IRA”) (Arcadis 2009 and Golder 2013c). Specifically, these areas included the area west of Highway 180, the South Golf Course, the North Golf Course, the Alley Way adjacent to the South Golf Course, the area east of Highway 180, and other sites located at the end of the South and North Golf Courses. The objective of the interim actions was to remove soils where copper concentrations were greater than 5,000 mg/kg (lateral delineation) at 0 – 1” based on the 2005 NMED-approved residential RAC for the HSIU (and later formally issued in 2011 as pre-FS RAC for STSIU), as the acres were deemed future residential as extensions of the town of Hurley and fall within the current city limits. Within the areas with higher than or equal to 5,000 mg/kg, soils were removed vertically until the copper concentrations were less than 2,700 mg/kg or resistance met.

Later interim actions were also completed in 2014 and 2019 as summarized in the *STSIU Razorback Ridge Interim Remedial Action Completion Report* (Golder 2015) and the *B-Ranch Interim Remedial Action Completion Report* (Arcadis 2021). The Razorback Ridge IRA is located east of Lake One in areas adjacent to the Whitewater Creek Diversion Channel and James Canyon. The B-Ranch IRA is located adjacent to the reclaimed slag and the Golf Course IRA with Whitewater Creek bounding the east side of the site. The objective of the IRAs was to remove soils in areas with copper concentrations higher than 5,000 mg/kg, delineated laterally, as per the STSIU pre-FS RAC. As the soil removed from the Razorback Ridge Area was being utilized as borrow fill material, removal depths were at 1 to 2 feet depth and only post removal sampling was performed. Post-confirmation and post-removal data associated with the above-described removals were incorporated into Chino’s maps and databases, as summarized in Table 3-2.

### 3.1.3 FS Proposal Data Collection

The FS Proposal summarized additional sampling needed to refine the potential remedial areas for all constituents that exceed their respective Pre-FS RACs (Arcadis 2011a). Additional soil samples were collected to more accurately define the current extent of contamination for copper and pCu in surface soil. The exact sampling locations and methods were presented in the Upland Sampling Work Plan (Appendix A) of the FS Proposal (Arcadis 2011a) and discussed in Appendix D in Section 4.

Additionally, exposure units are delineated via desktop evaluation with field verification. The Upland Sampling Work Plan (Appendix A) of the FS Proposal described the methodology that is used to perform rangeland and drainage habitat surveys to better understand the potential habitat units in the STSIU. NMED required that the SGFB copper Pre-FS RAC be applied as a 95UCL area-weighted average (defined in next section) over a habitat unit. Through use of vegetation alliance polygons (Daniel B. Stephens & Associates [DBS&A] 1999, NewFields 2006a), SGFB habitat units are defined for the STSIU and presented herein. Similarly, the pCu Pre-FS RAC is

applied as an average concentration over a habitat unit. These copper and pCu exposure units are described in detail in Section 8 of Appendix D. Datasets and exposure units are summarized below in Section 3.2.

## 3.2 Nature and Extent of Contamination

According to the AOC, the soils of concern in the STSIU are the areas that exceed the Pre-FS RAC values described in Section 2.4 of:

- < 5,000 mg/kg copper at 0-1” bgs,
- 1,600 mg/kg copper area-weighted 95UCL of the mean surface soil concentration for a habitat unit for 0-6” bgs,
- 27 mg/kg arsenic at 0-1” bgs, and
- 100,000 mg/kg iron at 0-1”.

A “95UCL” is the upper end of a confidence interval around the mean concentration, where the confidence interval represents a specified probability (95% in this case) that the value of the mean lies within it. It is a conservative estimate (overestimate) of the mean concentration. “Area-weighted” indicates that each sample point’s concentration is weighted by the amount of ground area the sample represents.

In addition to the Pre-FS RAC described above, based upon new information submitted herein and summarized in Sections 1.3 and 2.4, pCu less than 5 with soil copper concentration greater than 327 mg/kg for 0-6” bgs as well as richness, cover and rangeland condition (measured with the OAT score) are used to identify acres of soils of concern.

There were four constituents evaluated (arsenic, copper, iron, pCu) and three constituents (excludes arsenic) that exceed NMED’s Pre-FS RAC for soil in the STSIU. Copper and iron have exceedances of the NMED human health criteria, while copper and pCu fail their respective ecological criteria. Arsenic did not have exceedances. Each constituent is described further in sub-sections below.

### 3.2.1 Arsenic

The RI dataset consisted of surface soil samples collected from 165 locations (SRK 2008a,b). For this FS, Chino removed 33 samples from the metals analytical dataset due to their location overlapping with interim action areas that were cleaned up to pre-FS RAC standards, as discussed in Appendix D, Section 6.1 (Sample IDs of 64 through 96, Table 3-1), resulting in a dataset of 132 samples available for the FS from the RI. Of the 132 RI surface soil samples with arsenic data not in remediated areas, there are no exceedances of the human health Pre-FS RAC of 27 mg/kg (Table 3-1). Therefore, arsenic will not be further evaluated in the STSIU FS.

### 3.2.2 Copper

There were 350 pre-IRA surface (0-1” bgs) and shallow (0-6” bgs) soil samples collected to analyze copper in the STSIU through December 2012. These samples were taken during the following investigations: AOC Background Report (Chino 1995), Site Wide ERA (NewFields 2006b), STSIU RI (SRK 2008a,b), and the STISU FS Upland Sampling Plan (Arcadis 2011a). The results from the FS Upland Sampling Plan are discussed in Section 6 of Appendix D. An additional 51 samples were collected between 2012 and 2018 as part of the pH monitoring study

(Arcadis 2023) and amendment study (Arcadis 2017)<sup>4</sup>. Together these 401 samples comprise the copper dataset not including IRA samples and are identified as “pre-IRA” in Table 3-2 (though there was some overlap chronologically in the collection of samples from these studies and the IRAs). Finally, 1,496 samples were collected during four IRAs: Golf Course (Arcadis 2009), Railroad (Golder 2013c), Razorback Ridge (Golder 2015), and B-Ranch (Arcadis 2021). Table 3-2 summarizes all 1,931 copper samples collected in STSIU; development of the copper dataset shown in Table 3-2 is discussed in detail in Section 6 of Appendix D. All IRA delineation and confirmation samples were analysed using x-ray fluorescence (XRF) and corrected using the regression equation based on a subset of the samples analysed by a laboratory. The regression equation and methodology were presented in the IRA Completion Reports. The Razorback Ridge and B Ranch Completion Reports did not report corrected XRF data; however, paired datasets with XRF and laboratory copper results were reported and a regression was developed from those data for this FS and used to correct the data, which are reported in Table 3-2. All borrow areas, which were excavated at depths of 1 to 12 feet, were assumed to have copper concentrations at background, equal to 327 mg/kg. This background value of 327 mg/kg for copper is consistent with the soil copper portion of the pre-FS RAC for plants and was derived from the 95UCL on the 95<sup>th</sup> percentile concentration of nine upland reference soil samples (see Tables 6-1 and Table 2-1 in the STSIU HHRA [Gradient 2008]).<sup>5</sup> As part of the borrow excavation, some surficial WBTs were also removed but not sampled. Any older samples previously located in these areas were substituted with 327 mg/kg. The ‘post-IRA’ dataset analyzed in this FS is comprised of 1,851 samples including all copper data shown in Table 3-2 after substitutions for background values where appropriate, with the exception of samples removed due to occurring in inaccessible areas or occurring outside of the extent of the vegetation alliance polygons, or because the sample was from dust wiped off bedrock (as shown in Table 3-2 and described further in Section 6 of Appendix D; binned copper concentrations and source studies for the full post-IRA dataset are shown on Figure D-7 in Appendix D). The risks from copper in the STSIU are driven by exceedances of the pre-FS RAC for both human health and ecological receptors. The human health risks were evaluated with copper sampled in the 0-1” bgs interval, sieved to < 0.25 millimeters (mm). The risk to ecological receptors was evaluated using the soil horizon of 0-6” bgs, sieved to < 2 mm. Historically, most copper samples were sampled from the 0-1” bgs interval to determine potential risk to human health. Using the methodology presented in Appendix A of the FS Proposal (Arcadis 2011a), for ecological evaluations, the copper results from the 0-1” were adjusted to 0-6” using the median ratio between the two depths (0.7 unless in windblown tailing area, where it was 1.5, Table 3-2). This adjustment (described in more detail in Section 6.1 of Appendix D) provided a more robust 0-6” dataset to evaluate the remedial alternatives. The locations and concentrations of the copper samples evaluated in the pre-IRA dataset are shown in Figure 3-2. As shown in Figure 3-2, copper concentrations were the greatest to the east of the historical smelter and decrease with increased distance from the historical smelter’s location. There were also areas of elevated copper on the north, south, and west of the town of Hurley from smelting and other historical mineral processing activities. Similar to the elevated areas to the east, copper concentrations to the north, south and west decrease with increased distance from Chino’s operational area.

As discussed in Section 2.4, the March 3, 2011 NMED Chino AOC Informal Dispute Resolution STSIU memo states that the copper SGFB Pre-FS RAC value of 1,600 mg/kg is intended as the 95UCL area-weighted average concentration within an exposure unit, and the exposure unit should be delineated based on habitat. The habitat

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<sup>4</sup> Samples collected as part of the phytotoxicity and community study (Arcadis 2018) were not included in the pre-IRA dataset evaluated in the FS because the copper dataset collected for the RI provided adequate spatial coverage of the sampling area for this study.

<sup>5</sup> The data were originally from Table 6.4-6 in Chino (1995). The nine upland reference soil samples included seven near the airport and two located 2 miles west of Bayard.

polygons represented by the vegetation alliance map developed by DBS&A (1999) (and referenced by NewFields [2006a]) were used as the exposure units for upland areas (Figure 3-3) as discussed in the approved FS Proposal (Arcadis 2011a).<sup>6</sup> During the 2011 field effort, a subset of the STSIU drainages was evaluated to determine if the avian habitat along the drainage banks was different from the adjacent upland. It was determined that the STSIU drainage banks were not significantly different from the adjacent upland (Section 8 of Appendix D); therefore, the existing vegetation alliance polygons that encompass the upland and banks of the drainages were used as the ecological exposure unit for SGFB, without differentiating between upland and banks of drainages.

The datasets for each exposure unit for the SGFB were obtained by intersecting ArcGIS interpolated copper Thiessen polygons constructed from the post-IRA dataset (includes effect of borrow removal) with DBS&A vegetation alliance polygons (Figure 3-4). Thiessen polygons are the polygons that result from the method used to identify the ground area each sample represents, used to weight the 95UCL (as discussed in Section 3.2). Thiessen polygons are generated from the set of sample points. Each Thiessen polygon defines an area of influence around its sample point, so that any location inside the polygon is closer to that point than any of the other samples. It is a commonly used method of interpolating chemical concentrations across the landscape, and a method that lends itself to calculating 95% confidence intervals as required by the pre-FS RAC. Details on the choice of using the Thiessen polygon method and calculation of the 95UCL are presented in Appendix D, Section 6.2.

All Thiessen polygons with either lab or XRF sample results<sup>7</sup> that exceeded the human health RAC of 5,000 mg/kg have already been remediated (if possible) to be in compliance with the human health RAC. Some sample points in Table 3-2 had copper greater than the 5000 mg/kg RAC but were in areas that could not be remediated, including (1) areas where bedrock was noted and the sample result was dust wiped off the rock, (2) inaccessible steep sloped areas, (3) around infrastructure, or (4) around ROW.

Based on the 1,100 mg/kg monitoring criteria for the avian pre-FS RAC, exposure units (or habitat polygons) which intersected a copper Thiessen polygon with a concentration > 1,100 mg/kg were retained for further evaluation. For each retained exposure unit, a 95UCL area-weighted average concentration was calculated, using 500 area-weighted bootstrapped datasets. Specifically, the copper data in the exposure unit were re-sampled 500 times with replacement, with probability of a sample being selected in the bootstrap sample equal to the relative area of its Thiessen polygon. The 500 datasets were each entered into ProUCL to select the recommended 95UCL for each dataset. The final 95UCL was the average of the 500 95UCLs (Casmalia Resources Site Steering Committee 2011).

Table 3-3 includes the exposure units that have 95UCL concentrations in excess of 1,100 mg/kg. The colors of the exposure units in Figure 3-5 identify which of these post-IRA exposure units require remedial technology evaluation, additional monitoring, or no further action to be in compliance with the avian pre-FS RAC, depending on the 95UCL copper concentration of the unit. All exposure units had 95UCL concentrations less than or equal to the avian RAC of 1,600 mg/kg (ratios  $\leq 1.0$  in Table 3-2). Exposure units in yellow have a 95UCL between 1,100 mg/kg and 1,600 mg/kg and will require additional monitoring for the protection of the small ground-feeding bird (vegetation alliance polygons 1-3, 10-16, and 88-15). Exposure units in green do not require any additional action.

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<sup>6</sup> Alliance polygon size shifted for several polygons based on updates to the habitat map and an addition of “mine facilities/other” category.

<sup>7</sup> Results based on 0-1” and sieved to < 0.25 mm.

### 3.2.3 Iron

Two of the 132 soil samples collected for the RI and used in the FS that were analysed by USEPA 6010 exceed the human health Pre-FS RAC of 100,000 mg/kg (Table 3-1). The locations of these samples (SS145 & SS148) are along the eastern border of Tailing Dam 7, as seen in Figure 3-6. The iron results from SS145 and SS148 only slightly exceed the human health Pre-FS RAC with concentrations of 141,000 mg/kg and 123,000 mg/kg, respectively. However, the area encompassing these two sample locations falls within the currently operational Tailing Pond 7 footprint and will be addressed under the CCP and DP-1340. Thus, iron is not retained for further evaluation.

### 3.2.4 pCu

As discussed in Section 1.3, the results of the 5-year pH Monitoring Study showed that the 2008 white rain benefitted the STSIU soils by increasing the pH and pCu of the acidic soils, making copper less bioavailable. The increase in pCu led to a decrease in the uptake of copper into living organisms, decreased wildlife dietary exposure, and significantly improved plant richness but not plant cover. The potential of STSIU soils to generate acid was found to be consistently low after the white rain through 2014 in most areas, although the long-term persistence of this effect is unknown (Arcadis 2017 in Appendix B). The white rain improved richness on the untreated test plots but did not significantly improve cover (Arcadis 2017 in Appendix A). These changes needed to be considered when estimating the current nature and extent of pCu on the landscape.

Cupric ion activity (pCu) is difficult to measure directly in soil. It was directly measured in 0-6" bgs soil at 17 sites in the STSIU in 1999 in the laboratory and was found to be strongly related ( $R^2 = 0.97$ ) to copper concentrations and pH in the following equation (upland equation from Sitewide ERA, NewFields 2006b):

$$pCu = 7.34 + (0.93 * pH) - (1.15 * \ln[Cu_{total}])$$

Because it did not require measuring copper activity in solutions, this equation was applied to available total copper and pH data in the soil. The soil was sampled at 0-6" bgs and sieved to < 2 mm and then the spatial distribution of pCu was estimated using an interpolation method (Table 3-4, methods discussed in detail in Section 6 of Appendix D). Only post-white rain samples were used in impacted areas because they best represent current conditions, and remedial decisions should be based on current chemistry of the soils. Of 155 pCu locations, 102 were sampled in the STSIU after the white rain event. However, these 102 samples did not fully cover the outer edges of the STSIU needed to bound the interpolation of pCu across the landscape (prevent the interpolation from erroneously assigning low pCu values on the edges where pre-white rain samples show they are high pCu). Therefore, 56 pre-white rain samples located on the edges were included (see Figure 3-7 and Table 3-4). These 56 samples, though collected pre-white rain, are likely representative of current conditions because they are at the edges of the interpolation where the effect of the lime in the white rain was minimal because they were already well-buffered soils with high pCu, essentially at background levels (all pre-white rain samples used to bound the post-white rain samples had pCu > 5 with the exception of a few locations directly north of Hurley and just east of Tailing Pond 7). This approach is conservative in areas where liming possibly had a small beneficial effect in the STSIU edges. The impacted areas in the non-edge areas that have low buffering capacity and thus benefitted from the white rain were always represented only by the post-white rain samples to avoid recommending remediation in areas that already have been treated with alkalinity from the white rain (had increased pCu above the pre-FS RAC). Areas not included for remediation because the white rain improved pCu will be monitored in the future to ensure the white rain effect was sustained.

Using the above calculated pCu dataset *before* incorporating effects of interim remedial actions, pCu contours and a grid of pCu values across the STSIU were developed using natural neighbor interpolation as shown in Figure 3-8 (see Appendix D, Section 6.2 for details on interpolation method). IRA areas were then converted to pCu of 6, representing a conservative background value after IRAs were completed. As proposed in the FS Proposal, Chino selected the pCu exposure units to be the polygon boundaries defined by combinations of different soil and vegetation types. These are the same boundaries that were used for the rangeland condition analysis in Woodward Clyde (1997) and thus are referred to as “rangeland polygons” (Figure 3-9). The pCu values in the grid that intersected with the rangeland polygons were averaged to obtain the exposure point concentration (EPC) for each polygon.

The process of identifying rangeland polygons considered for remediation (destroys plants initially) to protect plants from pCu exposure was a sequence of steps that not only considered the pCu pre-FS RAC but also considered the current condition of the plant community relative to reference areas in the same soil category and relative to toxicity of a soil category. Plant sensitivity to pCu varies by four soil categories as discussed in the phytotoxicity and community study in Appendix C. The four soil/topography categories (referred to as soil categories) and vegetation survey and associated soil sampling locations from the study are shown in Figure 3-10. They include flat rocky, flat granular, slope > 13%, and bedrock soils.

The sequence of steps for identifying polygons considered for remediation is outlined as follows:

1. Identify rangeland polygons that fail the pre-FS RAC, with mean pCu EPC < 5 in areas with copper > 327 (identified in Figure 3-11 after IRAs). This pre-FS RAC, developed by NMED, is a generic pre-FS RAC applied regardless of soil category.
2. Identify rangeland polygons post-IRA that failed the generic pre-FS RAC that also have unacceptable wildlife habitat or rangeland habitat (Figure 3-12).
3. Identify remaining rangeland polygons that also failed a pre-FS RAC refined to be soil-category specific (Figure 3-13).

The area failing the pre-FS RAC for pCu in step 1, after IRA and borrow area removals, covers 2,459 acres in the STSIU (Figure 3-11). However, the pre-FS RAC to protect plants did not take into consideration whether the plant community in an area with pCu < 5 was of similar habitat quality to reference areas with high pCu. As such, the FS Work Plan outlined an approach for step 2 below for comparing areas to reference and excluding them from consideration for remediation if they have similar characteristics to reference areas, where the characteristics evaluated are important to wildlife and livestock.

Only rangeland polygons that intersected average pCu < 5 and copper > 327 mg/kg (i.e., below the Pre-FS RAC) that failed the criteria outlined below for rangeland and wildlife habitat quality were retained for further remedial evaluation. Rangeland was quantified using the OAT score, and wildlife habitat quality was quantified based on richness and cover of the vegetation community. The OAT method is a rapid assessment technique promoted by the Bureau of Land Management and Natural Resources Conservation Service whereby the investigator walks through a defined area and visually estimates scores based on vegetation vigor and soil condition (explained in detail in Appendix D, Section 5.2). The method was used to estimate “apparent” trend in rangeland condition without sampling more than one time period. A high score represents good rangeland condition.

The criteria outlined below based on OAT score, richness and cover, remove rangeland polygons from remedial consideration where the destruction of the existing vegetation and inevitable increase in soil erosion associated with remediation could lead to a loss of long-term environmental benefits to the vegetation, causing more harm than good.

- Except for bedrock, if the OAT score of the rangeland polygon was  $\geq 22$  for all soil categories, the polygon was considered to have “fair-good” rangeland condition (acceptable condition) and, therefore, was excluded from further evaluation. If the OAT score was  $< 22$  (or  $<$  than 13 for bedrock, see Section 9.3 of Appendix D for details), the polygon’s rangeland condition was considered “poor” and it was retained for further evaluation. The cutoff of 22 was based upon data in 1997 rangeland condition datasheets (see Woodward Clyde 1997) showing most rangeland polygons had fair to good condition with  $OAT \geq 22$  on STSIU, as described in the approved FS Workplan (Arcadis 2011a). The 2011 OAT scores (see Section 5.2 of Appendix D) were used to determine current rangeland condition for this FS.
- If the percent vegetation cover of a rangeland polygon was  $\geq$  a targeted percentage of a reference area mean, the polygon was considered to have “acceptable” cover for wildlife habitat. The targeted percentage was based on variability of the background value of the soil category; the less variable, the higher the target was (closer to the mean value, see Section 9.3 of Appendix D for details on target development). Flat granular soil, flat rocky soil, bedrock ( $> 60\%$ ), and steep slopes ( $>13\%$  slope) are the four soil categories with different targeted percentages. Unacceptable polygons based on cover were retained for further evaluation.
- Similar to vegetation cover, if the vegetation species richness of a rangeland polygon was  $\geq$  a targeted percentage of a reference area mean, the polygon was considered to have “acceptable” richness for wildlife habitat. Unacceptable polygons based on richness were retained for further evaluation.

OAT scores, percent vegetation cover, and species richness for each polygon were obtained using binary classification techniques on remote-sensing-based maps (Section 7 of Appendix D provides more detail on remote sensing methods to create the maps). The maps were developed from IKONOS satellite imagery that covered the site on September 4, 2011. These maps were classified into two categories: acceptable and unacceptable, as discussed above. Those classification results were then ground-truthed and evaluated for accuracy using the percent cover and species richness that were sampled in the field on 100 ft<sup>2</sup> plots and the OAT score data collected on 200-m transects that were alongside the plots in 2011, 2014, or 2018 (plots are shown in Figure 3-10, see Section 5.2 of Appendix D for details). All field sample locations had been grazed, as almost all of the STSIU has been grazed and, therefore, the criteria for acceptability was based on comparison to grazed reference plots.<sup>8</sup>

All plots with field data for vegetation cover and species richness were used to evaluate map accuracy because the plot locations were independent of the high and low value endpoints used to calibrate the vegetative cover and richness remote-sensing models (calibration of cover and richness is explained in Section 7 of Appendix D). However, for rangeland condition (OAT score), 3/4 of the ground samples were used for training to develop a good classification model and 1/4 of the samples were set aside to be independent and used for accuracy assessment (see Section 7 of Appendix D). The accuracy of the OAT score, vegetation cover, and species richness maps in identifying the two classes (acceptable vs. unacceptable) based on locations independent of the

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<sup>8</sup> Based on the FS work plan that specified grazed plots will be evaluated for habitat quality by comparing to reference plots. The initial reference plots included one north- and one south-facing grazed “wildlife reference” plot (Figure 3-10). However, the “wildlife reference plot south” was downgraded to a site plot with “de minimis” impacts after reviewing its pCu estimate when chemistry was later collected (see Appendix C and D) and subsequently was not used as a reference plot. Eight additional reference plots, representing the four major soil categories on the site were sampled in 2018 and, along with previously collected data on two other reference plots (STS-PT-2013-26 and wildlife reference plot north in Figure 3-10) were used to estimate target thresholds, as described in Section 5.3 in Appendix D.

model data was relatively good at 88%, 74%, and 71%, respectively, meeting the FS workplan goal of at least 70% accuracy (Tables 3-5, 3-6, and Section 7.2 in Appendix D).

For the final classification of a rangeland polygon as containing vegetation that may have been degraded in quality by pCu, the polygon had to be rated unacceptable for the OAT score (rangeland condition) or unacceptable for either richness or cover (wildlife habitat quality). The retained rangeland polygons classified as unacceptable (overlaid on the four soil categories) with pCu lower than the pre-FS RAC of 5 where copper is > 327 mg/kg are shown on Figure 3-12. The mapped acres in Figure 3-12 with pCu < 5 that had greater than 327 mg/kg copper that were unacceptable were 230 acres of flat granular areas, 1,097 acres of bedrock, 146 acres of steep sloped areas (steep slope defined as > 13%), and 346 acres of flat rocky soil areas (total of 1819 acres).

The generic pre-FS RAC did not consider sensitivity of a soil category in the field to pCu toxicity. The generic pre-FS RAC is unrefined because it is equivalent to the estimated PEL in the 1999 Site-wide ERA study of toxicity of pCu to growth endpoints of grasses and forbs using greenhouse data and site soils in pots as well as field-based plant community data. The site soils selected in that study represent a narrow range of habitat types, mainly on level or very shallow sloped areas. Much of the STSIU is topographically rugged with varying soil conditions and types not represented by the soils studied in the 1999 ERA. Thus, in 2012, Chino submitted a work plan to further investigate native seeds and soil type in a phytotoxicity study which NMED subsequently approved in 2014 (Arcadis 2014a). NMED agreed with the definition that specifies use of the effects concentration at 50% (EC50) for the PEL based on their comments and Chino's responses, as follows:

- The September 2010 pre-FS RAC letter indicates that the PEL in the Site Wide ERA (Table 2.2-1) was selected as the pre-FS RAC for pCu because the ecological significance of effects between the DEL and PEL is uncertain.
- The phytotoxicity study in Appendix C fit curves to 1999 ERA data (Figures 10 and 11 in Appendix C) that demonstrate that the 1999 PELs in the sitewide ERA upon which the pre-FS RAC was based, approximate an EC50 (e.g., for richness, cover, and shoot weight/height), not a lowest-observed-adverse-effect level.
- An EC50 definition is consistent with USEPA's use of the term "probable effects level" for aquatic sediment toxicity assessments. USEPA endorses using PELs for aquatic toxicity thresholds because they are better at predicting toxicity than ERLs or TELs (effects range low and threshold effect levels similar to DELs), as stated in Long et al. (1998): "the probabilities of highly toxic responses occurring when one or more ERLs or TELs are exceeded and no ERMs or PELs are exceeded are 16 to 18% in amphipod tests alone.... The probabilities of either marginally or highly toxic responses occurring are 48 to 52%...when concentrations exceed one or more ERMs or PELs; consistent with their original intent, the ERMs and PELs are considerably better at predicting toxicity than are the ERLs and TELs." When used for remedial decisions, the threshold should be the more predictive one to avoid causing more harm with habitat destruction than good with reducing COPC concentrations.
- Protection of plants requires different criteria in the southwest than for wildlife because the plants are directly destroyed in the remediation, whether by soil removal or tilling. The destruction of the plants "to benefit plants" and long recovery period required for the entire plant community (decades, see Appendix B-3 in Appendix A) in the semi-arid southwest outweighs the impact of a "low" (DEL) reduction in emergence or survival, difficult to even observe in the field. A reduction at a threshold that creates a more probable effect (PEL), above which a more certain effect occurs, is needed before deciding to destroy the very same species targeted for protection. Protectiveness of the plants as a food source is an important criteria to consider in the FS, which may be lost until the community recovers from remediation.



This approved phytotoxicity and community study was conducted in the field and greenhouse in 2014 and 2015, and all soil categories were evaluated. The soil category was found to change the PEL, and the PEL for each community endpoint (richness, cover, and OAT score) significantly differed among the four soil categories (Section 4.2.1 and Figure 12 in Appendix C). Based on NMED comments on the draft phytotoxicity study, reference soils for each category were also surveyed and sampled in 2018 in order to strengthen the study conclusions (Attachment A of Appendix D). Specifically, while Chino had worked with NMED in September 2012 to compare bedrock areas within the STSIU to an agreed upon reference bedrock area outside of the STSIU, NMED later deemed the area selected to not to be an adequate bedrock reference area because it was found to have pH lower than background (Section 4 in Appendix D). A reference area study was initiated in 2018 to identify adequate reference areas east of the former smelter that would represent all four soil categories, including bedrock (see Attachment A of Appendix D). The reference soils were assumed to represent no toxic effect and the PEL approximated a 50% effect level relative to the reference (following the general approach from the 1999 ERA work).

Because vegetation response to pCu varied by soil category (Section 4.2.1 in Appendix C) and its buffering capacity (Section 9.3 in Appendix D), the retained rangeland polygons after step 2 were further evaluated in step 3 as to whether their average pCu exceeded their respective PEL based on the polygon's soil category, as shown in Table 3-7 and described in Section 10 of Appendix D. For this FS, the PELs were re-calculated from the field community data collected in 2018. The PELs were 4.40, 4.98, 4.11, and 4.11 for bedrock, flat rocky, flat granular, and slope > 13% soil categories (see Section 9.3 in Appendix D for details on methods, data and calculations of PELs) based on selecting the highest PEL for either wildlife habitat (represented by average of richness and cover PELs) or rangeland condition (OAT score).<sup>9</sup> Because vegetation response to pCu varies by soil category (Section 4.2.1 in Appendix C) and its buffering capacity (Section 9.3 in Appendix D), the retained rangeland polygons after step 2 were further evaluated in step 3 as to whether their average pCu exceeded their PEL based on the polygon's soil category, as shown in Table 3-7 and described in Section 10 of Appendix D. If a retained rangeland polygon had an average pCu  $\geq$  its PEL, it was removed from further analysis.

The pCu map in Figure 3-8 represents the post-white rain conditions before interim actions and soil removals described in Section 3.1.2 occurred. The map does not represent current conditions after interim actions. Therefore, areas remediated, reclaimed or with WBT or borrow material removal (shown in Figure 3-13, referred to generically as "IRA" areas) were assumed to have pCu above the pre-FS RAC of 5 because depths over one to 12 feet were excavated (copper and pH effects are in shallower layers). Also, some operational areas that had clean soil such as Bolton Draw diversion were also considered "remediated" and removed. These IRA areas were assigned a pCu of 6 (conservatively representing background) before averaging pCu across map pixels within each rangeland polygon to create a map and table of pCu EPCs by polygon (see Table D-2 in Appendix D for all polygon EPCs in the STSIU). Using the soil category-specific PELs calculated from that reference data, approximately 47, 313, and 17 acres of bedrock, flat rocky, and slope >13%, respectively are retained to consider for remediation (Figure 3-13). All acres of flat granular soils were below the flat granular PEL of 4.11 and thus none in this category were retained.

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<sup>9</sup> Basing the calculation on the mean of all three endpoints of richness, cover, and OAT score is a second way one could calculate the PEL that produces lower PELs (see Appendix D, Section 10.2).

### 3.3 Soil Exposure Units to be Evaluated for Remedial Alternatives

Before evaluating technologies and remedial alternatives, this FS evaluates all exposure units where (1) copper in soil is greater than 1600 mg/kg, (2) copper in soil is within the monitoring range of 1100 – 1600 mg/kg, and (3) pCu in soil fails the pre-FS RAC, acceptability criteria, and PEL comparison for each soil category. All exposure units had a 95UCL area-weighted average for copper less than or equal to the Pre-FS RAC of 1,600 mg/kg (ratios  $\leq 1.0$  in Table 3-2), but three units (alliance polygons 1-3, 10-16 and 88-15) summing to 140 acres fell in the copper range for monitoring (Figure 3-5). Nine exposure units summing to 377 acres from the pCu evaluation are retained for consideration for remediation (Figure 3-13).

## 4. Description of Current Situation: Surface Water

### 4.1 Conceptual Site Model

As discussed in the AOC Background Report (Chino 1995) and the STSIU RI (SRK 2008a,b), the surface waters<sup>10</sup> and drainages in the STSIU include Bolton Draw (called Drainage C-2 in this FS), Rustler Canyon, Martin Canyon, Lucky Bill Canyon, and other un-named tributaries of Whitewater Creek (referred to by letters) and Lampbright Draw. Previous investigations including the RI and ERA have noted the majority of these drainages are considered ephemeral (Figure 4-2), with surface water flow generally occurring during and immediately after a precipitation event. In 2011, Arcadis applied NMED Surface Water Quality Bureau's (SWQB's) Hydrology Protocol to determine the appropriate hydrologic regime of STSIU surface water drainages. Hydrologic classifications of STSIU drainages were established in the Application of the Hydrology Protocol to STSIU Drainages report (Arcadis 2014b). Non-ephemeral drainages include Rustler Canyon, Martin Canyon, Lucky Bill Canyon, headwaters of Drainage C (upgradient of Bolton Canyon confluence), southeastern portion of Drainage D-1, northern drainage of Lucky Bill Canyon, and the drainage immediately downstream of Ash Springs to confluence with Drainage A. All other STSIU drainage areas are designated as ephemeral. Other than these drainages, the remaining surface water bodies present in the STSIU are livestock watering tanks. The livestock tanks are generally designed to catch surface water run-off and/or spring or well water, and to hold water for extended periods of time.

#### ***Fate and Transport of COCs in STSIU Drainages***

As discussed in the FS proposal, the primary sources of surface water impacts in the STSIU drainages are due to the presence of COCs in surface water runoff, leaching of COCs from sediments (see conceptual model, Figure 3-1), and possibly re-emergence of subsurface alluvial water.

#### Surface Water Run-off

For the STSIU drainages, runoff from the upland environment is considered to be a source of potential COC impacts to surface water. COCs from COC-affected soils, from either smelter emissions or fugitive dust, have the potential to be sorbed to/incorporated into suspended particles or dissolved in runoff during precipitation events, leading to increased total and dissolved surface-water concentrations of COCs in the receiving drainages.

#### Leaching from Sediments

STSIU surface water also has the potential to be affected by the remobilization of COCs from sediments in drainages during stream flow events. Constituents in COC-affected sediments, derived from smelter emissions, historical upland runoff, or fugitive dust, may be remobilized during precipitation events, leading to possible increased total and dissolved surface water COC concentrations in the receiving drainages. In addition, leaching of COCs from COC-affected sediment may occur during times when water is present in a drainage after a storm event.

The leaching potential of sediments in STSIU drainages was evaluated for samples collected at select locations within the STSIU during a December 2010 sampling event, and the results were reported in the *Groundwater Quality Pre-Feasibility Study Remedial Action Criteria for Drainage Sediments* (Arcadis 2011b). These results

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<sup>10</sup> The term "surface waters" is used in this document to reflect the common meaning of the term, not any legal meaning. Use of this term is not intended to imply that there has been a determination that any particular surface water or drainage constitutes a "surface water of the state" or "waters of the United States."

demonstrate that, with the exception of copper, sediments in the STSIU drainages do not have the potential to leach COCs to surface water at concentrations that might adversely affect the surface water quality at concentrations above acute aquatic life criteria. NMED approved the report and indicated that sediment does not need to be further evaluated in the FS (NMED 2010b).

#### Re-emergence of Subsurface Alluvial Water

There are a few locations in the STSIU where subsurface alluvial water is re-expressed as surface water via seeps. If this re-expressed groundwater has COC concentrations that are greater than the NM Water Quality Standards (WQS), then there is a potential that this subsurface alluvial water could have implications on how to manage surface water.

The FS proposal stated that it is unclear which of these three fate and transport mechanisms is driving metals – and specifically copper – loading to STSIU drainages. Appendix E discusses use of stormwater samplers, which were intended to sample run-off to allocate metal loads in surface water to upgradient sources, soil sources, or legacy sediment sources. However, the samplers entrained sediment in the samples not representative of surface water, and the amount of time after storm samples were in the bottle before samples were retrieved was uncertain. These issues led to the study results being inconclusive, and the results do not rule out slope runoff as a primary source of COCs.

#### Summary

The primary mechanisms for loading of COCs to surface water in STSIU drainages is transport of metals in hill slope runoff (dissolved COCs as well as COCs associated with soil/sediment) derived from COC-affected soil in the STSIU. Since sediments in drainages are derived from hill slopes above the drainage, the presence of COCs in hill slope runoff, surface water, and sediments are intimately linked.

#### ***Fate and Transport of COCs in Stock Tanks***

Stock tanks that receive the majority of their water from surface runoff are a key focus for the CSM because the COC sources described above for surface water will affect the quality of water in the stock tanks. Stock tanks might also receive water from springs or windmill-pumped groundwater, but these sources do not have a large surface-water runoff component. The water quality in stock tanks may also be affected by the potential regeneration of COCs stored in stock tank sediments.

#### Surface Water Inflow

Stock tanks receive their water from ephemeral surface drainages, which in turn likely receive the majority of their water from surface runoff. As stated above, there is potential for COCs in soils to be sorbed to/incorporated into suspended particles or dissolved in runoff during precipitation events. In addition, leaching of COCs from drainage sediments and mobilization of COC-affected sediments may occur during the brief period of time when stream flow occurs. These mechanisms may lead to input of dissolved and solids-associated COCs to the stock tanks during runoff events.

#### Cycling of COCs from Sediments

For stock tanks, the cycling of COCs from sediment to surface water may present a continued source of COC impacts to stock tank surface water. If there is no outflow from a stock tank, all COC-affected sediments transported to a stock tank during a runoff event will accumulate in the stock tank. These COC-affected sediments may present a long-term COC source for stock tank surface water. When a new inflow event occurs, the already present mass of COCs in the stock tank is increased by the constituents present in the incoming surface water.

The combination of COC-affected surface water and the remobilization of COCs from COC-affected sediments in the stock tanks may result in the potential for COCs in stock tank water to exceed surface water criteria.

### Summary

The source of the majority of the mass of COCs in stock tank surface water is unknown but likely is derived from inflow during runoff events. Cycling of COCs from COC-affected sediments that accumulate in the stock tanks may result in release of additional COCs to surface water in the stock tanks.

### ***Surface Water Conceptual Site Model***

STSIU surface water, contained in drainages and stock tanks, has the potential to be affected by runoff interacting with COC-affected surface soil and sediment during storm events, and ongoing leaching of COCs from COC-affected sediments during times when water remains within a drainage or stock tank.

The CSM for STSIU surface waters (Figure 3-1) identified the following potential exposure pathways for current and future humans associated with COC-impacted surface water:

- Direct Contact: Exposure associated with direct dermal contact with affected surface water;
- Ingestion: Exposure associated with incidental ingestion of affected surface water, and COCs from affected surface water might bioaccumulate in foods that could be consumed by grazing animals. COCs could, in turn, be ingested through consumption of these animals.

The potential exposure routes for current and future ecological receptors include;

- Direct contact (dermal contact, aquatic organism contact with water);
- Ingestion (drinking water and diet-borne exposure of food sources drinking water).

The above potential human health and ecological exposure routes were evaluated in the STSIU (Gradient 2008) and ERA (NewFields 2006a, 2008).

## **4.1.1 Previous Investigations**

### ***Remedial Investigation***

The STSIU RI was completed in 2004, and then revised with the inclusion of new data and resubmitted in 2008 (SRK 2004; SRK 2008a). Twenty-four surface water samples were collected at 16 locations in the STSIU and were used to evaluate surface water quality during the RI (SRK 2008a). Those results were presented in the Addendum to the RI (SRK 2008b).

During the RI sampling, aluminium, barium, beryllium, boron, cadmium, calcium, cobalt, copper, iron, lead, manganese, molybdenum, nickel, vanadium, and zinc were detected in STSIU surface waters. Of these detected constituents, only cadmium, copper, and lead were detected at concentrations that exceeded the chronic aquatic life criteria at that time (SRK 2008b). Lead concentrations were not detected above acute aquatic life criteria. Cadmium concentrations exceeded acute criteria in three drainage locations and two stock tanks. Copper concentrations were consistently above acute criteria in drainage and stock pond locations.

### ***Human Health Risk Assessment***

The HHRA (Gradient 2008) evaluated data generated from surface water samples collected within the STSIU. The risk assessment calculated cancer risks and non-cancer hazards for potential receptors on the site. The methodology used for this risk assessment was consistent with USEPA guidelines and used conservative, default

assumptions whenever site-specific data were not available. Gradient evaluated two dermal contact scenarios, the recreational swimmer and the trespasser swimmer in stock ponds, to determine the potential risk from exposure to STSIU surface water. They also evaluated incidental ingestion of water during swimming.

Gradient determined that site exposure to surface water did not result in unacceptable cancer risks or non-cancer risks for either the recreational swimmer or the trespasser swimmer scenarios for any COCs identified for the STSIU.

### ***Ecological Risk Assessment***

The ERA (NewFields 2008) evaluated the risks from surface water to aquatic receptors in the STSIU.<sup>11</sup> The methodology used for this risk assessment was consistent with USEPA guidelines and used conservative, default assumptions, whenever site-specific data were not available. The receptors evaluated in NewFields' ERA included amphibians, aquatic invertebrate communities, and fish communities.

The ERA showed that where surface water exists in the STSIU, the COCs of most concern were cadmium, copper, and lead (NewFields 2008). These three constituents exceed the chronic aquatic life NM WQS and/or the selected amphibian TRVs (Harfenist et al. 1989; Schafer and Associates 1999a) at one or more RI surface water locations.

NewFields (2008) concluded that where water is present in the STSIU, copper concentrations are elevated above acute and chronic water quality criteria at most locations. Risks to aquatic life from copper in surface water are predicted for the limited aquatic habitat within the STSIU. However, the risks to aquatic receptors from chemical exposure need to be qualified given the overall physical quality of the aquatic habitat. Without persistent aquatic habitat in the area, aquatic life is limited to invertebrate species that breed relatively rapidly in water and to potentially breeding amphibians and their larval offspring, particularly in the stock tanks. Fish have not been observed in any surface water features associated with the STSIU, except in Rustler Canyon during a 2007 sampling event (NewFields 2008). Therefore, as stated by NewFields (2008), the risk estimations presented for aquatic life are highly uncertain.

### ***Remedial Investigation Technical Memorandum***

The STSIU RI Surface Water Sampling and Analysis of Rainfall Pools Addendum was completed in 2008 (Chino 2008). Seventeen surface water samples were collected at surface water drainage and stock tank locations in the STSIU. These samples were used to evaluate occurrence of select metals (aluminium, cadmium, copper, iron, lead, and manganese) within different size fractions based on filter pore size and to assess potential associations of metals with dissolved organic matter.

During the RI Addendum sampling, all of the analysed metals were detected in surface water samples, but only aluminium, cadmium, and copper were detected at concentrations that exceeded the chronic aquatic life criteria (Chino 2008). Aluminium and cadmium were primarily associated with particulate form and were mostly excluded by filtration with 0.45 micron filter pore size. Concentrations of copper generally decreased with decreasing filter size fraction. However, ultra-filtered (~0.001 micron effective pore size) copper concentrations exceeded aquatic life criteria for about one half of the locations. Copper concentrations were positively correlated with dissolved organic carbon (DOC) concentrations, suggesting that copper may be bound with DOC and thus may have a

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<sup>11</sup> Risk was higher from surface water than sediments (NewFields 2008). Sediment exposure is deemed a relatively insignificant pathway with no pre-FS RAC so not addressed in the FS (NMED 2011b).

lower toxicity. Based on these results, a more formalized surface water toxicity study was undertaken, as described in the following section.

### **Water Effect Ratio Study**

As discussed in Section 1.3, Arcadis conducted a copper WER study for the STSIU surface waters in 2011 and results from the WER study were described in the *Development of Site-Specific Copper Criteria Interim Report* submitted to NMED in March 2013 (Arcadis 2013a). WER study locations are shown in Figure 4-1. A site-specific copper WER model was subsequently developed to derive adjusted copper criteria in STSIU surface waters in the *Revised Site-Specific Copper Toxicity Model Report* submitted to NMED in October 2013 (Arcadis 2013b). The site-specific criteria for STSIU surface waters were adopted by NMED and are contained in NMAC §20.6.4.808 - 809.

### **Hydrology Protocol Use-Attainability Analysis**

In 2011, Arcadis conducted a Use Attainability Analysis (UAA) based on NMED SWQB's Hydrology Protocol to determine the appropriate hydrologic regime of STSIU surface water drainages. Hydrologic classifications of STSIU drainages were proposed in the draft *Application of the Hydrology Protocol to STSIU Drainages* report submitted in October 2012 (Arcadis 2012). The report was finalized in 2014 after review and comment from the New Mexico Surface Water Quality Control Board and field reconnaissance by the Ground Water Quality Control Board as well as by NMED (Arcadis 2014b). The hydrologic classifications of STSIU drainages are depicted in Figure 4-2. Non-ephemeral drainages identified in the Hydrology Protocol UAA include those in Rustler Canyon, Martin Canyon, headwaters of C-1 (Bolton Canyon), southeastern portion of Drainage D-1, headwaters of C-2 (Bolton Draw), and drainage immediately downstream of Ash Springs to confluence with Drainage A. All other STSIU drainage areas in the UAA study area<sup>12</sup> are designated as ephemeral and were treated as ephemeral in this FS. Conclusions regarding use attainability analysis were promulgated by NMED, see NMAC §20.6.4.808 for list of intermittent/perennial drainages and §20.6.4.809 for list of ephemeral drainages.

### **2010, 2011, and 2013 Wet Season Surveys**

Before the WER study work plan was developed, a preliminary 2010 Wet Season Survey was conducted to gain a better understanding of water persistence, copper toxicity, and water chemistry variability in the STSIU surface water drainages (Arcadis 2013a). A total of 12 STSIU surface water samples were collected in the 2010 study. Surface water samples collected were analyzed for a complete set of water chemistry affecting metal bioavailability, including parameters such as total and dissolved organic carbon, alkalinity, and total dissolved solids. Previous site investigations included only a limited number of chemical parameters, which precluded estimation of site-specific metal toxicity using predictive models such as the biotic ligand model. In 2011, in addition to the WER study sampling, five samples were collected to increase the spatial distribution of these full set chemistry samples in the STSIU (no toxicity tests were performed on the water at those locations). In August 2013, a year with more rainfall, additional full set chemistry sampling was conducted to further understand the variability in surface water chemistry with the collection of 13 more samples. In Appendix E of Arcadis (2013b), 48 locations were evaluated (including the WER locations), with the purpose of assessing whether the chemistry range used to develop the WER model sufficiently represented the range of water chemistries in the STSIU study area (2010 to 2013 data in Tables 4-1 to 4-3). The analysis showed that the ranges of chemistry parameters used to develop the WER model were representative of STSIU surface waters.

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<sup>12</sup> Lucky Bill Canyon was not in the UAA study area.

### ***Surface Water Runoff Quality and Duration Sampling***

Appendix E documents the collection and analysis of surface water runoff samples proposed in the STSIU FS Proposal (Arcadis 2011a). The objectives of the surface water runoff investigation included providing additional surface water quality data to support refinement of the conceptual site model and to support the STSIU FS. The objectives of the sampling also included defining the duration of flow and presence of water to support classification of drainage channels in the STSIU. However, the data associated with the duration of flow and presence of water were used in the expedited UAA described above and are thus not included in Appendix E.

As described in Appendix E, both dissolved and total concentrations of metals in stormwater samples were typically higher in samples collected from stormwater samplers installed at higher elevations above the creek channel. This may be due to a longer contact time between runoff and COC-containing soil and sediment or may be due to greater entrainment of COC-containing sediments at higher flows. Concentrations of total metals were generally substantially higher than concentrations of dissolved metals because most samplers contained substantial quantities of sediment entrained within the samplers and the sample bottles at the time of sample retrieval and the presence of these sediments may have resulted in elevated concentrations of total metals in the stormwater samples. These sediments coupled with the uncertainty regarding the amount of time samples were in the sample bottles prior to retrieval introduced uncertainty in the quality of the data. Because of this, the data in Appendix E were evaluated qualitatively to guide the refinement of the conceptual site model.

## **4.1.2 Other Activities**

### ***Rangeland Improvements***

Chino and their lessees periodically perform maintenance activities on the rangelands lying within site boundaries, including removing earthen stock tanks that are no longer required either because they are no longer used or because they are redundant due to other stock tanks located nearby. As part of these rangeland improvements, stock tanks 15, 26, 29, and 60 were removed between 2013 and 2022 and were not replaced.

Additionally, the approximately 230-acre drainage basin feeding stock tank 06 was improved in 2023 to provide for stable stormwater conveyance to the tank inlet. The drainage channel upgradient of the tank has been improved for approximately ½ mile, beginning at stock tank 06 and extending northeast to the railroad crossing immediately west of the intersection of North Hurley Road and A Street. During these improvements, material was identified along the railroad track that could act as a source of copper to the drainage feeding stock tank 06; this material was removed, and the area was regraded. In addition, five check dams were installed along the drainage to slow stormwater flows entering stock tank 06. These improvements will achieve their desired effect, and an improvement to water quality is also expected.

## **4.2 Nature and Extent of Contamination**

All surface water data in the STSIU were compiled and compared to current New Mexico Water Quality Control Commission (NMWQCC) aquatic life criteria for the constituents identified as a concern in the 2008 RI (SRK 2008a,b) and ERA (Newfields 2008). The constituents of concern in the RI addendum included copper, cadmium, and lead after accounting for filtration at 0.45 and 0.001-micron size fraction to ensure dissolved concentrations were evaluated (SRK 2008b). Aluminum is compared to criteria based on total concentrations as required by current NMWQCC criteria and thus, even though aluminum is not known to be detected in historic smelter emissions, was included to be complete. Nickel and silver were mentioned as a possible concern for aquatic life



in SRK (2008a) and thus were also included. Tables 4-1 to 4-3 (COCs other than copper) and 4-5 to 4-7 (copper results) present the results and comparisons to criteria (locations shown in Figure 4-3), expressed as hazard quotients (HQ) for ephemeral drainages, non-ephemeral drainages and stock tanks. The HQ is the observed concentration divided by the hardness-adjusted criteria concentration. Exceedances (HQ > 1) of chronic criteria for one or more locations occurred for aluminum, cadmium, copper, and lead for both ephemeral and non-ephemeral drainages. Aluminum, cadmium and copper exceeded acute criteria for ephemeral drainages, and only copper and aluminum exceeded acute criteria for non-ephemeral drainages. For stock tanks, aluminum, copper and lead exceeded chronic criteria and aluminum and copper exceeded acute criteria. Aluminum, however, is not sourced from historical emissions (SRK 2008b) and is not discussed further. Only acute criteria are applicable to (1) ephemeral drainages listed in Tables 4-1 and 4-5 and (2) stock tanks, which are fed by ephemeral drainages shown in Table 4-3 and 4-7 (see NMAC §20.6.4.808 – 809). Because the exceedances are insignificant when compared to those for copper, and that any measure taken to address copper in STSIU surface water would also address the exceedances of these metals, copper is the constituent of focus for STSIU surface water.

As described in Section 4.1, additional STSIU surface-water studies were completed after the RI Addendum report was accepted. These studies include the Wet Season Survey conducted in 2010 - 2013, the Criteria Adjustment Study conducted in 2011, and the Expedited UAA Study and Surface Water Runoff Quality and Duration Study conducted in 2011. Metal concentration data collected during these studies are included in this nature-and-extent section. Copper WERs calculated in accordance with NMAC §20.6.4.808/809<sup>13</sup> are also considered herein for evaluating the nature and extent of copper contamination. Based on hydrologic classifications shown in Figure 4-2, chronic aquatic life criteria are only considered for the non-ephemeral drainages; whereas, acute criteria are considered applicable to the ephemeral drainages.

## 4.2.1 Copper – Surface Water Drainages

Most ephemeral surface water drainage locations in the STSIU (Figure 4-2) exceed the standard hardness-adjusted NMWQCC acute aquatic life standards for copper (Table 4-5). Two WER methods were applied to adjust the criteria to more site-specific conditions: the watershed-specific WERs and the geometric mean for the STSIU WER (developed from data in Table 4-4). Acute criteria, not chronic criteria, apply to the ephemeral drainages listed in Table 4-5 (see NMAC §20.6.4.808 – 809; chronic hardness-adjusted criteria shown for completeness). Fewer ephemeral locations exceeded the site-specific WER-based acute criteria for copper than before applying the WER, and most had HQ < 2. Those with HQ > 1 comprised only one sample of many in subdrainages with exceedances, except Drainage D3. Drainage D3 had all acute criteria HQs up to 5 for the more site-specific Acute criteria WER-based HQs (Table 4-5). Ephemeral drainages B, D-1, G, and Lampbright Draw had HQs < 1, whereas Drainages A, C1, C2, D2, and D3 had at least one sample with HQ > 1. Drainages with HQ > 1 are depicted on Figure 4-3 using dashed lines. The WER-based Chronic HQs that were > 1 on non-ephemeral streams were uncommon, with only one sample (of 11) in Martin Canyon (but later years showed HQ in the same area was < 1 in Martin Canyon so not of concern), one sample (of 12) in Rustler Canyon (SW05-2007 in Table 4-6 had a very high HQ but should not be included because a slope failure caused the high copper temporarily at that location in 2007), one sample in headwaters of subwatershed C-1, and none in Lucky Bill Canyon exceeding

<sup>13</sup> Specifies criteria in NMAC §20.6.4.900 will be multiplied by a site-specific WER value calculated from the equation:

$$\text{WER} = \frac{10^{0.588 + (0.0703 \times \log \text{DOC}) + (0.395 \times \log \text{Alkalinity}) \left(\frac{100}{\text{Hardness}}\right)^{0.9422}}}{19.31}, \text{ where DOC} = \text{dissolved organic carbon.}$$

criteria (Table 4-6). The more general WER-based HQs (last column in Tables 4-5 and 4-6) generally give similar results for Tables 4-5 and 4-6.

Figure 4-3 shows the nature and extent of copper contamination in the drainages based on these WER results. Areas of no concern (solid lines) and potential concern (dashed lines) are shown based on HQs > 1, with ephemeral drainages depicted as yellow lines and non-ephemeral drainages as green lines.<sup>14</sup> Recognizing that aquatic habitat in STSIU surface-water drainages in bedrock sections of drainage channels consists mostly of isolated and distal pools (with the exception of Stock Tanks in these areas), Figure 4-3 also distinguishes the bedrock-dominated channels as bright yellow lines. The alluvium-dominated drainages are depicted with pale yellow lines. This depiction is intended to highlight drainage areas in which significant pool habitats do not occur due to substrate limitations in the alluvial habitat where habitat for biota is minimal. Pools in the alluvial, ephemeral habitat are rare. In summary, the solid green lines shown in Figure 4-3 represent non-ephemeral drainages in which copper contamination is not expected to be of concern based on compliance with the NMAC §20.6.4.809 based on the copper WER-based HQs ( $HQ \leq 1$ ) in Table 4-6. Similarly, the solid yellow lines represent drainages in which copper contamination is not expected to be of concern based on compliance with the NMAC §20.6.4.808 based on the copper WER-based HQs ( $HQ \leq 1$ ) in Table 4-5.

After the drainages of potential concern ( $HQ > 1$ ) were initially identified (A, C1, C2, D2, D3, Martin and Rustler Canyon), drainages that should be carried forward in the FS for consideration for remediation were identified based on the most current conditions. If a drainage had  $HQ > 1$  in 2006-2007 or earlier, and the exceedance was no longer present in the same general locations in future years (2010-2013), then the drainage was not considered for remediation as it already has achieved the pre-FS RAC. This was true for C1 ephemeral (downgradient of Bolton Canyon confluence), Martin Canyon, and Rustler Canyon. Therefore, Drainages A, non-ephemeral C1 upgradient of Bolton Canyon confluence, C2 (ephemeral only), D2, and D3 were carried forward for consideration for remediation.

## 4.2.2 Copper – Stock Tanks

Water chemistry has been evaluated for some of the STSIU stock tanks during previous Site investigations, including the STSIU ERA, STSIU RI, and Wet Season Survey, and from select stock tanks in 2021. The STSIU stock tanks consist of concrete or steel tanks that receive water from springs or windmill-pumped groundwater, and earthen tanks that receive the majority of their water from surface runoff. Earthen tanks in STSIU are typically located in surface water drainages and thus collect water from upgradient drainage channels during precipitation events. A few earthen tanks, however, are located outside of drainage channels and receive overland flow directly from adjacent upland areas instead of upgradient drainage channels. Figure 4-4 shows the current and historical locations and type of known STSIU stock tanks and distinguishes individual tanks of potential concern based on the WERs, as described below.

All sampled stock tanks are on ephemeral drainages, and while they generally recharge during the monsoon season, they dry up by the end of the year. Therefore, STSIU-specific NMAC §20.6.4.809 is used herein to evaluate the nature and extent of copper contamination in stock tanks. As indicated in Figure 4-4, earthen stock tanks are the major focus of this evaluation because these tanks receive their water from surface runoff, which can potentially be contaminated by upland COC sources (in contrast to steel and concrete tanks that receive groundwater via springs or windmill pumps). Surface water chemistry data are available for 11 earthen stock

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<sup>14</sup> Lucky Bill Canyon was designated as ephemeral in southern drainage and non-ephemeral in northern drainage based on observations.

tanks, which includes the majority of earthen stock tanks located in the general area of potential concern based on proximity to the historic smelter location (i.e., within subwatersheds A through E) (Figure 4-4; Table 4-7). WERs determined in drainage pools for each subwatershed were used for the stock tanks in the corresponding watershed. These WERs are expected to represent conservative estimates of potential WERs for stock tanks because organic carbon, which greatly mitigates copper toxicity, is generally greater in stock tanks than in surface water pools (due to additional organic matter input from livestock use and aquatic vegetation).

Application of the standard NMAC §20.6.4.900 for the stock tanks results in exceedances of the standard hardness-adjusted copper criteria with the greatest hardness-adjusted acute HQ equal to 29 in the subwatershed E (west drainage, Figure 4-3) stock tank 6; Table 4-7).<sup>15</sup> Adjusted copper HQs were calculated for each stock tank following the STSIU-specific NMAC §20.6.4.809 using two WER methods: the watershed-specific WERs and the geometric mean STSIU WER (Table 4-4). For stock tanks that were sampled multiple times throughout previous investigations, geometric mean adjusted copper HQs were also calculated to provide a more robust estimate of tanks of potential concern. Using the above WER methodology and evaluating the geometric means when available, stock tanks 6, 15, 29, 60, and West Drainage were determined to have concentrations of potential concern to aquatic life with HQ > 1 (Table 4-7). However, stock tanks 15, 26, 29, and 60 were removed between 2013 and 2022 as part of rangeland improvements, as discussed in Section 4.1.2, and are not considered further, leaving only stock tank 6 in the west drainage of concern.

## **4.3 Surface Water Locations to be Evaluated for Remedial Alternatives**

### ***Drainages***

Surface water locations described above as areas of potential concern to aquatic life uses were evaluated for remedial alternatives. Drainages A, C1, C2, D2 and D3 are carried forward..

### ***Stock Tanks***

Stock tank 6 was determined to have concentrations of potential concern to aquatic life uses and will be evaluated for remedial technologies.

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<sup>15</sup> This stock tank 06 was sampled in 2004 and 2006 for the STSIU RI to represent E2 drainage in subwatershed E, with HQs of 1 and 6, respectively (Table 4-7). Usually no water is found in the rest of this ephemeral drainage and because sampling was always in the stock tank location where water was found, E drainage is not included in Table 4-5.

## 5. Identification and Screening of Potentially Applicable Technologies

### 5.1 Soil

As part of the FS Proposal, a range of potential soil remedial technologies have been identified and summarized for the upland areas of the STSIU. In 2006, an extensive literature search was conducted consisting of a review of over 500 abstracts related to potential soil remedial technologies for treatment and/or removal of the primary COC for the site (copper). As a result of this review, 12 technologies were identified and included in this FS as Table 5-1 for a preliminary screening and evaluation. The preliminary screening and evaluation of the potential soil remedial technologies has been performed to determine which remedial technologies should be retained for consideration as part of the FS alternatives analysis, which will include a comprehensive alternatives evaluation for remedial alternatives for the site. The preliminary screening of each remedial technology is based on USEPA Guidance for Conducting Remedial Investigations and Feasibility Studies under CERCLA (USEPA 1988) and includes an evaluation of the effectiveness, implementability, and cost. Potential use of institutional controls consistent with CERCLA guidance (USEPA 2012) may be warranted for implementation of specific remedial technologies as described below.

Based on the preliminary screening conducted in the FS Proposal, only a brief summary of the technology is presented, with a summary of the effectiveness, implementability and cost presented in Table 5-1. If the remedial technology is considered viable, it will be retained for consideration as part of the site-wide remedial alternatives analysis in Section 6.

A brief explanation of these soil remedial technologies described below and a preliminary screening of each technology is presented in Table 5-1.

#### 5.1.1 No Action

This remedial technology consists of leaving the site soils in their current condition without performing any soils/vegetation removal or treatment, engineering controls, or institutional controls as part of the remediation efforts. This technology is provided as a baseline for screening other technologies and is summarized as Technology No. 1 in Table 5-1.

##### **Screening Result**

No Action is being retained as a baseline for comparison with other remedial technologies in the FS and for potential use in conjunction with other technologies where implementability may be low.

#### 5.1.2 Monitoring

This remedial technology consists of leaving the site soils in their current condition without performing any soils/vegetation removal or treatment, engineering controls, or institutional controls as part of the remediation efforts. As part of this technology, a monitoring program would be implemented to observe and document the occurrence of natural attenuation of site contaminants. Natural attenuation may occur due to (1) cessation of processes that contributed to copper deposition in soil and (2) continuation of natural processes that deposit clean

dust/soil, rain, and litter on the soils. The cessation of operation of the smelter in 2007 reduced copper deposition in the soils, and reclamation of many of the tailing ponds has reduced windblown deposition in the non-operational areas. Natural processes that deposit clean dust/soil or reduce its toxicity include wind deposition, periodic regular alkaline rains that reduce copper bioavailability over time (increase pH and pCu), unusual white rain events (windstorms picking up lime from playas and depositing them on the site which increases pH), and soil erosion from uplands into drainages. Rainfall also can cause leaching of copper into deeper soils, reducing exposure at the surface to ecological and human receptors. Monitoring is designed to detect if these processes are occurring, reducing the copper concentrations and increasing pH (and thus pCu) over time. The rate of attenuation cannot currently be estimated due to limited number of years with data in the same location and variability in the data.

Monitoring would include collection of qualitative and quantitative samples of STSIU media such as surface soils, vegetation and other biotic media. This technology is provided as a baseline for screening other technologies and is summarized as Technology No. 2 in Table 5-1.

#### **Screening Result**

Monitoring is being retained as a baseline for comparison with other remedial technologies in the FS and for potential use in conjunction with other technologies where implementability may be low.

### **5.1.3 Excavation and Reuse**

This remedial technology consists of removal of soils with contaminants above the soil pre-FS RAC levels located 0- to 1-inches to 6-inches bgs, depending on the potential receptor. Depth of potential removal could vary depending on site features such as the presence of bedrock. Final lateral and vertical removal extents would be determined during the remedial design. This technology would remove soils considered to be impacted for final onsite management at the waste rock stockpiles or for future use as soil fill material for the Chino Mine tailings pond closure activities or other operational areas of the site. The soils would be ripped if needed (or filled with clean soil) to improve their texture for plant growth and areas seeded to meet CCP guidelines. This technology is summarized as Technology No. 3 in Table 5-1.

#### **Screening Result**

Soil excavation could be applied at the site for the overall site-wide remedy or applied at targeted locations with higher concentrations of contaminated soils in conjunction with another technology. Although this technology may not be implementable at all areas, this technology is being retained due to its high effectiveness at reducing site contaminants.

### **5.1.4 Excavation and Disposal**

This remedial technology consists of removing soils above the soil RAC levels in the same manner as described in the excavation and onsite management remedial technology described in Section 5.1.3. However, instead of final onsite management of soils at the waste rock stockpiles or re-using the removed soils for fill material in the operational areas of the site, the soils would be disposed of at an offsite commercial disposal facility.

Removed soils would be characterized in accordance with Resource Conservation Recovery Act regulations to determine final offsite transportation and disposition requirements. This technology is summarized as Technology No. 3a in Table 5-1.

### **Screening Result**

Soil excavation could be applied at the Site for the overall site-wide remedy or applied at targeted locations with higher concentrations of contaminated soils in conjunction with another technology. This technology would only be implemented on a limited or case-by-case basis, in select areas. Although this technology may not be implementable at all areas, this technology is being retained due to its high effectiveness at reducing site contaminants.

## **5.1.5 Soil Amendments – Limestone and/or Organic Matter**

Many soil amendment technologies exist for reducing metals bioavailability, toxicity, and mobility in soils. They rely on changing soil chemistry to affect the solubility or mobility of site contaminants within the soil column, and/or improve vegetative cover or speciation. Several soil amendments are described further below including, pH adjustment via lime addition and/or organic matter, tilling (Section 5.1.6), ferrihydrite (Section 5.1.7) and chelating agents (Section 5.1.8). The pH adjustment and/or organic matter addition technology is summarized as Technology No. 4a in Table 5-1. As included in the Amendment Study Work Plan (Arcadis 2008a), data collected over the five years of monitoring was assessed to determine if a particular amendment, or combination thereof, can be successfully applied to particular areas at the Site.

The analysis in Arcadis (2017) in Appendix A indicates that liming is recommended for soils with pH < 2 but liming is not that helpful above that pH, and soils pH identified herein are not less than pH 2. Organic matter was not recommended after the amendment study was completed because cow manure brought in weedy plants that degraded the habitat (however, other forms of organic matter could be considered). Use of organic matter was screened out in this FS because other amendments were more effective (Appendix A).

### **Screening Result**

Soil amendment using lime will be retained as a remedial technology to be considered as part of the comprehensive remedial alternative for further evaluation in the FS.

## **5.1.6 Soil Amendments – Tilling**

Soil mixing by using mechanical tilling or ripping technology is being evaluated as part of this FS for use at the site as part of the comprehensive remedial alternative. This technology does not include the addition of other amendments such as lime and/or organic matter or include seeding; however, some of the subsurface soils at STSIU may include Gila Conglomerate Formation or other geological formations which have high alkalinity and geochemistry that may serve as a naturally-occurring amendment from wind-borne and erosional deposition. Tilling would be conducted in a similar manner as performed in the soil amendment study. Initially, the ground surface vegetation is cleared and grubbed using a bulldozer and/or excavator. Following vegetation clearing, the tilling is conducted using a 140 blade (or similar) attached to a bulldozer to mix to a pre-determined depth of soil. In areas requiring soil mixing with limited access to larger equipment, hand tilling equipment can be used as an alternative to the bulldozer to mix soils. Tilling is less intrusive in general than excavation; it lowers disruption to habitat and lowers carbon footprint compared to alternatives relying on excavation.

Tilling, which mixes higher concentration surface soils with deeper soils, has the potential to provide additional reduction of metal concentrations and to raise pH conditions to more neutral pH conditions pending existing pH levels within the soil treatment area being tilled. Plant coverage, pH, and soil chemistry would be monitored post-tilling operations. As part of the remedial design phase, additional soil sampling (contaminant levels and soil

chemistry) within the soil treatment column would be conducted to determine if tilling alone would be appropriate technology and what is the appropriate soil mixing depth within each soil treatment area to raise pH conditions. The soil amendments - tilling technology is summarized as Technology No. 4b in Table 5-1.

### **Screening Result**

Tilling is being retained for comparison with other remedial technologies in the FS and for potential use in conjunction with other technologies where implementability may be low. Tilling could potentially be applied at the majority of the site areas containing sufficient equipment access and appropriate terrain slopes or at targeted locations (e.g. in areas where lime and organic matter soil amendments are not applied, if applicable).

## **5.1.7 Soil Amendments – Ferrihydrite**

The use of the soil amendment ferrihydrite as a potential soil remedial technology is being evaluated for use at STSIU as part of a comprehensive remedial alternative. The addition of ferrihydrite to soils containing copper has been observed to bind copper, reduce free  $\text{Cu}^{2+}$  activity, and total soluble and labile concentrations of copper. The soil amendments - ferrihydrite technology is summarized as Technology No. 4c in Table 5-1.

### **Screening Result**

A site-specific data gap for this particular technology currently exists. And if this technology were to be considered a pilot and/or bench-scale treatability study would be proposed. While pilot and/or bench scale treatability studies are not being proposed to be conducted at this time, this technology is being retained as a remedial technology to be considered as part of the comprehensive remedial alternative for further evaluation in the FS.

## **5.1.8 Soil Amendments – Chelating Agents**

The application of chelating agents as a potential soil remedial technology is being evaluated for use at STSIU as part of a comprehensive remedial alternative. Specifically, chelating agents are being evaluated for use in the following soil remedial technologies:

- Phytoextraction;
- Soil Washing (Ex-Situ); and
- Soil Washing (In-Situ).

Chelating agents are compounds that are added to the soil to either assist in increasing the uptake of the contaminant (i.e., copper) into plants for the phytoextraction process or for removing a metal from soils as part of a soil washing technique. The use of chelating agents in the phytoextraction and soil washing processes are discussed in detail below and are summarized as Technologies Nos. 4d1, 4d2, and 4d3 in Table 5-1.

### **5.1.8.1 Soil Washing (Ex-Situ)**

Ex-situ soil washing is a soil remedial technique consisting of removing and concentrating contaminants from bulk soil using separation methodologies. Soil washing can be applied to soils containing heavy metals. The resulting concentrated soil containing the contaminants must be characterized for further treatment and/or offsite disposition. The “clean” portion of the separated soil is also characterized to determine if it meets the criteria for on-site reuse to be returned to the excavations or if it requires further treatment and/or offsite disposition.

The design of the soil washing process, including the size of scrubber unit, type of soil washing detergent, and soil handling requirements, would be determined via a pilot treatability study and during the remedial design.

**Screening Result**

Although the cost of soil washing may be moderately lower than excavation, the uncertainty of this technology and the access limitations of water and equipment to certain areas of the STSIU does not make it a viable remedial technology. Soil washing is not being retained as a remedial technology for consideration as part of the comprehensive remedial alternative for further evaluation in the FS.

**5.1.8.2 Soil Washing (In-Situ)**

In-situ soil washing consists of introducing a chelating agent into the soil. The chelating agent assists in mobilizing the contaminant within the soil column and allows it to become more soluble in the groundwater. The groundwater, containing the site contaminant, is then extracted with a groundwater extraction system for treatment and/or disposal.

**Screening Result**

Due to the lack of infrastructure required for the groundwater extraction system, the high costs, and the uncertainty in the effectiveness, in-situ soil washing is not considered a viable remedial technology. In addition, site accessibility issues, including the remoteness of the site and incongruous nature of areas needing treatment, will make soil washing less implementable or potentially infeasible for certain portions of STSIU. Therefore, in-situ soil washing is not being retained as a remedial technology for consideration as part of the comprehensive remedial alternative for further evaluation in the FS.

**5.1.8.3 Phytoextraction**

Phytoextraction is the process of plants taking up contaminants (i.e., copper) located in the soils via the plant root system. Once the metals have been transferred through the root system, the contaminants are subsequently transferred and accumulated into the aboveground portions of the plant tissue. Once the phytoaccumulation (contaminants transferred to the above ground plant tissue) has occurred, the contaminants are removed from the site by harvesting the plants. An additional phytoextraction technology consists of the contaminant accumulation occurring just in the root system, resulting in the need to harvest the entire plant (including the roots) to remove the contaminants from the site. The plant species selected for phytoextraction must be able to tolerate the site contamination (i.e., copper), must be fast growing, have a high biomass, and easily be harvested.

**Screening Result**

Although phytoextraction costs significantly less during the initial phase of treatment and is less invasive as compared to other soil remedial technologies, the uncertainty of this technology being effective in the natural environment, the potential of SGFBs consuming impacted vegetation and seeds, and the significantly increased treatment duration, does not make it a viable option. Phytoextraction is not being retained as a remedial technology for consideration as part of the comprehensive remedial alternative for further evaluation in the FS.

**5.1.9 Containment – Soil Cover**

The use of a soil cover is being considered as a potential remedial technology at the site to contain the impacted soils and prevent exposure of site contaminants to potential site receptors. As part of this remedial technology, a soil cover would be placed over existing soils with site contaminant levels above the Pre-FS RAC values. In



addition, areas of the site that currently consist of soils intermixed with exposed bedrock but have been determined to historically contain fully established soil and vegetative covers would be considered for a soil cover.

The purpose of the soil cover is to place imported, approved soils upon impacted soils or exposed bedrock areas to provide a layer of clean soil that would provide protection of SGFBs and would promote growth of local, naturally occurring vegetation. Clean, approved soil cover material would be spread, graded, and compacted to promote positive drainage. The minimum thickness and type of soil, including percentage of organic matter, to be used for the soil cover would be selected during the remedial design. The containment – soil cover technology is summarized as Technology No. 5a in Table 5-1.

#### **Screening Result**

Installation of a soil cover is considered to be a viable, cost-effective, and easily implementable remedial technology and will be retained as a remedial technology to be considered a part of the comprehensive remedial alternative for further evaluation in the FS.

### **5.1.10 Containment – Impermeable Cover**

This remedial technology consists of placing an impermeable cover over targeted areas with soil contaminant levels above the pre-FS RAC. This technology would include preparing the ground surface and site grades to accommodate the installation of the impermeable cover layer. Ground surface preparation activities would include clearing and grubbing existing vegetation and smoothing and compaction of the ground surface using general construction equipment such as excavators, graders, and rollers. Once proper ground surface conditions have been met (i.e., site grade and subgrade compaction), the impermeable cover would be installed. Details of the impermeable cover, including subgrade preparation requirements, subbase layers, and final impermeable cover material and thickness would be determined during the remedial design. The containment – impermeable cover technology is summarized as Technology No. 5b in Table 5-1.

#### **Screening Result**

Because it is not practical to implement this technology on a large scale and due to the impacts of surface water infiltration and stormwater run-off, and limitations on future vegetated growth and grazing, an impermeable cover is not considered a viable remedial technology to be retained for further evaluation in the FS.

### **5.1.11 Surface Soil Controls – Phytostabilization**

The purpose of surface soil controls is to further stabilize the surface soils to prevent or greatly reduce airborne dispersion of soils and provide overall erosion control for storm water. Phytostabilization is considered a remedial technology that would provide increased surface soil controls. Phytostabilization consists of vegetating the ground surface with plant species that are targeted at increasing long term soil stabilization as compared to existing vegetative conditions and/or other revegetation options that are not focused specifically on soil stabilization. The seed and plant species, planting locations and density, thinning requirements, and operation and maintenance (O&M) activities would be determined during the remedial design. The surface soil controls - phytostabilization technology is summarized as Technology No. 6 in Table 5-1.

#### **Screening Result**

Phytostabilization may be challenging because plant species must be tolerant to low pCu soils but shrubs with deep roots could grow. This will be retained as a remedial technology to be considered as part of the comprehensive remedial alternative for evaluation in in the FS.

### 5.1.12 Phytoremediation

Phytoremediation consists of planting vegetation (trees and/or plants) that can uptake the contaminants located in the soil and subsequently remediate the soils. Trees and/or plants remove the site contaminants when the roots take in water and nutrients from the surrounding impacted soils. Metals are stored in the roots, stems, or leaves of the vegetation, effectively removing them from the soil. Activities that are associated with the implementation of phytoremediation include selection of the proper tree and plant species, site preparation (potentially clearing and grubbing existing vegetation), planting, and O&M to ensure that the trees and plants are being established. The phytoremediation technology is summarized as Technology No. 7 in Table 5-1.

#### **Screening Result**

Due to the extended time period it would initially take to reduce site contaminants from the soils and the fact that the site remedy effectiveness is directly dependent on the success of the trees and plants that can grow in low pCu soils, phytoremediation is not being retained for further evaluation in the FS.

### 5.1.13 Electrokinetic Remediation

Electrokinetic remediation is being evaluated as part of this FS for use at the site as part of the comprehensive remedial alternative. Electrokinetics is based on the principle that when direct current (DC) is passed through contaminated soil, certain (negatively charged) types of contaminants will migrate through the soil pore water to a place where they can be removed. This alternative uses electrode assemblies that are installed in the ground in a square array and connected to a DC voltage power supply. When the DC power supply is activated, a current passes through the soil. As the electric current is applied to the soil between the electrodes, water flows by electroosmosis in the soil pores usually toward the cathode. When the contaminated water reaches the cathode, it is pumped to the surface by circulating water within the ceramic casing.

Electrokinetic remediation has the potential to provide remediation of copper contamination without physically removing any soil. The lack of physical soil removal allows for remediation without impacting the surface vegetation at the cost of increasing acidity in the surface soil. This increase in acidity would compound the already low overall pH in Chino soils east of the historic smelter. The electrokinetic technology is summarized as Technology No. 8 in Table 5-1.

#### **Screening Result**

It is assumed that it is not practical to implement this remedial technology over large-scale area, and therefore this technology will not be retained for further evaluation in the FS Report. Even on a small scale, the potential decreased soil pH and difficulties implementing this alternative in areas with low soil moisture are considered significant enough that this alternative will not be retained for evaluation in the FS.

### 5.1.14 Summary and Identification of Data Needs

The following soil remedial technologies were evaluated in the preliminary screen:

- No Action [retained];
- Monitoring [retained];
- Excavation and Reuse [retained];

- Excavation and Disposal [retained];
- Soil Amendments – Lime and/or Organic Matter [retained];
- Soil Amendments – Tilling [retained];
- Soil Amendments – Chelating Agents [not retained];
- Soil Amendments – Ferrihydrite [retained];
- Containment – Soil Cover [retained];
- Containment – Impermeable Cover [not retained];
- Surface Soil Controls – Phytostabilization [retained];
- Phytoremediation [not retained]; and
- Electrokinetic Remediation [not retained].

There are no additional data needs that need to be considered based on this preliminary screen of remedial alternatives for soil.

## **5.2 Surface Water**

As part of the FS Proposal, a range of potential surface water remedial technologies have been identified and summarized for the drainage areas and stock tank associated with the Site. A preliminary screening and evaluation of the potential surface water remedial technologies has been performed to determine which remedial technologies should be retained for consideration as part of the FS, which will include a comprehensive alternatives evaluation for the site. The preliminary screening of each remedial technology is based on USEPA (1988) and includes an evaluation of the effectiveness, implementability, and cost.

Based on the preliminary screening, if the remedial technology is considered viable, it will be retained for consideration as part of the site-wide remedial alternatives analysis. As a result of this review, seven technologies were identified and are described in detail below and a preliminary screening of each technology is presented in Table 5-2.

### **5.2.1 No Action**

This remedial technology consists of leaving the drainage areas which are known to contain surface water with levels of site contaminants above surface water Pre-FS RAC values, in their current condition without performing any soil, sediment, vegetation, groundwater and/or surface water removal or treatment. This technology is being retained to serve as a baseline control to compare to other potential surface water remedial technologies. This technology is summarized as Technology No. 1 in Table 5-2.

#### **Screening Result**

No Action is being retained as a baseline control for comparison with other remedial technologies in the FS and for potential use in conjunction with other technologies where implementability may be low.

## 5.2.2 Monitoring

This remedial technology consists of leaving the drainage areas which are known to contain surface water with levels of site contaminants above surface water Pre-FS RAC values, in their current condition without performing any soil, sediment, vegetation, groundwater and/or surface water removal or treatment. As part of this technology, a monitoring program would be implemented to observe and document the occurrence of natural attenuation of site contaminants, as described in Section 5.1.2, above. Monitoring would include collection of qualitative and quantitative samples of site media such as surface water, in-drainage sediments, and/or vegetation. This technology is being retained to serve as a baseline control to compare to other potential surface water remedial technologies. This technology is summarized as Technology No. 2 in Table 5-2.

### **Screening Result**

Monitoring is being retained as a baseline control for comparison with other remedial technologies in the FS and for potential use in conjunction with other technologies where implementability may be low.

## 5.2.3 Excavation

This remedial technology consists of the removal of soils and/or sediments from the specified drainage areas. Based on the work presented in Appendix E, historical sediments and soils within the drainages may be contributing to surface water quality issues. Although a specific loading rate has not been determined for historical sediments for each drainage, sediments have been identified as a potential source to surface water. Therefore, excavation will be considered a viable remedial technology to consider for the comprehensive remedial alternative. This technology is summarized as Technology No. 3 in Table 5-2.

### **Screening Results**

Excavation of sediments is an effective and technically implementable way of removing contaminated sediments from surface water although the costs associated with the excavation of sediments are considered to be high. Therefore, excavation of sediments is being retained for further evaluation as part of a comprehensive remedial alternative in the FS.

## 5.2.4 In-Stream Removal of Suspended Sediments

This remedial technology consists of in-stream removal of suspended sediments via construction of settling basins within the stream drainage area pathway. The contaminants are adhered to the suspended sediments located within the surface water, subsequently contributing to the exceedances of the surface water Pre-FS RAC values. Removal of the suspended sediments containing the contaminants will result in lowering the total contaminant concentrations in the surface water. There may still be a potential for dissolution of contaminants from sediments into the dissolved phase.

Multiple settling basins would be constructed at specified locations along the drainage area to capture sediments at different points along the surface water drainage pathway. The location, size, and materials of the settling basins would be determined during the remedial design. The settling basins would be located in areas that are easily accessible by construction equipment for removal of the accumulated sediments. The frequency of sediment removal from the settling pools will depend on the rate of sediment accumulation and would be determined during the remedial design. This technology is summarized as Technology No. 4 in Table 5-2.

### **Screening Results**

In-stream removal of sediments seems to be an effective, technically implementable, and cost-effective way of removing contaminated sediments from surface water. Therefore, in-stream removal of sediments is being retained for further evaluation as part of a comprehensive remedial alternative in the FS.

## **5.2.5 Limestone Treatment**

This ex-situ remedial technology consists of the installation of limestone features within the surface water drainage area to passively treat surface water with contaminant levels above the Pre-FS RAC levels. Contaminants which either adhere to the suspended sediments in surface water or are within the dissolved phase contribute to lowering the pH of the water which creates a more acidic environment when compared to naturally occurring surface water in the region that do not contain the contaminants.

Limestone features would require installation at multiple locations along the surface water drainage areas. The multiple locations of the limestone features would provide increased treatment of the surface water as it progresses down the drainage area. The limestone features installation may consist of the construction of a waterfall using limestone masses to increase surface water contact of the water with the limestone. In addition, limestone may be installed as armoring and/or chips. The final design and location of the limestone features would be determined during the remedial design. This technology is summarized as Technology No. 5 in Table 5-2.

### **Screening Results**

Limestone treatment is an effective and technically implementable way of removing contaminated sediments from the surface water. Therefore, installation of limestone features is being retained for further evaluation as part of a comprehensive remedial alternative in the FS.

## **5.2.6 Alkaline Washing**

This in-situ remedial technology consists of the insertion of an alkaline fluid into the active channel and bar sediments in the drainages of the STSIU to treat surface water with contaminant levels above the Pre-FS RAC levels.

In-situ treatment would need to be evaluated using a pilot study to determine the effectiveness of this technology on the STSIU sediments. This technology would only be an effective remedial technology if the majority of metals loading for the STSIU drainages are related to constituents in legacy sediments. If in-situ treatment is determined to be an effective remedial technology, however, this technology would require extensive infrastructure installation along the drainages for transport and treatment of an alkaline fluid. The final design and location of this infrastructure would be determined during the remedial design. This technology is summarized as Technology No. 6 in Table 5-2.

### **Screening Results**

In-situ treatment may be effective and technically implementable way of neutralizing contaminated sediments prior to leaching to surface water, but due to the low implementability and high cost, this technology is not being retained for further evaluation as part of a comprehensive remedial alternative in the FS.

## 5.2.7 Sediment and Erosion Control

Based on the surface water quality data, presented in Section 4, stormwater run-off originating from the upland areas of the site is likely a contributing factor to surface water impacts within the drainage areas. A comprehensive sediment and erosion control system could be constructed to minimize impacts from entering the drainage areas. The sediment and erosion control system would be considered a permanent remedial technology that would reduce contaminants from entering the surface water located within the drainage area systems.

The sediment and erosion control system would consist of adjusting surrounding grade elevations where necessary and installing various types of Best Management Practices (BMPs) to redirect upland run-off away from the surface water drainage areas and allow it to infiltrate directly into the upland soils. Grading adjustments may include, but may not be limited to, construction of swales, drainage ditches, and/or catch ponds to capture stormwater run-off for infiltration and/or redirect stormwater run-off from entering the drainage areas.

Temporary and permanent BMPs may be utilized as part of the sediment and erosion control system. Temporary BMPs may include silt fences, straw bales, and/or erosion control mats to be installed during construction and be maintained for a limited time after construction until vegetation has sufficiently been established. Permanent BMPs may include placement of gravel, stone, and/or riprap to be placed on newly graded areas, as warranted. The erosion and control system, including grading adjustments and BMP details would be determined during the remedial design. This technology is summarized as Technology No. 7 in Table 5-2.

### **Screening Results**

Construction of sediment and erosion control systems near the drainage areas are considered to be effective, technically implementable, and cost effective. Therefore, installation of sediment and erosion control systems is being retained for further evaluation as part of a comprehensive remedial alternative in the FS.

## 5.2.8 Summary and Identification of Data Needs

The following surface water remedial technologies were evaluated in the preliminary screen:

- No Action [retained];
- Monitoring [retained];
- Excavation [retained];
- In-stream Removal of Suspended Sediments [retained];
- Limestone Treatment [retained];
- Alkaline Washing [not retained]; and
- Sediment and Erosion Control [retained].

Besides ongoing sampling activities, there are no additional data needs that need to be considered based on this preliminary screen of remedial alternatives for surface water.

A summary of retained remedial technologies is included in Table 5-3.

## 6. Assembly and Development of Remediation Alternatives

In this Section, remediation technologies retained after screening in Section 5 are assembled into remediation alternatives to identify one or more options that will address the remedial action objectives. A list of the remediation technologies and alternatives are presented in Tables 6-1 through 6-4. A more detailed description of the technologies is presented in Section 5.

### 6.1 Soils Alternatives – Total Metals

Five alternatives have been developed to consider for remediation of the soils for total metals within the STSIU. The regulatory standards for human health and ecological risk are different. Human health standards are 5,000 mg/kg of copper and 100,000 mg/kg for iron. Avian risk is based on 95UCL area-weighted average concentration within an exposure unit with concentrations greater than 1,600 mg/kg. Concentrations greater than 1,100 mg/kg requiring monitoring. The technologies of the alternatives are listed in Table 6-1.

Alternative 1: No Action

- Alternative 2: Monitoring
- Alternative 3: Excavation and Monitoring
- Alternative 4: Soil Cover and Monitoring
- Alternative 5: Soil Amendments/Tilling and Monitoring

#### 6.1.1 Alternative 1: No Action

A no action alternative is included as a baseline comparison to the other remedial alternatives. This alternative would leave the site for upland soil for arsenic, lead and copper in its current state.

#### 6.1.2 Alternative 2: Monitoring

In this alternative upland soils would be monitored for potential natural attenuation of copper concentrations in areas that fail the human health or ecological Pre-FS RAC including 95UCL area-weighted average copper concentrations within exposure units with concentrations between 1,100 and 1,600 mg/kg of copper would be conducted for avian risk. NMED (2011b) indicates:

*Chino shall propose risk-based monitoring of abiotic and/or biotic media to help ensure that risk from copper exposure of small ground-feeding birds in the exposure unit do not exceed acceptable levels. A protectiveness assessment consistent with CERCLA guidance will be conducted on a population basis...This monitoring shall include, at a minimum, ground-dwelling invertebrates, an indicator of copper exposure to birds that feed primarily on invertebrates during critical life stage. This requirement is intended to offset uncertainties in estimated copper exposure between 1,100 and 1,600 mg/kg. Monitoring shall be conducted in intervals no greater than 2 years for the first 5 years to establish a trend within the first 5 year review period. Monitoring frequency beyond the first years will be determined based on the first 5 year review recommendations.*

Monitoring would be completed once NMED issues the ROD.

### 6.1.3 Alternative 3: Excavation and Monitoring

In this alternative, excavation is a technology that can reduce soil copper concentrations greater than 5,000 mg/kg for human health and 1,600 mg/kg for avian risk. As discussed under Alternative 2, monitoring of 95UCL area-weighted average concentration within an exposure unit with concentrations between 1,100 and 1,600 mg/kg of copper would be conducted for avian risk.

Excavated soil, greater than the RAOs, would be taken to the operational area for subsequent reuse.

### 6.1.4 Alternative 4: Soil Cover and Monitoring

This alternative is a combination of soil cover long-term maintenance and monitoring. It would be implemented for the same area as Alternative 3, by covering all areas needing protection for human health (> 5,000 mg/kg copper) and avian risk (> 1,600 mg/kg copper). This alternative would include the placement of a clean cover (minimum 6-inches of soil) over the exposure areas and revegetation. Long-term maintenance and monitoring of the cover would be required, and institutional controls would be required.

### 6.1.5 Alternative 5: Soil Amendments/Tilling and Monitoring

This alternative is a combination of soil amendments, with potentially tilling, implemented for the same areas as Alternative 3 and 4. This alternative would implement soil amendments based on the results of the soil amendment study, which was completed in 2013. The study does not recommend manure amendments, but recommends lime amendments in areas with very low pH. Therefore, this alternative includes a combination of tilling and soil amendments.

## 6.2 Soils Alternatives – Cupric Ion Activity

Six alternatives have been developed to consider for remediation of the soils for pCu within the STSIU. The areas being considered for remedial alternatives are the areas presented in Figure 3-13 that have pCu concentrations less than a predefined PEL, where copper concentrations are greater than 327 mg/kg, and also meet the evaluation criteria described in Section 3.2.4. There is some overlap between the areas discussed in Section 6.1 (total metals) and Section 6.2 (pCu). The selection of a remedy for each remedial component may have an impact on the other. This interaction will be described in Section 7. The technologies of the alternatives are listed in Table 6-2.

- Alternative 1: No Action
- Alternative 2: Monitoring
- Alternative 3: Excavation (Reuse or Disposal) and Monitoring
- Alternative 4: Soil Amendments (Lime), Tilling and Monitoring
- Alternative 5: Tilling and Monitoring
- Alternative 6: Soil Amendments (Ferrihydrite), Tilling and Monitoring



- Alternative 7: Soil Cover and Monitoring
- Alternative 8: Surface Soil Controls – Phytostabilization

### **6.2.1 Alternative 1: No Action**

A no action alternative is included as a baseline comparison to the other remedial alternatives. This alternative would leave the site for upland soil for pCu in its current state.

### **6.2.2 Alternative 2: Monitoring**

In this alternative, upland soils for pCu would be monitored for potential natural attenuation of copper concentrations in areas that have a pCu less than 5.0, where copper concentrations are greater than 327 mg/kg.

### **6.2.3 Alternative 3: Excavation (Reuse or Disposal) and Monitoring**

In this alternative, to the extent possible, low pCu areas would be excavated until the pCu of the top six inches of soil meet the criteria of greater than 5.0 or copper drops below 327 mg/kg. Excavated soil, less than the pCu RAO, would either be taken to the operational area for subsequent reuse or taken to the West Stockpile for recycling (disposal option). Excavation of low pCu soils would be limited because of the infeasibility of excavating in areas of bedrock and steep slopes. Monitoring of the areas would be conducted to determine if the pCu values are stable over time.

### **6.2.4 Alternative 4: Soil Amendments (Lime), Tilling and Monitoring**

Based on the results of the soil amendment study plots (Appendix A) and white rain studies (Appendix B), lime could be added to areas where pCu is less than 5.0 or to low pH areas (pH < 2). This alternative is Alternative 5 with the addition of amendments, which can be in low pCu areas or in areas where soils cannot be tilled (e.g., too shallow or steep slopes). This alternative may be appropriate as it provides additionally for amendments in areas where tilling is not feasible. Some areas may still be inaccessible, even on foot, for application of amendments. Monitoring of the areas would be conducted to determine if the pCu values are stable over time.

### **6.2.5 Alternative 5: Tilling and Monitoring**

In this alternative, to the extent possible, low pCu areas would be tilled using mechanical means. No soil amendments or addition of organic material would be included in this alternative, based on the amendment study (Appendix A) finding that tilling alone can be very effective. Tilling would not be conducted on areas inaccessible to the mechanical equipment, or on bedrock outcrops. Monitoring of the areas would be conducted to determine if the pCu values are stable over time.

### **6.2.6 Alternative 6: Soil Amendments (Ferrihydrite), Tilling, and Monitoring**

In this alternative, ferrihydrite would be added to areas where pCu is less than 5.0. This alternative is Alternative 5 with the addition of amendments in areas where soils cannot be tilled (e.g., too shallow or steep slopes). This alternative may be appropriate as it provides additionally for amendments in areas where tilling is not feasible.

However, some areas may still be inaccessible, even on foot, for application of amendments. Monitoring of the areas would be conducted to determine if the pCu values are stable over time.

### **6.2.7 Alternative 7: Soil Cover and Monitoring**

In this alternative, a soil cover would be placed over existing soils with site contaminant levels above the Pre-FS RAC values. In addition, areas of the site that currently consist of soils intermixed with exposed bedrock but have been determined to historically contain fully established soil and vegetative cover will be considered for a soil cover. The purpose of the soil cover is to place imported, approved soils upon impacted soils or exposed bedrock areas to provide a layer of clean soil that would provide protection of SGFBs and would promote growth of local, naturally occurring vegetation. Clean, approved soil cover material would be spread, graded, and compacted to promote positive drainage. The minimum thickness and type of soil, including percentage of organic matter, to be used for the soil cover would be selected during the remedial design. Monitoring of the cover would be conducted to verify the integrity of the remedy and to determine if the pCu values are stable over time.

### **6.2.8 Alternative 8: Surface Soil Controls – Phytostabilization**

In this alternative, surface soil would be stabilized with vegetation to prevent or greatly reduce airborne dispersion of soils and provide overall erosion control for storm water. Phytostabilization consists of vegetating the ground surface with plant species that are targeted at increasing long term soil stabilization as compared to existing vegetative conditions and/or other revegetation options that are not focused specifically on soil stabilization. The seed and plant species (tolerant of low pCu), planting locations and density, thinning requirements, and OMM activities would be determined during the remedial design. Phytostabilization could also be effective when combined with remedial technologies that reduce pCu but is considered alone. Monitoring of the areas would be conducted to determine if the pCu values are stable over time and to verify vegetation establishment.

## **6.3 Surface Water Alternatives – Drainages**

Five alternatives have been developed to consider for remediation of the surface water drainages within the STSIU. The drainages being evaluated for remedial alternatives are the yellow drainages on Figure 8-1 that have water quality concentrations higher than water quality standards and have the potential for exposure to a receptor. Alternatives to address both drainage surface water run-off and drainage sediments are listed below. A more detailed description of the technologies is described in section 5. The technology types in the alternatives are listed in Table 6-3.

- Alternative 1: No Action
- Alternative 2: Monitoring
- Alternative 3: Excavation and Monitoring
- Alternative 4 Excavation, Sediment Control, Limestone Treatment, and Monitoring
- Alternative 5: Limestone Treatment and Monitoring
- Alternative 6: Sediment Control, Erosion Control, and Monitoring

### **6.3.1 Alternative 1: No Action**

A no action alternative is included as a baseline comparison to the other remedial alternatives. This alternative would leave the site the way it is for surface water drainages.

### **6.3.2 Alternative 2: Monitoring**

In this alternative surface water drainages would be monitored for natural attenuation for copper against surface water quality criteria parameters.

### **6.3.3 Alternative 3: Excavation and Monitoring**

This alternative would excavate deposited historical sediments from potential aquatic habitat in pools, and upstream of the pools to the extent that the drainages are accessible, in selected or targeted areas along each of the yellow drainages identified in Figure 8-1 (i.e., in hot spot removals or receptor locations). Excavation would be conducted only in the drainages and would not extend to the upland soils, as upland soils are addressed in the remedies discussed previously. Long-term monitoring of the surface water quality would be conducted in each of the drainages to document that the surface water quality is improving. This alternative would address historical sediments within the surface water drainages but would have limited impact on new sediments generated from upland soils.

### **6.3.4 Alternative 4: Instream Removal of Suspended Sediments and Monitoring**

In this alternative, sedimentation basins would be constructed within the yellow surface water drainages shown on Figure 8-1. The sedimentation basins would be constructed of limestone rock and slabs to increase the pH of surface water, thereby decreasing the bioavailability of copper in surface water. Construction of the sedimentation basins within existing areas of lower velocity would also provide an opportunity to excavate existing sediments within the surface water drainages in those areas and replace the material with limestone, thereby reducing total copper concentration of in-channel sediments. This alternative would address both the transportation of total copper, as well as dissolved copper concentrations. Long-term monitoring of the surface water quality would be conducted in each of the drainages to document that surface water quality is improving.

### **6.3.5 Alternative 5: Limestone Treatment and Monitoring**

In this alternative, limestone would be placed within the yellow surface water drainages shown on Figure 8-1. The limestone would increase the pH of surface water, thereby decreasing the bioavailability of copper in surface water. The limestone would be placed in areas within the drainage where surface water contact time with the limestone could be maximized (e.g., low gradient reaches, basins, at the bottom of waterfalls/knickpoints, etc.). This alternative would address both the transportation of total copper, as well as dissolved copper concentrations. Long-term monitoring of the surface water quality would be conducted in each of the drainages to document that surface water quality is improving.

### **6.3.6 Alternative 6: Sediment Control, Erosion Control, and Monitoring**

In this alternative BMPs and slope changes would be made to prevent the movement of suspended sediments from entering surface water drainages. This alternative would be most appropriate for areas where upland treatments or removals are not conducted. BMPs would be installed along select locations along the length of the yellow surface water drainages in Figure 8-1. Targeted stream length would be in areas with the highest potential for sediment loading. This alternative would address suspended sediment contribution to total copper concentrations and potential loading of sediments to the surface water drainages. This alternative would require long-term monitoring of surface water quality to determine if it is improving as well as monitoring and maintenance of the constructed/installed BMPs. Preference should be given to BMPs or controls that include a limestone component which may address the dissolved phase of metals in storm water.

## **6.4 Surface Water Alternatives – Stock Tanks**

Five alternatives have been developed to consider for remediation of the surface-water fed, earthen stock tank 6 within the STSIU. The remedial alternative for the stock tank is linked to the remedial alternatives selected for surface water drainages, discussed in Section 6.3. The alternatives are summarized in Table 6-4.

- Alternative 1: No Action
- Alternative 2: Monitoring
- Alternative 3: Excavation and Monitoring

### **6.4.1 Alternative 1: No Action**

A no action alternative is included as a baseline comparison to the other remedial alternatives. This alternative would leave the site for surface water drainages.

### **6.4.2 Alternative 2: Monitoring**

In this alternative monitoring of the surface water stock tank would be conducted for surface water quality parameters and potential natural attenuation.

### **6.4.3 Alternative 3: Excavation and Monitoring**

In this alternative the stock tank would be excavated to remove the historical sediment from the bottom of the tank. Excavated soil would be taken to the operation area for subsequent reuse. Monitoring of the stock tank would be required to document that the water quality in the stock tank meets regulatory standards.

## 7. Analysis of Alternatives

The remediation alternatives developed in Section 6 are evaluated in this Section. A final remediation alternative is recommended for each remedial component (soils – total metals, soils – pCu, surface water drainages, and a surface water stock tank).

The descriptions provided below include the major activities for each remedy at sufficient level of detail for the purposes of this FS. Detailed designs, sampling and analysis plans, inspection and monitoring plans, and other documents necessary for implementing the alternatives will be prepared at a later date after the remedy has been selected and documented in the ROD. Remedial alternative analysis is based on the following USEPA evaluation criteria:

- overall protection of human health and the environment;
- compliance with ARARs;
- long-term effectiveness and permanence;
- reduction of toxicity, mobility, or volume;
- short-term effectiveness;
- implementability; and
- cost;

In addition to the above standard USEPA evaluation criteria, the remedial alternatives will be evaluated using green remediation criteria. Factors for each remedial alternative that will be evaluated will also be evaluated as a green alternative, which may include, but may not be limited to, conservation of natural resources, carbon footprint, greenhouse gas emissions, and sustainability of the design.

The first two criteria are considered threshold criteria. Threshold criteria are minimum requirements that must be satisfied by an alternative. These criteria are applied to individual alternatives, but not used in the comparative evaluation of alternatives. The next five are the balancing criteria. Comparative evaluation is based on the balancing criteria used to assess tradeoffs between each alternative.

The USEPA includes two more criteria not in the above list, state and community acceptance, which are modifying criteria and are more difficult to assess at the initial FS stage. Typically, after the FS is out for public review, the FS is finalized with reference to these two criteria. Alternatively, after the FS is finalized, an alternative is selected as the proposed remedial action. The proposed remedial action is described along with the basis for its selection in the Proposed Plan. The evaluation of the modifying criteria is based on the state and public comments on the FS and the Proposed Plan. State and community concerns, and any resulting changes in the selected remedial actions, are documented in the ROD for the site. Therefore, the two modifying criteria are not evaluated in this document.

Each of the remedial alternatives have been summarized in Table 7-1 through 7-4 for: soils – total metals, soils – pCu, surface water drainages, and surface water stock tanks. A summary of the areas, drainage length and stock tanks involved is presented below. The selection of a remedial alternative for the upland soils will focus on the 140 acres identified for monitoring for copper concentrations in abiotic and biotic media (none were identified for active remediation for copper), and the 377 acres carried forward for pCu requiring active remediation. The areas requiring active remediation or monitoring for soils are presented in Figure 8-1. The selection of a remedial

alternative for surface water will focus on the drainages with watershed-specific HQs  $\geq 1$  for WER-adjusted copper concentrations in ephemeral drainages, which are A, C1, C2, D2, and D3 Drainages in addition to stock tank 06 (Figure 8-1).

## 7.1 Evaluation Criteria

### 7.1.1 Threshold Criteria

Under CERCLA, remediation alternatives must meet the following two threshold requirements:

- overall protection of human health and the environment;
- compliance with ARARs;

#### 7.1.1.1 Protection of Human Health and Environment

This criterion addresses the degree to which the remediation alternative is protective of human health and the environment, considering both long-term and short-term risks. Overall protectiveness is a threshold criterion, in that those alternatives that do not achieve adequate protection of human health or the environment are eliminated from further consideration. The ability of the alternatives to achieve remedial action objectives is part of the evaluation of this criterion. This criterion considers the evaluation of other criteria, especially long-term effectiveness and permanence; reduction of toxicity, mobility and volume; and short-term effectiveness. This is a summary of the overall evaluation of these other criteria. Because the overall protection of human health and the environment is a summary evaluation, it is evaluated for screening individual remediation alternatives, but not used in comparative evaluation of the alternatives.

#### 7.1.1.2 Compliance with ARARs

This criterion addresses whether or not the alternative meets ARARs, which were defined in Section 2. As with overall protectiveness, compliance with ARARs is a threshold criterion that must be met for an alternative to be selected.

### 7.1.2 Balancing Criteria

#### 7.1.2.1 Long-term Effectiveness and Permanence

This criterion addresses the results of remedial actions in terms of the risk remaining at the site after the response action objectives have been met and the reliability of the remedial action at reducing risks over an extended period of time. The primary focus of this evaluation is the extent and effectiveness of the control that may be required to manage the risks posed by the contaminants in the long-term.

#### 7.1.2.2 Reduction of Toxicity, Mobility, or Volume through Treatment

This criterion addresses the degree to which a remediation alternative, through treatment (not through removal or covering) reduces the toxicity of contaminants, the ability of the contaminants to migrate into the accessible environment, or the volume/quantity of the contaminated material. This criterion focuses the analysis of the

preference for treatment under CERCLA. Effectiveness and reliability of treatment are addressed under long-term effectiveness and permanence, and are not addressed under this criterion.

### 7.1.2.3 Short-term Effectiveness

This criterion addresses short-term effects on human health and the environment while the alternative is being implemented. The following factors should be addressed as appropriate for each alternative: protection of community and workers during construction, environmental impacts, amount of time to implement the remedial actions.

### 7.1.2.4 Implementability

This criterion addresses the degree of difficulty in implementing each alternative. Implementability can be divided into three categories: technical feasibility, administrative feasibility, and availability of services and materials. Implementability is a key criterion for more complex alternatives and reliance on innovative technology increases. Implementability issues are important because they address the potential for schedule delays, costs increases, and remedy failure to achieve the intended results. The evaluation considers the following:

- *Technical Feasibility.* Addresses site-specific factors that could prevent successful implementation of an alternative. As previously mentioned in Section 6 implementability issues could include physical interferences, such as bedrock, steep slopes or limited access.
- *Administrative Feasibility.* The degree of difficulty anticipated due to regulatory constraints such as permit approvals and degree of coordination between regulatory agencies and stakeholders.
- *Availability of Services and Materials.* The availability of labor, equipment and materials to implement the alternatives.

### 7.1.2.5 Cost

The criterion is used to consider the costs of implementing each alternative including capital costs and operating, monitoring, and maintenance costs. Costs that are excessive compared to the overall effectiveness may be considered as one of several factors used to eliminate an alternative. Alternatives providing effectiveness and implementability similar to that of another alternative, but at a higher cost, may be eliminated. Appendix F presents the supporting cost information. Components of Appendix F include:

- Capital cost estimates, with supporting narrative regarding the source(s) of component unit costs (e.g., RS Means, etc.);
- Monitoring cost estimates, with supporting narrative regarding the source(s) of the component unit costs;
- Periodic costs estimates, with supporting narrative regarding the source(s) of the component unit costs; and
- Net present value cost estimates for each alternative that combine the capital, annual monitoring, and periodic costs, including the selected discount factor and justification for its use, frequency for periodic costs, and overall time duration for the net present value calculations.

All component cost estimates are linked to specific references.

### 7.1.3 State and Community Acceptance

These last two evaluation criteria are not evaluated in the STSIU FS. These criteria will be addressed in the ROD.

### 7.1.4 Green Remediation

Factors for each remedial alternative that will be evaluated will also be evaluated as a green alternative, which may include, but may not be limited to, conservation of natural resources (fuel), carbon footprint, greenhouse gas emissions, and sustainability of the design.

## 7.2 Evaluation of Soils Alternatives – Total Metals

As presented in Section 3, all sample-specific concentrations were less than the 5,000 mg/kg copper and 100,000 mg/kg Fe, as specified by the Pre-FS RAC for human health, except in IRA areas where removal could not be accomplished due to bedrock, infrastructure or steep slopes. All habitat alliance polygons had a 95UCL area-weighted average copper concentration less than 1,600 mg/kg and meet the avian pre-FS RAC. Three polygons had 95UCL area-weighted average copper concentrations between 1,100 and 1,600 mg/kg (1-3, 10-16 and 88-15). It is therefore appropriate to evaluate non-active remedial alternatives for these acres.

The alternatives are summarized as:

- Alternative 1: No Action
- Alternative 2: Monitoring

Using the 95UCL area-weighted average concentrations, 140 acres were identified within the monitoring range of the avian pre-FS RAC (1,100 to 1,600 mg/kg), but no acres exceed the SGFB Pre-FS RAC of 1,600 mg/kg. Thus, Alternatives 1 and 2 are evaluated further below and alternatives with active treatment (Alternatives 3, 4, and 5) are not appropriate based on current conditions; Alternatives 1 and 2 are detailed further in Table 7-1.

### 7.2.1 Detailed Analyses of Soil Alternatives – Total Metals

This section presents a detailed analysis of the copper soil remediation alternatives for each of the threshold and balancing criteria described in Section 7.1, above.

#### 7.2.1.1 Alternative 1: No Action

##### Protection of Human Health and the Environment

The No Action alternative does not provide any additional protection of human health and the environment. However, this alternative meets the human health pre-FS RAC for copper and iron and the part of the avian pre-FS RAC that specifies copper concentrations should be less than 1600 mg/kg. It does not meet the monitoring requirement for copper concentrations between 1100 and 1600 mg/kg, and thus does not fully meet the RAOs.

##### Compliance with ARARs

There are no chemical-specific ARARs for surface soil except for the pre-FS RAC, which would not be met for monitoring of copper to protect wildlife. However, the portion of the pre-FS RAC that requires active remediation



when copper > 1600 mg/kg is met with the no action alternative because no exposure units exceeded 1600 mg/kg.

### **Long-term Effectiveness and Permanence**

The magnitude of residual risk (i.e., potential for unacceptable risk to wildlife over the long term) could change from current conditions. Therefore, this alternative would not be acceptable for long term effectiveness or permanence because it does not include monitoring.

### **Reduction of Toxicity, Mobility, or Volume**

The No Action alternative does not reduce the toxicity, mobility, or volume because no treatment would be implemented.

### **Short-term Effectiveness**

The No Action alternative would be effective in the short term because current conditions are protective as they meet the human health pre-FS RAC and the avian pre-FS RAC such that active remediation is not required. No actions would be implemented so there would be no risks to the community or workers or impacts to the environment during implementation.

### **Implementability**

No actions would be implemented so this alternative is implementable. The No Action alternative would not limit further actions at STSIU and no approval or coordination with other agencies would be necessary.

### **Cost**

There is no cost associated with the No Action alternative. The No Action alternative provides a baseline for comparison of other alternatives.

## **7.2.1.2 Alternative 2: Monitoring**

### **Protection of Human Health and the Environment**

The Monitoring alternative does not directly provide any additional form of overall protection of human health and the environment beyond current conditions. However, the ability of the Monitoring alternative to satisfy this threshold criteria is high and achievable because it meets the requirements of the pre-FS RAC. The acres involved do not require remediation because the concentrations currently do not pose risk to human health and the environment; however, they do require monitoring based on the pre-FS RAC. The monitoring in this alternative ensures current conditions between 1100 mg/kg and 1600 mg/kg copper do not change and adversely affect the small ground feeding bird. If conditions change adversely, monitoring may trigger a remedial alternative over time designed to improve protection of human health and the environment.

### **Compliance with ARARs**

There are no chemical-specific ARARs for surface soil except for the pre-FS RAC, which is met with this alternative by monitoring exposure units with copper in soils between 1,100 mg/kg and 1,600 mg/kg, and given no exposure units exceed the active remediation threshold of 1,600 mg/kg. The Monitoring alternative would meet the threshold criterion of compliance with ARARs.

### **Long-term Effectiveness and Permanence**

The magnitude of risk would be consistent with current conditions unless monitoring shows an adverse change in results that requires the implementation of additional remedial actions. More likely, natural attenuation will lower copper in soils due to erosion of surface soils with copper and deposition of cleaner windblown and colluvial soil over time. This alternative affords the opportunity to continue data collection and to take additional action if necessary.

### **Reduction of Toxicity, Mobility, or Volume**

The Monitoring alternative does not reduce the toxicity, mobility, or volume because no treatment would be implemented.

### **Short-term Effectiveness**

The Monitoring alternative would be effective in the short term because current conditions are protective as they meet the human health pre-FS RAC and the avian pre-FS RAC such that active remediation is not required. No remedial actions would be implemented so there would be no risks to the community or workers or impacts to the environment during implementation.

### **Implementability**

Monitoring would be readily implementable. This alternative would not limit further actions at STSIU and no approval or coordination with other agencies would be necessary beyond monitoring coordination.

### **Cost**

The present value cost of this alternative is \$1.04 million (see Appendix F).

## **7.2.2 Comparative Analyses of Soil Alternatives – Total Metals**

Table 7-1 provides a comparative analysis for the two alternatives for addressing total metals in soils. For the 140 acres for which monitoring is indicated based on 95UCL area weighted average concentrations between 1,100 and 1,600 mg/kg, Alternative 1 does not meet the threshold criteria because the acres do not meet the Pre-FS RAC for monitoring and, therefore, do not meet the state's criteria. Alternative 2 is monitoring and this remedial alternative meets the state's Pre-FS RAC. Alternative 2 is also considered the greenest remediation alternative behind Alternative 1. Based on the amendment, pH and insect studies, the white rain in 2008 reduced copper uptake into plants and thus reduced phytotoxicity, and improved plant richness. However, NMED (2011a) indicates that:

*If the 95UCL area-weighted average copper concentration in an exposure unit exceeds 1,100 mg/kg, Chino shall propose risk-based monitoring of abiotic and/or biotic media to help ensure that risk from copper exposure of small ground-feeding birds in the exposure unit do not exceed acceptable levels. A protectiveness assessment consistent with CERCLA guidance will be conducted on a population basis. This requirement is intended to offset the uncertainty in risk conclusions for this area due to the paucity of soil and invertebrate data available to characterize risk since the cessation of smelter operations and the 'white rain' event of January 2008. This monitoring shall include, at a minimum, ground-dwelling invertebrates, an indicator of copper exposure to birds that feed primarily on invertebrates during critical life stage. This requirement is intended to offset uncertainties in estimated copper exposure between 1,100 and 1,600 mg/kg. Monitoring shall be conducted in intervals no greater than 2 years for the first 5*

*years to establish a trend within the first 5 year review period. Monitoring frequency beyond the first years will be determined based on the first 5 year review recommendations.*

A monitoring work plan will be submitted as part of the basis of design after the ROD is issued. Total estimated cost is \$1.04MM.

## **7.3 Evaluation of Soils Alternatives – pCu**

As described in Section 3, nearly 4,000 acres of upland have been remediated or reclaimed through interim actions and soil removals under reclamation. Of the remaining acres, the flat granular soil category had mean pCu concentrations less than its respective PEL, and no acres of this soil category are retained for remediation. The acres associated with retained rangeland polygons in bedrock, flat rocky, and slope >13% soils types with PEL values of 4.4, 4.98, and 4.11 respectively, amounted to 377 acres, with 47, 313, and 17 acres in bedrock, flat rocky, and slope soils. These acres brought forward for remedial analysis were evaluated across the eight remedial alternatives for the threshold and balancing criteria (Section 7.3.1, below).

As discussed in Section 6, the remedial alternatives are:

- Alternative 1: No Action
- Alternative 2: Monitoring
- Alternative 3: Excavation (Reuse or Disposal) and Monitoring
- Alternative 4: Soil Amendments (Lime), Tilling, and Monitoring
- Alternative 5: Tilling and Monitoring
- Alternative 6: Soil Amendments (Ferrihydrite), Tilling, and Monitoring
- Alternative 7: Soil Cover and Monitoring
- Alternative 8: Surface Soil Controls: Phytostabilization

### **7.3.1 Detailed Analyses of Soil Alternatives – pCu**

This section presents a detailed analysis of the pCu soil remediation alternatives for the plant community for each of the threshold and balancing criteria described in Section 7.1.

#### **7.3.1.1 Alternative 1: No Action**

##### **Protection of Human Health and the Environment**

The No Action alternative does not provide any additional protection of human health and the environment beyond current conditions. A number of acres (313 for flat rocky, 47 for bedrock, and 17 for slope soils) were carried forward for consideration for remediation for exposure units that were below the estimated mean pCu PELs of 4.98, 4.11, 4.11, and 4.40 for the flat rocky, slope, flat granular, and bedrock soil categories, respectively. However, in some flat rocky areas with pCu between 4.6 and 4.98, estimated risk of pCu exposure may be biased high, and the risk may not fully consider other adverse effects to the plant community from remediation (discussed below when comparing the other alternatives). Lower PELs could be technically justifiable, particularly for flat rocky soils, which would mean the No Action alternative could be protective in some of the acres carried forward.

The No Action alternative may be protective of human health and the environment for areas between 4.6 and pCu of 4.98. The reasons are outlined below.

- OAT score not included in PEL calculation. The average of richness and cover was used to estimate the PEL for wildlife habitat, the more sensitive parameter than rangeland condition. This average, empirically measured in the field to represent wildlife habitat, does not adequately incorporate compaction due to grazing that occurs in flat rocky soils. Richness is problematic, because it can increase with an increase in pCu in highly compacted soils; whereas, plant cover does not, as shown by the amendment (Appendix A) and pH monitoring studies (Appendix B), where tiny species increased but plant cover did not respond to an increase in pCu. An average PEL that included the OAT score in its calculation (4.6 for flat rocky soils, 3.83 for bedrock soils, and 2.97 for slope and flat granular soils, Table D-7 in Appendix D), as well as richness and cover likely would better represent the improvement of quality of the habitat for wildlife with a change in pCu.
- Grazing impacts independent of pCu not fully incorporated. Many areas of the STSIU with poor vegetative cover in flat rocky soils have likely *not* been impacted by pCu but rather were unsustainably grazed until the 1970s when management improved on public lands with passage of the Federal Lands Policy and Management Act, which also influenced practices on private lands. Based on Chino's records over 20 years starting in 1984, there was moderate level grazing for the area's productivity, and even though ranching practices were improving, moderate grazing may be unsustainable for the soil types that are present, even if pCu is well above 5. Soil productivity was surveyed by U.S Agricultural Survey and results published in a Soil Conservation Survey (NRCS 1983); the soil types present in the flat rocky portions of the STSIU include Encierro and Muzzler rock complexes, Santana Rock outcrop, and Lonti gravelly loam (Table F-4 of Attachment C). Specifically, NRCS classifies these soil types as "very poor" for grasses (rating of very poor indicates that restrictions for the element or kind of habitat are very severe and that unsatisfactory results can be expected, NRCS 1983). Therefore, these soils generally lack the productivity to support even moderate grazing, and the A horizon of soil types may erode under grazing pressure resulting in a loss of palatable herbaceous cover, as discussed in Schafer and Associates (1999a) and Bestelmeyer et al. (2004). The effect of grazing was supposed to be incorporated into the unacceptability criteria and PELs used to screen acres for this detailed remediation analysis by comparing vegetation in site flat rocky soils to a grazed flat rocky reference area, but the grazed reference area for flat rocky selected was a poor representation of this type of soil, as it was more of an intermediate soil category between bedrock and flat granular types and was a different soil type, Lehman's extremely rocky loam. Acres carried forward may be biased high for flat rocky if the flat rocky reference area is a more productive habitat type.
- Greenhouse results support lower PELs for flat rocky soils. The greenhouse results in Appendix C produce lower PELs than the field community study PELs used in this FS (but they only consider the chemical effect of pCu). Based on the seedling emergence endpoint, which is most correlated to richness and cover in the STSIU plant communities (Appendix C), the PEL would average 4.6 for the native, sensitive species (sideoats grama and tansyaster). The field-sourced seeds actually had lower average PELs (4.4) than the nursery sourced seeds (4.7) for the sensitive species. The average PEL for emergence across all species tested in the greenhouse is 4.4 and the alfalfa PEL alone was 3.7.
- Adjusting greenhouse results to match field conditions still suggests lower PELs for flat rocky soils. The greenhouse results rely on soil samples that were homogenized (in both the 1999 and 2014 studies), and the samples did not retain the structural effects of the soil condition as seen in the field (in bedrock or compacted). The results are less representative of real-world conditions than the plant community results collected in the field for richness, cover, and OAT score. As an example of discrepancies, the greenhouse

results for lower pCu soils often indicated zero emergence; whereas, the field location of the same soils had growing plants. This result suggests native plants in their natural condition are not as sensitive as the greenhouse results indicate. The greenhouse results do not include physical effects of micro-scale heterogeneity of pCu and soil compaction, nor erosion, on vegetation endpoints that may have occurred, either due to grazing or the effects of pCu or both. If such effects allow more emergence, then the greenhouse pCu is overestimated (i.e., PELs biased high). It is possible to statistically account for the effect of pCu on soil condition, where soil condition is calculated as the sum of the three soil factors in the OAT score; however, it requires analyses that separate the soil category effects (e.g., flat rocky vs. other types) from pCu effects when predicting soil condition and then an in-depth modelling method that adds a portion of the structural effect back. Such modelling could be conducted in the future and is not presented herein.

Therefore, given the uncertainties in the effects of pCu on vegetation communities ranging from the mean of all endpoint PELs for each soil category (Table D-7 in Appendix D) to the higher of either the wildlife habitat or rangeland condition PEL for each soil category (method discussed in Section 3.2.4), the No Action alternative may be more protective for the vegetation community as a whole compared to active remedies that may destroy vegetation communities in the short-term (short term may still be decades in semi-arid ecosystems) as discussed in the various remedial alternatives below.

#### **Compliance with ARARs**

There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Compliance with the ARARs is low because some acres do not meet the pre-FS RAC.

#### **Long-term Effectiveness and Permanence**

The magnitude of residual risk (i.e., potential for unacceptable risk to vegetation over the long term) would be consistent with current conditions. There could be changes to pCu in surface soil due to changes in pH over time, either improving or not improving with no action. Therefore, it is unclear if this alternative is acceptable for long term effectiveness or permanence.

#### **Reduction of Toxicity, Mobility, or Volume**

The No Action alternative does not reduce the toxicity, mobility, or volume because no treatment would be implemented.

#### **Short-term Effectiveness**

This alternative is more effective in the short term at protecting the plant community than active remedies, which destroy or damage the plant community. It is less effective at protecting the community in long-term where pCu exposure is the main risk, depending on the clean-up level selected. No actions would be implemented so there would be no risks to the plant community or impacts to the environment during implementation.

#### **Implementability**

No actions would be implemented. This Alternative 1 would not limit further actions at STSIU and no approval or coordination with other agencies would be necessary.

#### **Cost**

There is no cost associated with the No Action alternative. The No Action alternative provides a baseline for comparison of other alternatives.

### 7.3.1.2 Alternative 2: Monitoring

#### **Protection of Human Health and the Environment**

Similar to the No Action alternative, the Monitoring alternative does not provide any additional protection of human health and the environment beyond current conditions. The level of protection for plants is the same as discussed for the No Action Alternative 1, except monitoring provides a mechanism to determine if conditions change and a different action is required. As discussed for the No Action Alternative, the Monitoring alternative may be more protective for the vegetation community in areas of flat rocky soils at pCu between the average PEL of all endpoints (4.6) and the selected PEL in Section 3.2.4 (4.98) because it does not require destruction of the vegetation community during implementation. The net environmental benefit of more active remedies is more certain in flat rocky areas where pCu < 4.6. Similarly, monitoring may be more protective of bedrock and slope soils at pCu values between the average PEL (3.83, and 2.97 for bedrock and slope, respectively) and selected PEL (4.11 and 4.4 for bedrock and slope, respectively) where the effect to the vegetation community is more uncertain than in areas with pCu < 3.83 for bedrock and < 2.97 for slope soil. Additionally, monitoring may be more protective in these soil types because no polygons had bedrock pCu < 3.83 or slope soil pCu < 2.97, where benefits of active remediation would be more certain.

#### **Compliance with ARARs**

There are no chemical-specific ARARs for surface soil (except for the pre-FS RAC). Because no actions would be taken, there are no action-specific or location-specific ARARs for this alternative. The likelihood of compliance with the ARARs on an ongoing basis is low.

#### **Long-term Effectiveness and Permanence**

The magnitude of residual risk (i.e., potential for unacceptable risk to vegetation over the long term) would be consistent with current conditions. There could be changes to pCu in surface soil due to changes in pH over time, either improving or not improving. Therefore, it is unclear if this alternative is acceptable for long term effectiveness or permanence.

#### **Reduction of Toxicity, Mobility, or Volume**

The Monitoring Alternative does not reduce the toxicity, mobility, or volume because no treatment would be implemented.

#### **Short-term Effectiveness**

This alternative is more effective in the short term than active remedies because, unlike active remedies, it does not initially destroy the plant community being protected. No actions would be implemented so there would be no risks to the community or workers or impacts to the environment during implementation.

#### **Implementability**

Monitoring would be readily implementable. Alternative 2 would not limit further actions at STSIU; however, a monitoring plan would need to be coordinated with NMED.

#### **Cost**

The present value cost of this alternative is about \$1.9 million (see Appendix F).

### 7.3.1.3 Alternative 3: Excavation (Reuse or Disposal) and Monitoring

#### Protection of Human Health and the Environment

This alternative would remove soils with concentrations at levels that could pose a potential risk to vegetation from copper and pH (quantified as pCu). The removal of the topsoils could leave unproductive, compacted soils underneath that may be difficult for vegetation growth to establish quickly in this semi-arid environment, even after borrow is placed and the area seeded. Removal of soils has been effective at producing productive plant communities in interim action areas using such techniques (e.g., Arcadis 2014c) although not enough time has elapsed to establish the original plant communities. Also, it is also important to underscore the net environmental benefit of this remedial alternative. The plants are directly destroyed in the remediation, and the harm from destruction of the plants “to benefit plants” and the potentially long recovery period required for the entire plant community to return to and potentially exceed pre-remediation habitat function in the semi-arid southwest might outweigh the effects of pCu on the vegetation community. Protectiveness of the plants as a wildlife and livestock food source is an important criterion to consider in the FS, which may be lost until the community recovers from remediation. Any active remedy such as excavation of flat rocky soils would destroy the entire community, including both less sensitive species that are thriving and more sensitive species that would benefit in the long term. The entire community may require a significant period of time to return to pre-remediation conditions. Site-specific data show that pCu has a significant effect on the plant community. However, as discussed in the 7.3.1.1, there is some uncertainty in the degree of protection provided at the selected PELs versus PELs that incorporate the mean effects to both wildlife habitat and the less sensitive rangeland condition endpoint. Active remedy conducted at pCu between 4.6 and 4.98 for flat rocky soils (and between 3.83 and 4.4 for bedrock and between 2.97 and 4.11 for slope soils) may improve soil conditions for more sensitive species, but there may be a limited net benefit to wildlife habitat overall because all vegetation species are simultaneously destroyed the implementation of the action; the potential for effect from pCu in that range is uncertain. No areas of bedrock soils with pCu < 3.83 or slope soils with pCu < 2.97, bedrock and slope areas with more certain benefits from this Alternative, were identified. The number of acres identified for active remedy in flat rocky soils reduces 3-fold between PELs of 4.98 to 4.6, see Table 7-3 (below).<sup>16</sup> A lower PEL than 4.98 for flat rocky is technically justifiable as discussed in Section 7.3.1.1 and using a slightly lower PEL (e.g., 4.6) that incorporates both the wildlife habitat and rangeland condition endpoints would disturb less land and may be more beneficial to the overall ecosystem. A lower PEL would be consistent with the phytotoxicity and community study results overall and incorporate effects on the soil structure due to grazing and land practices over time.

Once the ecosystem re-establishes, eventually maturing to its previous successional stage (possibly after decades in this semi-arid ecosystem), Alternative 3 would meet the threshold criterion of overall protection of human health and the environment. The same is true for bedrock and slope soils, since the plant communities will be destroyed until recovery and a slightly lower PEL may be more protective.

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<sup>16</sup> Table has removed acceptable wildlife habitat or acceptable rangeland. The already remediated Bolton Diversion is part of a rangeland polygon HE-216 retained because pCu is still lower than the PEL even with that area averaged in with the other flat rocky acres (see Figure 3-12 vs. Figure 3-13). However, its acres are excluded from the acres in Table 7-3 because it will not require remediation.

**Table 7-3. Flat Rocky Acres less than PELs ranging from 4 to 4.98**

pCu	Acres
4	3.4
4.1	3.4
4.2	67.1
4.3	67.1
4.4	87.3
4.5	87.3
4.6	112.5
4.7	164.2
4.8	187.1
4.9	312.8
4.98	312.8

**Compliance with ARARs**

There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Action-specific and location-specific ARARs (Tables 2-2 and 2-3) would be met by remedial design. Alternative 3 is expected to meet the threshold criterion of compliance with ARARs if careful attention is given to restoring the soil productivity and plant community after soil removal.

**Long-term Effectiveness and Permanence**

This alternative would provide protection over the long term for soil categories that can be excavated. Removal is a proven and reliable method of preventing the potential for contact with materials of concern. Monitoring of erosion from removal areas and soil sampling would be required to ensure that recontamination did not occur. However, as described above, excavation at pCu at < PELs for the soil categories would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species may not return for an extended period of time. Alternative 3 would provide a high level of long-term effectiveness and permanence after species return.

**Reduction of Toxicity, Mobility, or Volume**

This alternative would not meet the statutory preference for treatment. Alternative 3 does not reduce the toxicity, mobility, or volume because no treatment would be implemented; only contaminated soil will be removed.



### **Short-term Effectiveness**

Alternative 3 would be ineffective at achieving the goal of protecting the vegetation community, wildlife habitat, and rangeland in the short term. As described above, excavation would destroy the entire community, including both less sensitive and more sensitive species, and many species may not return for an extended period of time. When active remedy occurs at < PELs, the whole community is affected by this remedy. While there is some gain to improve soil conditions for more sensitive species, there is limited short-term effectiveness because all species are simultaneously destroyed. Also, Alternative 3 would entail significant earthwork which would result in a risk to workers due to physical hazards, particularly working on steep and potentially unstable slopes. Placing heavy equipment on slopes to regrade them could increase the risk of slope failure. Site preparation, soil excavation, transport and disposal onsite are significant construction activities that will require transport of heavy equipment to the site. In addition, the use of heavy equipment for grading and earthwork would release diesel emissions into the environment.

### **Implementability**

This alternative is considered moderately implementable. Excavation would not be implementable in areas of bedrock outcrops or steep slopes. Of the 377 acres, 47 are on bedrock and 14 are on steep slopes. The presence of these two physical factors limits the implementability of this alternative in these areas. Soil excavation, transportation and onsite disposal can be achieved in flat rocky soils with standard equipment and materials. Site preparation will include building new roads for haulage. Excavated soil would either be taken to the operational area for subsequent reuse or taken to the West Stockpile for recycling (disposal option).

### **Cost**

The present value cost of this alternative is \$11.1 Million (see Appendix F).

## **7.3.1.4 Alternative 4: Soil Amendments (Lime), Tilling, and Monitoring**

### **Protection of Human Health and the Environment**

The ability of this alternative to satisfy this protection criterion is high on relatively level habitats, particularly if on poor rangeland. Success of this remedial technology in protecting the plant community was defined in the amendment study (Appendix A) as an increase in soil pCu, decrease in copper concentration in plant tissue, an increase in plant species richness and cover, an increase in rangeland grass cover, and a reduction in undesirable plant species (toxic, non-native). Lime (in form of white rain) alone increased pCu, decreased plant uptake of copper, and increased species richness but not cover. Tilling has been shown to be effective on the site at also increasing cover (e.g. for ripped roads, see Section 5.3.3.3 in Appendix A). Copper present in surface soils would be tilled and mixed deeper into the soil along with bringing higher pH soils to the surface. Tilling also provides the benefit of breaking up the soil to allow plant growth. The addition of lime in areas with tilling or where tilling is impractical could also increase pCu in soils and reduce the bioavailability of any remaining copper. Notably, the amendment study found the increase in pCu in steep slopes and benefit to plant community was less than in flatter slopes, likely due to runoff of the lime. Also, lime alone in form of white rain did not substantially increase plant or grass cover, but did improve richness. Thus, lime alone is expected to have a low to moderate ability to satisfy this criterion.

Similar to Alternative 3, it is important to underscore the net environmental benefit of this remedial alternative. The plants are directly destroyed in the tilling and are adversely impacted even with only lime amendment application (see Appendix A results). The destruction of the plants “to benefit plants function as wildlife habitat” and the

potentially long recovery period required for the entire plant community to return to and potentially exceed pre-remediation habitat function in the semi-arid southwest might outweigh the impact of effects of pCu on the vegetation community, depending on the pCu of the PELs. Protectiveness of the plants as a food source is an important criteria to consider in the FS, which may be lost until the community recovers from remediation. Any active remedy such as tilling of flat rocky soils would destroy the entire community, including both less sensitive species that are thriving and more sensitive species that would benefit in the long-term. The entire community may require a significant period of time to return to pre-remediation conditions. Site-specific data show that pCu has a significant effect on the plant community; however, there is some uncertainty in the degree of protection provided at the selected PELs versus PELs that incorporate the mean effects to both wildlife habitat and the less sensitive rangeland condition endpoint. Active remedy conducted at pCu between 4.6 and 4.98 for flat rocky soils (and between 3.83 and 4.4 for bedrock and between 2.97 and 4.11 for slope soils) may improve soil conditions for more sensitive species, but there may be a limited net benefit to wildlife habitat overall because all plant species are simultaneously affected by the action and the potential for effect from pCu in that range is uncertain. A lower PEL than 4.98 for flat rocky could be technically justifiable as discussed in Section 7.3.1.1 and using a lower PEL (e.g., 4.6) would disturb less land and may be more beneficial to the overall ecosystem. A lower PEL would be consistent with the phytotoxicity study and incorporate changes to the soil structure due to grazing and land practices over time. Once the ecosystem re-establishes, Alternative 4 would meet the threshold criterion of overall protection of human health and the environment.

### **Compliance with ARARs**

There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Action-specific and location-specific ARARs (Tables 2-2 and 2-3) would be met by remedial design. Alternative 4 would likely meet the threshold criterion of compliance with ARARs if tilling is included. It is less certain on steeper slopes where lime has run off and has been less effective. Tilling breaks up the compacted soils of flat rocky soils and is likely to improve the plant community in such soils, although pCu may not always reach the flat rocky PEL of  $pCu \geq 4.98$  because, although pH may increase, copper may not be reduced enough (see Amendment study in Appendix A where pCu change was not statistically significant). The breaking up of the compacted soils is the main benefit, and the compaction likely is due to grazing and/or pCu effects.

### **Long-term Effectiveness and Permanence**

The long-term effectiveness of this alternative to meet some if not all of these metrics is considered moderate for amendments for improving pCu but high for tilling for improving the plant community if in flat rocky soils. Lime (calcium carbonate) has been shown to significantly increase pH values at the site (Appendix A) but does not always increase pCu values in flat rocky soils at the site to pCu of 4.98. As described above, tilling at  $pCu < 4.98$  for flat rocky soils would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species may not return for an extended period of time. Alternative 4 would provide a high level of long-term effectiveness and permanence after species return. Its effectiveness in bedrock and on steep slopes at  $pCu < 4.4$  and  $< 4.11$ , respectively is low, given tilling cannot be done in bedrock and tilling equipment is difficult to use on steep slopes. Liming may be more effective in bedrock and steep slopes, but the effectiveness is uncertain because lime is easily washed off with rainfall in these areas. Additional applications of soil amendments may be required in bedrock outcrops or on steep slopes due to transportation of material from storm water flows.

### **Reduction of Toxicity, Mobility, or Volume**

This alternative would mix the upper soil layers with underlying soils, increasing pCu and decreasing toxicity and amount of acid in the soils. Also, adding lime would reduce the bioavailability and mobility and thus toxicity of the copper.

### **Short-term Effectiveness**

This alternative would not be considered effective in the short-term. Mixing of surface soils with deeper soil layers resulting from tilling is expected to increase pH, decrease soil copper, and improve conditions for plant growth. However, it may take several years for the soil amendments to increase the pCu values. Also, as described above, tilling or liming at pCu of < 4.98 in flat rocky soils (and < 4.4 in bedrock and < 4.11 in slope soils) would destroy or impact the entire community, including both less sensitive and more sensitive species, and many species may not return for an extended period of time. If lime is only applied (no tilling), as it was on steeper slopes in the amendment study (only up to 13% slope), it may still adversely impact the plant community in the short term (observed during amendment study), and only slightly increase the pCu. The benefit is likely greater for steep slopes in poor rangeland condition. The three step approach this FS used to identify acres for remediation of pCu often, but not always removed fair to good rangeland condition areas from consideration, increasing the likelihood that lime treatment alone might be effective, particularly on relatively level acres, and it may be somewhat effective on bedrock in the short term where tilling cannot occur (but may wash off with rainfall greatly reducing its short-term benefit). When active remedy occurs in areas < PELs, the whole community is adversely affected by the remedy in the short term. While there is some gain to improve soil conditions for more sensitive species, there is limited short-term effectiveness because all species are simultaneously destroyed by tilling or reduced by lime application if no tilling occurs on steeper slopes.

### **Implementability**

The implementability of soil amendments is considered high compared to the other alternatives. If tilling is excluded, it is an alternative that could be applied to areas with steep slopes and bedrock outcrop areas, which cannot be addressed by physical means. However, additional applications of soil amendments may be required in the future in bedrock outcrops or steep slopes due to the transportation of amendments in storm water. Tilling is considered moderately implementable as it could not be used in areas of bedrock outcrops or steep slopes but is effective in flat rocky soils.

### **Cost**

The present value cost of this alternative is \$6.4 million (see Appendix F).

## **7.3.1.5 Alternative 5: Tilling and Monitoring**

### **Protection of Human Health and the Environment**

The ability of this alternative to satisfy this criterion is high for slopes that are not too steep. As stated in the amendment study that evaluated this particular remedial technology over a 5-year period, success of this remedial technology in protecting the environment is defined as an increase in soil pCu, decrease in copper concentration in plant tissue, an increase in plant species richness and cover, an increase in rangeland grass cover, and a reduction in undesirable plant species (toxic, non-native). The results showed success in these factors except the study was not long enough to evaluate reduction in undesirable plant species (which increased in first five years). Tilling alone has demonstrated effectiveness on the site of improving grass cover in flat rocky soils (e.g. for ripped roads, see Section 5.3.3.3 in Appendix A). While natural attenuation rates for pCu have not

been determined, tilling would likely increase the rate of attenuation. Copper present in surface soils would be tilled and mixed deeper into the soil along with bringing higher pH soils to the surface, attenuating the bioavailable copper concentration. Tilling also provides the benefit of breaking up the soil to allow plant growth. Tilling in relatively level locations was the most effective. Tilling was not tested on steeper slopes because of difficulty operating the equipment on steeper slopes. This alternative is not likely protective in steeper sloped areas or on bedrock, which account for 64 acres carried forward to evaluate for remediation. Tilling is most beneficial in flat rocky soils, where soil compaction is an issue preventing grass establishment.

Similar to Alternative 4, it is important to underscore the net environmental benefit of this remedial alternative. The plants are directly destroyed in the tilling, and the destruction of the plants “to benefit plants” and the potentially long recovery period required for the entire plant community to return to and potentially exceed pre-remediation habitat function in the semi-arid southwest (especially with possibly undesirable species establishing first) might outweigh the impact of effects of pCu on the vegetation community, depending on the pCu of the PELs. Protectiveness of the plants as a wildlife and livestock food source is an important criterion to consider in the FS, which may be lost until the community recovers from remediation. Any active remedy such as tilling of flat rocky soils would temporarily destroy the entire community, including both less sensitive species that are thriving and more sensitive species that would benefit in the long-term.

Site-specific data show that pCu has a significant effect on the plant community; however, there is some uncertainty in the degree of protection provided at the selected PELs versus PELs that incorporate the mean effects to both wildlife habitat and the less sensitive rangeland condition endpoint. Active remedy conducted at pCu between 4.6 and 4.98 for flat rocky soils (and between 3.83 and 4.4 for bedrock and between 2.97 and 4.11 for slope soils) may improve soil conditions for more sensitive species, but there may be a limited net benefit to wildlife habitat overall because all species are simultaneously affected by the action and the potential for effect from pCu in that range is uncertain. The number of acres identified for active remedy in flat rocky soils reduces 3-fold between PELs of 4.98 to 4.6 (Table 7-3). A lower PEL for flat rocky is technically justifiable as discussed in Section 7.3.1.1 and using a lower PEL (e.g., 4.6) would disturb less land and be more beneficial to the overall ecosystem. A lower PEL would be consistent with the phytotoxicity study and would incorporate effects on the soil structure due to grazing and land practices over time. Once the ecosystem re-establishes, Alternative 5 would meet the threshold criterion of overall protection of human health and the environment.

### **Compliance with ARARs**

There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Action-specific and location-specific ARARs (Tables 2-2 and 2-3) would be met by remedial design. Alternative 5 may meet the threshold criterion of compliance with ARARs. The likelihood of achieving the desired pCu ( $\geq$  the PELs) is moderate with tilling but the likelihood of achieving a more diverse community with greater plant cover is high.

### **Long-term Effectiveness and Permanence**

Long-term effectiveness and permanence is similar to Alternative 4. It is expected that pCu values would increase over time with mixing with deeper soils with lower copper and higher pH, that would result from tilling. Monitoring with soil sampling would be required to ensure that recontamination did not occur. As described above, tilling at pCu of  $<$  the soil-specific PELs would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species would not return potentially for an extended period of time. However, Alternative 5 would provide a high level of long-term effectiveness in the very long term and be permanent after species return.

### **Reduction of Toxicity, Mobility, or Volume**

This alternative would mix the upper soil layers with underlying soils, decreasing copper concentrations and decreasing the amount of acidity in the soil. Toxicity in the soils would reduce. Mobility also may be reduced.

### **Short-term Effectiveness**

This alternative would not be considered effective in the short-term. Mixing of surface soils with deeper soil layers resulting from tilling is expected to increase pH, decrease soil copper, and improve conditions for plant growth. However, it may take several years for the tilling to increase the pCu values. That said, as described above, tilling in flat rocky (only soil category that this technology would be effective in of the three categories being evaluated) would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species may not return for an extended period of time. When active remedy occurs at pCu < the PELs, the whole community is affected by the remedy. While there is some gain to improve soil conditions for more sensitive species, there is limited short-term effectiveness because all species are simultaneously affected.

### **Implementability**

The implementability of this alternative is considered moderate. As with Alternative 3, tilling could not be used in areas of bedrock outcrops or steep slopes. Soil tilling can be achieved with standard equipment and materials. No extensive site preparation would be required.

### **Cost**

The present value cost of this alternative is about \$5.9 million (see Appendix F).

## **7.3.1.6 Alternative 6: Soil Amendments (Ferrihydrite), Tilling, and Monitoring**

### **Protection of Human Health and the Environment**

The ability of this alternative to satisfy this criterion is possibly similar to Alternative 4, but unlike lime, the effectiveness of ferrihydrite has not been tested on the site. Tilling has been shown to be effective on the site (e.g. for ripped roads, see Section 5.3.3.3 in Appendix A), as discussed above. Copper present in surface soils would be tilled and mixed deeper into the soil along with bringing higher pH soils to the surface. Tilling also provides the benefit of breaking up the soil to allow plant growth. The addition of ferrihydrite with tilling or in areas where tilling is impractical would be expected to decrease free cupric ion activity in soils and reduce the bioavailability of any remaining copper. Similar to Alternative 4, however, it is important to underscore the net environmental benefit of this remedial alternative. The plants are directly destroyed in the tilling, and the destruction of the plants “to benefit plants” and the potentially long recovery period required for the entire plant community to return to and potentially exceed pre-remediation habitat function in the semi-arid southwest might outweigh the impact of effects of pCu on the vegetation community, depending on the pCu of the PELs. Protectiveness of the plants as a food source for wildlife and livestock is an important criteria to consider in the FS, which may be lost until the community recovers from remediation. Any active remedy such as tilling of flat rocky soils would destroy the entire community, including both less sensitive species that are thriving and more sensitive species that would benefit in the long-term. The entire community may require a significant period of time to return to pre-remediation conditions.

Site-specific data show that pCu has a significant effect on the plant community; however, there is some uncertainty in the degree of protection provided at the selected PELs versus PELs that incorporate the mean effects to both wildlife habitat and the less sensitive rangeland condition endpoint. Active remedy conducted at

pCu between 4.6 and 4.98 for flat rocky soils (and between 3.83 and 4.4 for bedrock and between 2.97 and 4.11 for slope soils) may improve soil conditions for more sensitive species, but there may be a limited net benefit to wildlife habitat overall because all species are simultaneously affected by the action and the potential for effect from pCu in that range is uncertain. The number of acres identified for active remedy in flat rocky soils reduces 3-fold between PELs of 4.98 to 4.6 (Table 7-3 above). A lower PEL for flat rocky is technically justifiable as discussed in Section 7.3.1.1 and using a lower PEL (e.g., 4.6) would disturb less land and might be more beneficial to the overall ecosystem. A lower PEL would be consistent with the phytotoxicity study and incorporate changes to the soil structure due to grazing and land practices over time. Once the ecosystem re-establishes, Alternative 6 would meet the threshold criterion of overall protection of human health and the environment.

### **Compliance with ARARs**

There are no chemical-specific ARARs for surface soil, except for the pre-FS RAC, which might be met. Action-specific and location-specific ARARs (Tables 2-2 and 2-3) would be met by remedial design. The likelihood that Alternative 6 would meet the threshold criterion of compliance with ARARs is considered moderate for ferrihydrite alone and high if combined with tilling.

### **Long-term Effectiveness and Permanence**

The long-term effectiveness of this alternative is considered moderate to high to increase pCu values. Ferrihydrite has been shown to reduce free cupric ion activity at similar sites and is less mobile than lime and thus may remain longer on steep slopes and be more effective than lime when used alone on steeper slopes. However, additional applications of soil amendments still may be required in bedrock outcrops or on steep slopes due to transportation of material from storm water flows. As described above, tilling at pCu of < 4.98 for flat rocky soils, < 4.4 for bedrock soils, and < 4.11 for bedrock soils would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species may not return for an extended period of time. Alternative 6 would provide a high level of long-term effectiveness and permanence only after species return. Because the beneficial effects of tilling in flat rocky soils (the soil type in which tilling is most effective) at pCu between 4.6 and the PEL (4.98) is uncertain, the long-term effectiveness of tilling in those areas with pCu between 4.6 and 4.98 may be lower than in the areas with pCu < 4.6 in flat rocky soils.

### **Reduction of Toxicity, Mobility, or Volume**

This alternative would mix the upper soil layers with underlying soils, increasing pCu and thus reducing toxicity and amount of acid in the soils, plus adding ferrihydrite would reduce the bioavailability and mobility and thus toxicity of the copper.

### **Short-term Effectiveness**

This alternative would not be considered effective in the short-term. Mixing of surface soils with deeper soil layers resulting from tilling with soil amendments is expected to increase pH, decrease soil copper, and improve conditions for plant growth. However, it may take several years for the soil amendments to increase the pCu values and the increase may be small if copper concentrations are high. As described above, tilling in areas at pCu of < PELs would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species may not return for an extended period of time. When active remedy occurs at < PELs, the whole community is affected by the remedy. While there is some gain to improve soil conditions for more sensitive species, there is limited short-term effectiveness because all species are simultaneously affected.

### **Implementability**

The implementability of soil amendments is considered high. Like alternative 4, use of the ferrihydrite in this alternative means the treatment could be applied to areas with steep slopes and bedrock outcrop areas, which cannot be addressed by physical means. Additional applications of soil amendments may be required in the future in bedrock outcrops or steep slopes due to the transportation of amendments in storm water. Tilling is considered only moderately implementable as it could not be used in areas of bedrock outcrops or steep slopes but should be effective in flat rocky soils.

### **Cost**

The present value cost of this alternative is \$6.4 million (see Appendix F).

## **7.3.1.7 Alternative 7: Soil Cover and Monitoring**

### **Protection of Human Health and the Environment**

Soil cover would reduce but not eliminate plant contact with the contaminants. Also, the pathways would only be incomplete if the soil cover remains in place. The soil cover could be breached by digging, construction, or similar activities. This could potentially pose a risk to the vegetation. Most importantly, many plants have deep roots and would still access the contaminated soil beneath the cover, and those plants may have reduced uptake but still be exposed. It is important to also underscore the net environmental benefit of this remedial alternative. The plants are directly destroyed during the soil cover process. In soils with pCu between the selected PELs and the lower PELs identified when thresholds for rangeland wildlife habitat are averaged (the “uncertain” pCu range), the net environmental benefit may be lower than alternatives without destructive remediation, as soil cover destroys the entire community, including both less sensitive species that are thriving and more sensitive species that would benefit. The destruction of the plants “to benefit plants” and recovery period for an undetermined but possibly long time frame required for the entire plant community in the semi-arid southwest might outweigh the effect of the pCu on the plants in the uncertain pCu range. Protectiveness of the plants as a food source for wildlife and livestock is an important criterion to consider in the FS, which may be lost until the community recovers from remediation. Any active remedy such as implementing a soil cover would initially destroy the entire community, including both less sensitive and more sensitive species. When active remedy occurs in areas < PELs, the whole community is affected by the remedy. While there is some gain to improve soil conditions for more sensitive species, there may be limited net environmental gain overall for areas in the uncertain pCu range because all species are simultaneously destroyed when the cover is placed, whether in bedrock, steep slopes or on flat rocky soils. A lower PEL in flat rocky soils is more consistent with the phytotoxicity study and incorporates changes to the soil structure due to grazing and land practices over time. Covering bedrock and steep areas with soil in areas < PELs may not be effective enough to restore the community due to soil erosion, and may not be beneficial until pCu is at the lower end of the uncertain PEL range (and no areas have pCu below the lower alternative PELs for bedrock and slope soils). Once the ecosystem re-establishes, Alternative 7 would meet the threshold criterion of overall protection of human health and the environment.

### **Compliance with ARARs**

There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Action-specific and location-specific ARARs (Tables 2-2 and 2-3) would be met by remedial design. Alternative 7 would meet the threshold criterion of compliance with ARARs.

### **Long-term Effectiveness and Permanence**

A soil cover would require long-term operations, maintenance, and monitoring to continue to be effective. A soil cover would be considered a moderately effective solution. As described above, a soil cover placed in soils with pCu < soil-specific PELs would destroy the entire community, including both less sensitive and more sensitive species, and even with seeding, many species may not return for an unknown, but likely extended period of time. Because the beneficial effects of placing a soil cover at pCu between 4.6 and the PEL (4.98) for flat rocky soils and between 3.83 and the PEL (4.4) for bedrock and between 2.97 and the PEL (4.11) for slope soils is uncertain, the long-term effectiveness of cover placement in those areas with uncertain benefits may be lower than in the areas with pCu < the lower end of the pCu range (determined by means of the three vegetation endpoints)..Alternative 7 would provide a high level of long-term effectiveness and permanence after species return.

### **Reduction of Toxicity, Mobility, or Volume**

This alternative would not meet the statutory preference for treatment. Alternative 7 does not reduce the toxicity, mobility, or volume because no treatment would be implemented; only contaminated soil would be covered.

### **Short-term Effectiveness**

This alternative is not considered to be effective in the short-term. It would address all of the pCu areas < PELs, but, as described above, a soil cover placed over acres with pCu below the PELs would temporarily destroy the entire plant community, including both less sensitive and more sensitive species, and, even after seeding, many species may not return for an extended period of time. When active remedy occurs at the PELs, the whole community is affected by the remedy. While there is some gain to improve soil conditions for more sensitive species, there is no short-term effectiveness because all species are simultaneously affected. Vegetation in borrow areas would be adversely impacted in the short-term, too.

### **Implementability**

This alternative is considered moderately implementable. Unlike excavation, placing a soil cover could be implementable in areas of bedrock outcrops but would be difficult in steep slopes, and might erode off of bedrock, also. Covering bedrock on 47 acres would require large amounts of soil and large borrow areas. Of the 377 acres addressed for pCu exposure, 14 acres are on steep slopes. More maintenance is required with this option than for other alternatives.

### **Cost**

The present value cost of this alternative is \$13.2 million (see Appendix F).

## **7.3.1.8 Alternative 8: Surface Soil Controls – Phytostabilization**

### **Protection of Human Health and the Environment**

The overall protection of this alternative would be low. Contaminant reduction from planting vegetation in order to contain contaminants within the soils and plant roots would not necessarily lower pCu in the area needing remediation nor help reduce toxicity to the plants, although it may reduce erosion and wind dispersion of contaminated soil to adjacent areas or into drainages. It is also important to underscore the net environmental benefit of this remedial alternative. The existing plants may be directly destroyed and replaced during the phytostabilization and planting process. The destruction of the plants “to benefit plants” and unknown but potentially long recovery period required for the native plant community to return in the semi-arid southwest likely



outweighs the benefit of stabilizing the soil, because stabilization does not reduce phytotoxicity in the targeted area nor is it likely to increase pCu in the area stabilized. Protectiveness of the plants as a food source for wildlife and livestock is an important criterion to consider in the FS, which may be lost until the desired plants of the community that remediation is supposed to protect recover from the disruption from the remediation. Any active remedy such as implementing phytostabilization would temporarily disrupt the entire community, including both less sensitive and more sensitive species. While there is some gain to improve soil stability, there is no net environmental gain overall because existing species are disrupted during the planting and low pCu is not effectively addressed. Overall benefit is unlikely within the uncertain PEL range between the selected PEL and alternative lower PELs discussed in previous sections. The lower PEL for flat rocky is technically justifiable as discussed in Section 7.3.1.1 and using a low PEL (e.g., 4.6) would disturb less land and be more beneficial to the overall ecosystem. A lower PEL for bedrock and slope soils than the selected PELs is also justifiable because the benefit of soil stabilization, which is very minor for pCu, likely does not outweigh the temporary disruption to the plant community. Once the ecosystem re-establishes, Alternative 8 would still not meet the threshold criterion of overall protection of human health and the environment.

### **Compliance with ARARs**

There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Action-specific and location-specific ARARs (Tables 2-2 and 2-3) would be met by remedial design, but Alternative 8 would not meet the threshold criterion of compliance with ARARs because it is unlikely to meet the pre-FS RAC.

### **Long-term Effectiveness and Permanence**

The long-term effectiveness for this alternative is considered to be very low compared to the other alternatives. Phytostabilization has the potential to be effective at stabilizing the soils and reducing transport of impacted soils. However, this technology used alone would not be effective at reducing site contaminant levels and/or exposure to plants the remediation is supposed to protect.

### **Reduction of Toxicity, Mobility, or Volume**

This alternative would not meet the statutory preference for treatment. Alternative 8 does not reduce the toxicity, mobility, or volume because no treatment would be implemented.

### **Short-term Effectiveness**

This alternative would be considered not to be effective in the short-term because of the length of time required for the disrupted vegetation community to become re-established relative to an uncertain benefit in that the pCu onsite may not improve. Additionally, as described above, phytostabilization would destroy many species in the community and many species may not return for an extended period of time. When this active remedy occurs at < PELs, species insensitive and sensitive to pCu are affected by the planting of other species used for phytostabilization. While there is some gain to improve soil stability, there is no short-term effectiveness because the community is disrupted by the planting and remedial activities.

### **Implementability**

Revegetating soils with species that are targeted at soil stabilization is generally considered implementable at the Site. There may be certain areas of the site that may not support the phytostabilization species (due to slopes, percent soil coverage, and soil conditions). Plant species and potential locations would be determined during the remedial design process.

## **Cost**

The present value cost of this alternative is \$6.6 Million (see Appendix F).

### **7.3.2 Comparative Analyses of Soil Alternatives - pCu**

A comparison of soil alternatives for pCu is provided in Table 7-2. Although Alternatives 1 and 2 are the greenest of the remediation alternatives considered, they do not meet the threshold criteria because the acres do not meet the Pre-FS RAC for pCu and, therefore, do not meet the state's criteria for reduction of risk to vegetation. Chino carried forward for remedial consideration 377 acres associated with the estimated PELs of 4.98, 4.11, 4.11, and 4.40 for the flat rocky, slope, flat granular, and bedrock soil categories, respectively. Lower PELs by soil type could be technically justifiable, particularly for flat rocky soils. Lower PELs that average in effects of rangeland condition that can be technically justifiable are 4.6, 2.97, 2.97, and 3.83 for flat rocky, slope, flat granular, and bedrock soil categories, respectively. Of the 377 acres carried forward, the No Action or Monitoring alternative may be protective of human health and the environment in areas where remedial benefits are uncertain, which is between the selected PEL and the alternative lower PELs. In areas with pCu below the lower PELs, benefits of active remedy are more certain.

Alternative 2 would not satisfy the criteria of "overall protection"; however, due to the presence of naturally occurring calcium carbonate (which may occur in Gila conglomerate and other areas), natural attenuation may be occurring at some areas of the site because as wind and erosion occur, the calcium carbonate in dust particles can buffer the soil pH over time and increase pCu. The rate of natural attenuation is currently unknown but was enhanced by the 2008 white rain, which deposited lime on the soil that also buffers the soil pH. The long-term effectiveness of the monitoring alternative is dependent on the unknown natural attenuation rate in the future. Based on the amendment, pH and insect studies, the white rain in 2008 decreased cupric ion activity (increased pCu), reduced copper uptake into plants and thus reduced phytotoxicity, and improved plant richness. Without active remediation, improvements to low pCu soils beyond the improvement in 2008 from the white rain, however, are not anticipated in the short-term to meet the RAOs. Total estimated cost for Alternative 2 is \$1.8 million.

Alternative 3, which is not considered to be a green remediation alternative, involves removal and monitoring of low pCu soils for areas indicated for remediation of pCu in Figure 8-1. This alternative would meet overall protection for ecological receptors and the targeted receptor of the plant community if measures are taken to ensure soil quality remaining is high by tilling if the soil is compacted and by seeding or by implementing other ways of improving the soil; however, excavation initially physically destroys habitat which must be restored (Redente 2004). The effectiveness of this remedial alternative in the long-term is high because it increases pCu values and, as long as the excavation leaves tilled, granular soils, should benefit the plant community. Once historically deposited soils are excavated, the source of the material would be removed; however, short term monitoring and sampling may be required to document that the excavation was sufficient. This alternative does not reduce toxicity, mobility, and volume of the concentration of pCu in surface soil because there is no treatment and this alternative is also not considered effective in the short-term because the remedy itself destroys the very receptor it aims to protect and though seeding can accelerate return, it will take an undetermined, but likely extended period of time to mature into the plant community that was lost. The implementability of this alternative is considered moderate compared to other alternatives. Total estimated cost is \$11.1 million.

Alternative 4 is amendments (lime) coupled with tilling and monitoring of low pCu soils. This alternative would meet overall protection for vegetation. Amendments include lime but not organic matter. The long-term effectiveness of this alternative is considered moderate to high to increase pCu values. While lime (calcium

carbonate) has been shown to increase pCu values at the site, Arcadis (2017) indicates that liming is recommended for soils with pH < 2 but liming is not that helpful above that pH, and soil pH identified herein are not less than pH 2. However, it is uncertain if liming would be ineffective at pH > 3 in other soils than used in the amendment study and tilling plus lime may be beneficial. Organic matter was not recommended after the amendment study was completed because cow manure brought in weedy plants that degraded the habitat; however, other forms of organic matter could be considered but are not because lime can be effective without it. This alternative is not effective in the short-term as it may take several years for the soil amendments to increase the pCu values and vegetation must re-establish when destroyed during tilling or damaged with lime slurry amendments sprayed on them, and it takes an undetermined but potentially extended period of time to return to a mature community. The tilling aspect of the remedy initially destroys the plant community. The long-term effectiveness of this alternative is considered moderate to high (is high in non-steep soils) for increasing pCu values but has been found to be most effective at improving the grass cover due to breaking up compacted soils. Toxicity would be decreased with increase in pCu. The implementability of tilling is high in flat rocky soils, but not implementable in bedrock and steep soils. Implementability of soil amendments is considered high compared to the other alternatives; however, additional applications may be required if steep slopes or bedrock are encountered that experience high storm water runoff. This alternative is considered to be a moderately green remediation alternative. Total estimated cost is \$6.4 million.

Alternative 5, which is a moderately green remediation alternative, is tilling and monitoring of low pCu soils. This alternative would meet overall protection for vegetation. The ability of Alternative 4 to satisfy the criteria of “overall protection” is high in non-steep and non-bedrock areas that can be tilled. Tilling mixes cleaner soils with surface soils, reducing copper concentrations and can expose more of the soil to calcium carbonates at depth if calcium carbonates are present. The long-term effectiveness of this alternative is considered moderate for increasing pCu values but has been found to be most effective at improving the grass cover due to breaking up compacted soils. This alternative would decrease concentrations by mixing the topsoil layers with underlying soils, and mobility and toxicity may be reduced. That said, this remedy is not effective in the short term as tilling initially destroys the plant community. Implementability in flat rocky soils is better than Alternative 3 because less equipment and movement of soil are required but, because it does not include amendments, is not implementable in bedrock and steep soils. Total estimated cost is \$5.8 million, which is roughly half the cost of Alternative 3.

Alternative 6 is amendments (ferrihydrite) coupled with tilling and monitoring of low pCu soils and this alternative would meet overall protection goals for vegetation. The amendment is ferrihydrite. The long-term effectiveness of this alternative is considered moderate to high to increase pCu values, as ferrihydrite has been shown to reduce cupric ion activity at similar sites. However, liming the soils (Alternative 4) would achieve similar results, is more readily available, and the dosing is easier to determine without the need for a pilot study, which ferrihydrite would require. Organic matter was not recommended after the amendment study was completed because cow manure brought in weedy plants that degraded the habitat; however, other forms of organic matter could be considered but lime in white rain has been found to be effective without organic matter added. Similar to lime amendments and tilling, this alternative with ferrihydrite and tilling would not be considered moderately effective in the short-term because it initially destroys the plant community, and the amendments can damage the community initially. The implementability of soil amendments is considered high compared to the other alternatives; additional applications may be required if steep slopes and bedrock are encountered. This alternative is considered to be a moderately green remediation alternative. Total estimated cost is \$6.4 million.

Alternative 7 includes the installation of a soil cover to protect plants and monitoring of low pCu soils and the cover system. This alternative might meet overall protection for vegetation, but plants with deeper roots may still

access the contaminated soils. The overall protection to plants of this alternative could be moderate as it protects shallower-rooted plants but not deep-rooted plants, but those deep-rooted plants may have more tolerance to the low pCu (uncertain). The cover also has the advantage of eliminating direct human and avian contact with the contaminants. The long-term effectiveness for this alternative for protecting plants is considered at most to be moderate, compared to the other alternatives. The toxicity and volume of the material would not decrease with the addition of a soil cover in the layers underneath the cover but exposure on the surface would be reduced and the mobility of the contaminants by either wind erosion, overland flow, or infiltration would be reduced. This alternative would not be considered effective in the short-term because the soil cover initially destroys the established plant community. This alternative is not a preferred treatment designed to reduce toxicity, mobility and volume. The implementability of this alternative is considered to be moderate but would require extensive borrow areas to bring in soil to cover bedrock areas, and the borrow areas would further destroy plant communities, reducing short-term effectiveness of protecting plants. Because of the resource and fuel use required for implementation, and because of the long-term cover maintenance requirements, this alternative is not considered to be a green option. Total estimated cost is \$13.2 million, which is slightly more than Alternative 3.

Alternative 8 is the use of phytostabilization as a surface soil control. The long-term effectiveness for this alternative is considered to be low. This technology used alone would not be effective at reducing site contaminant levels and/or exposure to the plants onsite, nor would it be effective in the short term because the phytostabilization initially disrupts the established plant community and it also takes a recovery period to re-establish the ecosystem. This alternative is not a preferred treatment designed to reduce toxicity, mobility and volume. This alternative is considered to be implementable, though there may be areas around the site that would present difficulties (due to slopes, percent soil coverage, and soil conditions). Because of the limited resources and fuel use required for implementation, this alternative is considered a green option. Total estimated cost is \$6.6 million.

The recommended remedial alternative is Alternative 5: tilling and monitoring of low pCu soils. Alternative 5 uses the mixing of deeper cleaner soils and natural alkalinity of the Gila Conglomerate occurring in some areas to restore low pCu soils, which the no action and monitoring alternatives do not do. Compared to active remediation alternatives, it involves a lower carbon footprint compared to alternatives relying on excavation and trucking in a soil cover. Tilling would increase pCu but most importantly break up the compacted soils in flat rocky areas that may have become impacted from pCu as well as grazing. Breaking up the soils with tilling or ripping has been demonstrated to improve the plant community, especially grasses, on haul roads without seeding (Appendix A). This alternative addresses the flat rocky soil acreage but does not address bedrock or steeper slopes where tilling cannot occur, but amendments (lime, organic matter) that are part of other alternatives have not been particularly effective in steeper slopes in the amendment study and may not be in bedrock when storm runoff occurs, either. The bedrock and slope soil polygons carried forward for remedial consideration are not recommended for active remediation at the selected PELs (4.4 and 4.11, respectively) because of expected limited effectiveness. Although amendments applied to bedrock and slope soils could be considered for acres less than the lower PELs discussed in section 7.3, which are more certain to provide net environmental benefit, no bedrock or slope areas are lower than the lower range PELs. Implementability of Alternative 5 is better than Alternative 3 (excavation) because less equipment and movement of soil are required. That said, there is no short-term effectiveness due to tilling causing disruption of the plant community, which must re-establish itself as a mature community, which may be an extended period of time; the balance between the harm from plant destruction and benefit from increased pCu to produce the most net benefit is likely at a lower pCu than the selected soil-specific PELs. NMED should evaluate the net environmental benefit when selecting the final PEL (e.g., from Table 7-3 or Table D-5 or Table D-7) to use in the ROD, as the selected PELs likely will do more harm than good. A flat rocky PEL area with pCu <

4.6 is more certain to provide environmental benefit if tilled. Total estimated cost of this preferred alternative at the PELs identified in Section 7.3 is \$5.8 million, which is roughly half the cost of Alternative 3.

## 7.4 Evaluation of Surface Water Alternatives – Surface Water Drainages

### 7.4.1 Detailed Analyses of Surface Water Alternatives – Surface Water Drainages

This section presents a detailed analysis of the surface water drainage alternatives for each of the threshold and balancing criteria described in Section 7.1, above. The drainages evaluated for remediation include the pool habitat areas in Drainages A, non-ephemeral C1, ephemeral C2, D2, and D3, as discussed in Section 4.2.1<sup>17</sup> and indicated on Figure 8-1. Alluvial sections of the drainages lack pool habitat and are not included.

#### 7.4.1.1 Alternative 1: No Action

##### Protection of Human Health and the Environment

The No Action alternative does not provide any additional form of overall protection of human health and the environment. However, current exposure for all but the D3 drainage on average likely is low and protective overall, because only a small percentage of the samples in the drainages had an HQ > 1 (14% of non-ephemeral C-1, 20% of ephemeral C-2, and 25% of D2), except Drainage A, which had 50% of the two locations sampled with HQ > 1. Also, HQ exceedances were often low (2 and at most 3), lower than in D3 with HQs of 5. Evaluation of stock tanks on Drainage A indicates that Drainage A has improved in the more recent years (HQ < 1 in 2010 and 2021) and may be meeting the pre-FS RAC. Recent attempts to re-sample key drainages have failed because all were dry, which also indicates impacts to the environment are small because pool habitat is limited in these areas (see pale yellow areas that have limited habitat in Figure 4-3), particularly in dry years. However, no action does not address the uncertainty of the current conditions, whereas monitoring would. The no action alternative, therefore, does not meet the RAOs.

##### Compliance with ARARs

Alternative 1 for Surface Water Drainages is not compliant with ARARs because it is no action and there are some exceedances of copper based on NMAC §20.6.4.808 – 809 that need to be addressed with monitoring or other actions. Copper is the focus because it is main driver of risk, having significantly higher HQs than cadmium and lead, the other COCs for the drainages.

##### Long-term Effectiveness and Permanence

Alternative 1 for Surface Water Drainages is not effective in the long-term or permanent because it is no action and there are some exceedances of NMAC §20.6.4.808 – 809 for copper. However, natural attenuation may be happening, but without monitoring to ascertain the rates, this alternative is not protective. This alternative has the benefit of preserving the existing upland vegetation and aquatic ecosystem, both of which would take decades to recover following the implementation of an active remedial alternative.

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<sup>17</sup> Subwatershed E pool area is actually a stock tank and addressed in Section 7.5.

### **Reduction of Toxicity, Mobility, or Volume**

Alternative 1 for Surface Water Drainages does not reduce toxicity, mobility or volume because it is the no action alternative and does not provide treatment.

### **Short-term Effectiveness**

Alternative 1 for Surface Water Drainages is not effective in the short-term because it is no action and there are some exceedances of NMAC §20.6.4.808 – 809 and uncertainty about the effect of the exceedances on the biota in the habitat.

### **Implementability**

Alternative 1 is implementable at the site, as no action is required.

### **Cost**

There are no costs associated with this alternative.

## **7.4.1.2 Alternative 2: Monitoring**

### **Protection of Human Health and the Environment**

Alternative 2 does not provide any form of additional overall protection of human health and the environment because it only involves monitoring. However, current exposure for all but the D3 drainage on average is low and protective overall, because only a small percentage of the samples in the drainages had an HQ > 1 (14% of non-ephemeral C-1, 20% of ephemeral C-2, and 25% of D2), except Drainage A, which had 50% of the two locations sampled with HQ > 1. HQs were low (usually 2 and at most 3), lower than in D3, which had HQs of 5. Evaluation of stock tanks on Drainage A indicates that Drainage A has improved in the more recent years (HQ < 1 in 2010 and 2021) and may be meeting the pre-FS RAC. Recent attempts to re-sample key drainages have failed because all were dry, which also indicates impacts to the environment are small because pool habitat is limited in these areas (see pale yellow areas that have limited habitat in Figure 4-3), particularly in dry years. Therefore, monitoring may be protective because it provides more data to evaluate need for remediation and may trigger a remedial alternative over time.

### **Compliance with ARARs**

The Monitoring alternative is not immediately compliant with ARARs or the pre-FS RAC for some point locations, but may be meeting it for drainages on average. Monitoring would allow for adaptability and may trigger remedial activities over time. In this case, the likelihood of compliance with ARARs is dependent on whether pre-FS RAC are met when evaluated across the drainages currently (appears true for all but D3, particularly in areas where pools are limited and gone in dry years) and dependent on the management of upstream sources, suspended sediment and dissolved metals concentrations, as well as overall water quality and attenuation rates.

### **Long-term Effectiveness and Permanence**

The effectiveness of this remedial alternative in the long-term may be limited, as likelihood of reducing concentrations to below aquatic criteria standards depends on natural attenuation, which appears to be occurring or may have already occurred for all but Drainage D3. This alternative has the benefit of preserving the existing upland and riparian vegetation and aquatic ecosystem, both of which would take decades to recover following the implementation of an active remedial alternative, while climate change may be drying up the limited pool habitat.

Additionally, this alternative affords the opportunity to continue data collection and to take additional action if necessary.

### **Reduction of Toxicity, Mobility, or Volume**

Alternative 2 for Surface Water Drainages does not reduce toxicity, mobility or volume because it does not involve any active remedy, similar to the no action alternative.

### **Short-term Effectiveness**

Alternative 2 for Surface Water Drainages is not effective in the short-term because it takes no action other than monitoring, and there are some exceedances of NMAC §20.6.4.808 - 809. However, current conditions may be protective for all but Drainage D3. Monitoring in the short term of all the retained drainages may trigger an action if conditions warrant it.

### **Implementability**

This remedy is implementable at the site with limited additional effort. Long-term monitoring would be required to show improvements in water quality in the surface water drainages.

### **Cost**

The present value cost of this alternative is \$313,000 (see Appendix F).

## **7.4.1.3 Alternative 3: Excavation and Monitoring**

### **Protection of Human Health and the Environment**

The ability of this alternative to meet the overall protection of water quality is moderate to high. The overall protection is based on the removal of historical sediments, but also a reduction of potential upstream contributors in the upland areas feeding the drainages. This alternative would destroy the existing upland and riparian vegetation and aquatic ecosystem initially, and the vegetation could take decades to recover following the implementation of an active remedial alternative. The pools are limited and dry in drier years, particularly in alluvial drainages (pale yellow areas in Figure 4-3), providing limited habitat. The trade-off of remediation benefits and loss of riparian vegetation supports this remedy only for Drainage D3 that has high HQs.

### **Compliance with ARARs**

The likelihood of compliance with the ARARs on an ongoing basis is moderate to high with this alternative. There is the potential for reduced long-term effectiveness of this alternative if upstream contributions continue. Additional excavation may be required at a future time.

### **Long-term Effectiveness and Permanence**

The effectiveness of this remedial alternative in the long-term is moderate to high to reduce concentrations to aquatic life standards. In recent efforts to re-sample, areas of concern were dry so long-term effectiveness at protecting aquatic biota could be low if frequently dry. Once excavation of historical sediments is complete, monitoring will be required to document that water quality criteria have been met. Long-term effectiveness of this alternative may be dependent on addressing upstream contributors (suspended sediments and storm water). This alternative will be most effective in Drainage D3 as the net environmental benefit in other areas with limited exceedances in the long-term may be limited.

### **Reduction of Toxicity, Mobility, or Volume**

This alternative would not meet the statutory preference for treatment. Alternative 3 does not reduce the toxicity, mobility, or volume because no treatment would be implemented.

### **Short-term Effectiveness**

This alternative is considered moderately effective in the short-term. Surface water quality may take some time to improve, as was seen on Tributary 2 in the Lampbright Investigation Unit (Golder 2016). Normal risks associated with nuisance dust, excavation hazards, noise and traffic will impact short-term effectiveness.

### **Implementability**

In specific surface water drainages, this alternative will potentially be difficult to implement compared to other alternatives. Drainage D3 is heavily vegetated over bedrock and tightly confined by the surrounding cliffs, particularly in the upstream portions of the drainage. Because of this, it is unclear if an access road could be built to allow access for this type of remedy. Drainages A, D2, C1 and C2 are similarly vegetated, though less confined with less bedrock in many areas than Drainage D3 and may be implementable but will damage riparian vegetation and wildlife habitat that may take decades to recover. Additional permitting and coordination would also be required compared to other alternatives. For other lower-gradient drainages, the excavation of historical sediments is considered implementable. Some of the drainages may not be accessible to large equipment, due to steep slopes or tight access. However, overall, this alternative is implementable at the lower-gradient sites. Equipment and materials to complete the job are readily available.

### **Cost**

The present value cost of this alternative is \$1 million (see Appendix F). Note this cost does not include building an access road in the drainages to implement this remedy, as the steep and vegetated nature of the drainages requires survey and design information that is currently unavailable.

## **7.4.1.4 Alternative 4: Instream Removal of Suspended Sediments and Monitoring**

### **Protection of Human Health and the Environment**

The ability of this alternative to meet the overall protection of water quality is moderate to high. The overall protection is based on the installation of limestone sedimentation basins within the drainage channels that trap and remove sediments, thereby decreasing the amount and bioavailability of copper in surface water pools. Long-term monitoring is suggested for the implementation of this alternative.

### **Compliance with ARARs**

The likelihood of compliance with the ARARs on an ongoing basis is high, and potentially higher than Alternative 3. There is the potential for reduced long-term effectiveness of this alternative if upstream contributions continue. Excavation should be anticipated within the sedimentation basins periodically over time. The addition of limestone to treat bioavailability of copper increases the potential compliance with ARARs.

### **Long-term Effectiveness and Permanence**

The effectiveness of this remedial alternative in the long-term is moderate to high to reduce concentrations to aquatic life standards. Once excavation of historical sediments in the installation location is complete and sedimentation basins are installed to collect suspended sediment, monitoring will be required to document that water quality criteria have been met. Overall long-term effectiveness will also depend on any additional potential



upstream sources of contaminants and frequency of removing contaminated sediments from the basins. Chemical treatment of the surface with limestone would increase the potential effectiveness of this alternative compared to Alternative 3. Vegetation damage will be limited to sediment basin locations and the constructed road leading to the basins. This alternative will be most effective in Drainage D3 as the net environmental benefit in other areas with limited exceedances in the long-term may be small.

#### **Reduction of Toxicity, Mobility, or Volume**

This alternative addresses toxicity by reducing the bioavailability of copper due to the presence of limestone in the sedimentation basins. It also reduces the mobility of the contaminants by removing them from contact with surface water. The volume of material will be reduced with this alternative and it will also prevent potential contamination of surface water due to the presence of historical sediments.

#### **Short-term Effectiveness**

This alternative is considered moderately effective in the short-term. Surface water quality may take some time to improve, and building a road into these areas to install the sediment basins will destroy vegetation in the short-term. Normal risks will occur associated with nuisance dust, excavation hazards, noise, and traffic.

#### **Implementability**

In specific surface water drainages this alternative will be potentially difficult to implement compared to other alternatives. Drainage D3 is heavily vegetated and tightly confined by the surrounding cliffs, particularly in the upstream portions of the drainage. Because of this, it is unclear if an access road could be built to allow access for this type of remedy. Drainages D2, non-ephemeral C1 and C2 are similarly vegetated, though less confined with less bedrock in many areas than Drainage D3. Additional permitting and coordination would be required compared to other alternatives due to the excavation component. Specifically for the critical habitat transect, excavation may be omitted and only limestone sedimentation basins placed in the surface water drainages. For other drainages, the excavation of historical sediments is considered implementable. Some of the drainages may not be accessible to large equipment, due to steep slopes or tight access. These factors make the implementability of this alternative more challenging than other alternatives. Equipment and materials to complete the job are readily available.

#### **Cost**

The present value cost of this alternative is \$2.8 million (see Appendix F). Note this cost does not include building an access road in the drainages to implement this remedy, as the steep and vegetated nature of the drainages requires survey and design information that is currently unavailable.

### **7.4.1.5 Alternative 5: Limestone Treatment and Monitoring**

#### **Protection of Human Health and the Environment**

The ability of this alternative to meet the overall protection of water quality is moderate to high. The overall protection is based on the installation of limestone in low-lying, low-energy areas within the drainage channels, thereby decreasing the bioavailability of copper in surface water. Long-term monitoring is suggested for the implementation of this alternative.

### **Compliance with ARARs**

The likelihood of compliance with the ARARs on an ongoing basis is moderate to high, as lime would be placed in the stream channels without excavation in areas where velocities are low. The limestone would decrease the bioavailability of copper and increase the potential compliance with ARARs.

### **Long-term Effectiveness and Permanence**

The effectiveness of this remedial alternative in the long-term is moderate to reduce concentrations to aquatic life standards. This technology allows water to pass through the limestone but without the benefit of basins to increase the contact time with surface water, so effectiveness is moderate, not high. Similar to Alternative 4, overall long-term effectiveness will also depend on any additional potential upstream sources of contaminants. This alternative will be most effective in Drainage D3 as the net environmental benefit in other areas with limited exceedances in the long-term may be limited, especially given the limited habitat value in the drainages.

### **Reduction of Toxicity, Mobility, or Volume**

This alternative addresses toxicity by reducing the bioavailability of copper due to the presence of limestone in the drainage channels.

### **Short-term Effectiveness**

This alternative is considered moderately effective in the short-term. Surface water quality may take some time to improve. Normal risks associated with nuisance dust, excavation hazards, noise, and traffic. Vegetation damage is expected to be significant, as an access road would need to be placed along the entire length of the drainage to allow for limestone placement and periodic maintenance.

### **Implementability**

In specific surface water drainages this alternative will be potentially difficult to implement compared to other alternatives. Drainage D3 is heavily vegetated and tightly confined by the surrounding cliffs, particularly in the upstream portions of the drainage. Because of this, it is unclear if an access road could be built to allow access for this type of remedy. Drainages D2, non-ephemeral C1 and C2 are similarly vegetated, though less confined with less bedrock in many areas than Drainage D3. Some of the drainages may not be accessible to large equipment, due to steep slopes or tight access, which would be required to place limestone in as much of the channel bottoms as possible. These factors make the implementability of this alternative more challenging than other alternatives. Equipment and materials to complete the job are readily available.

### **Cost**

The present value cost of this alternative is \$9.6 million (see Appendix F). Note this cost does not include building an access road in the drainages to implement this remedy, as the steep and vegetated nature of the drainages requires survey and design information that is currently unavailable.

## **7.4.1.6 Alternative 6: Sediment Control, Erosion Control, and Monitoring**

### **Protection of Human Health and the Environment**

The ability of this alternative to meet the overall protection of water quality is low to moderate. The overall protection of this alternative is based on the control of suspended sediments and potentially dissolved concentrations in storm water. This alternative does not address historical sediments. The addition of limestone as part of the BMPs to treat dissolved/bioavailable copper would increase the potential compliance with ARARs.

### **Compliance with ARARs**

The ability of this alternative to comply with this criterion is low to moderate.

### **Long-term Effectiveness and Permanence**

The long-term effectiveness of this alternative is low to moderate. A significant amount of maintenance and monitoring will be required to implement a large enough sediment and erosion control program across all of the drainages. Potential excavation and maintenance of the BMPs would be required on a regular basis to keep this alternative effective. This alternative would damage vegetation and wildlife habitat which would take decades to recover. This alternative will be most effective in Drainage D3 as the net environmental benefit in other areas with limited exceedances in the long-term may be limited, especially given the limited habitat value in the drainages.

### **Reduction of Toxicity, Mobility, or Volume**

This alternative (without limestone treatment) would not meet the statutory preference for treatment. Alternative 6 does not reduce the toxicity, mobility, or volume because no treatment would be implemented. If limestone or other material is used, toxicity and volume of the constituents may be reduced. Mobility of the constituents may be reduced.

### **Short-term Effectiveness**

This alternative would be low to moderately effective in the short-term for aquatic habitat but may damage upland habitat. This alternative does not address the presence of historical sediments in the drainages. This alternative would take time for water quality to improve.

### **Implementability**

In specific surface water drainages, this alternative will be potentially difficult to implement compared to other alternatives. Drainage D3 is heavily vegetated and tightly confined by the surrounding cliffs, particularly in the upstream portions of the drainage. Because of this, it is unclear if an access road could be built to allow access for this type of remedy. Drainages D2, non-ephemeral C1 and C2 are similarly vegetated, though less confined with less bedrock in many areas than Drainage D3. Some of the drainages may not be accessible to large equipment, due to steep slopes or tight access. Equipment and materials to complete the job are readily available.

### **Cost**

The present value cost of this alternative is \$4.2 million (see Appendix F). Note this cost does not include building an access road in the drainages to implement this remedy, as the steep and vegetated nature of the drainages requires survey and design information that is currently unavailable.

## **7.4.2 Comparative Analyses of Surface Water Alternatives – Surface Water Drainages**

As presented in Section 4, four drainages were determined to have concentrations of potential concern to aquatic life uses, including Drainage A, Drainages C1 (non-ephemeral only) and C2 (ephemeral), Drainage D2, and Drainage D3. It is therefore appropriate to evaluate remedial alternatives for these five drainages.

### **Drainage A**

The surface water in Drainage A is ephemeral and it is thus appropriate to use the WER-based Acute HQ when evaluating surface water quality. As included on Table 4-5 (watershed-specific WER-adjusted acute HQ), surface water results for copper ranged between an HQ of 1 and 3 in 2010, which is relatively low. Further, HQs in samples collected from stock tanks in Drainage A have been low. Stock tank 60 has historically been less than or equal to 3, HQs in samples collected from stock tanks 15, 25, and 26 have historically been less than 2, and HQs in samples collected from stock tank 16 have been less than 1 since 2007 (Table 4-7). Stock tanks 15, 26, and 60 were removed as part of rangeland improvements completed between 2013 and 2022 (see Section 4.1.2), so current data for these tanks are unavailable. However, a water quality sample was collected from stock tank 16 in 2021 and the WER-based Acute HQ for copper was less than 1. As a result of the relatively low HQs historically measured in this drainage, and because the latest sample indicates the copper HQ is less than 1, the monitoring remedy (Alternative 2) included on Table 7-4 is the most appropriate remedy for this drainage that may be naturally attenuating. Alternative 2 is also the greenest remediation alternative for Drainage A behind no action.

### **Drainages C1 (Non-ephemeral)**

The surface water in the area of potential concern in Drainage C1 is non-ephemeral, and it is thus appropriate to use the watershed-specific WER-based Chronic HQ when evaluating surface water quality in this drainage. As included on Table 4-6, surface water results for copper have ranged between an HQ of 1 and 2 since 2010. Because the watershed-specific WER adjusted copper chronic HQs are around 2 or below in this drainage, and because more aggressive remedies may involve damaging existing vegetation that is currently thriving in these drainage systems, the monitoring remedy (Alternative 2) included on Table 7-4 is the most appropriate remedy for these drainages. Alternative 2 is also the greenest remediation alternative for Drainage C1 behind no action.

### **Drainage C2 (Ephemeral)**

The surface waters in areas of potential concern in Drainage C2 are ephemeral in downstream sections (Figure 4-3) and thus it is appropriate to use the watershed-based WER-based Acute HQ when evaluating surface water quality in this drainage. The ephemeral portion has not been sampled but is included to be conservative. Because the copper HQs are around 2 or below in this drainage, and because more aggressive remedies may involve damaging existing vegetation that is currently thriving in this drainage systems, the monitoring remedy (Alternative 2) included on Table 7-4 is the most appropriate remedy. Alternative 2 is also the greenest remediation alternative for Drainage C2 behind no action.

### **Drainage D2**

The surface water in Drainage D2 is ephemeral, and it is thus appropriate to use the WER-based Acute HQ when evaluating surface water quality. As included on Table 4-5 (watershed-specific WER-adjusted acute HQ), surface water results for copper ranged between an HQ of 0.5 and 2 between 2010 and 2013. An HQ of 2 was observed in 2013, which is anomalously high when compared to the HQs measured in 2010 and 2011. More data are needed in this drainage to determine if HQs return to their pre-2013 levels. The monitoring remedy (Alternative 2) included on Table 7-4 is therefore the most appropriate remedy for this drainage, as it prevents prematurely implementing an active remedy based on one data point and allows for the collection of additional data and the flexibility to act in the future, if needed. Alternative 2 is also the greenest remediation alternative for Drainage A behind no action.

### Drainage D3

The surface water in Drainages D3 is ephemeral and it is thus appropriate to use the WER-based Acute HQ when evaluating surface water quality in this drainage. As described in Section 4, one surface water drainage location in Drainage D3 exceeds the WER-based Acute HQ for copper (see Section 4.2.1). The HQ for this location is 5 (Table 4-5). The water quality results in this drainage support one of the active remedies included in Table 7-4. However, Drainage D3 is steep, heavily vegetated, and little access is available to implement an active remedy. In fact, access constraints have limited additional sample collection higher up in the drainage. A road would need to be constructed to access the upgradient portions of the drainage to implement the active remedies, and the steep walls confining the drainage and heavy vegetation make this technically impractical. As a result, the monitoring remedy (Alternative 2) included on Table 7-4 is the most appropriate remedy for this drainage. Alternative 2 is also the greenest remediation alternative for Drainage D3 behind no action.

The total estimated cost for implementing a monitoring remedy for Drainages A, C1, C2, and D3 is \$313K.

## 7.5 Evaluation of Surface Water Alternatives – Stock Tanks

### 7.5.1 Detailed Analyses of Surface Water Alternatives – Stock Tanks

This section presents a detailed analysis of the surface water stock tank remediation alternatives for each of the threshold and balancing criteria described in Section 7.1, above. The stock tank evaluated for remediation includes stock tank 06, as discussed in Section 4.2.2 and indicated on Figure 8-1.

#### 7.5.1.1 Alternative 1: No Action

##### Protection of Human Health and the Environment

The No Action alternative does not provide any additional form of overall protection of human health and the environment and does not currently meet the RAOs. However, current exposure at stock tank 06 likely is low and protective overall, because HQs have oscillated between 1 and 2 since 2007 and improvements were made in 2023 to the approximately 230-acre drainage basin feeding stock tank 06 to provide for stable stormwater conveyance to the tank inlet (see Section 4.1.2). Because of these improvements, it is expected that water quality will continue to improve in stock tank 06 with time and may eventually meet the pre-FS RAC and RAOs.

##### Compliance with ARARs

Alternative 1 for Surface Water Drainages is not compliant with ARARs because it is no action and there are some exceedances of copper based on NMAC §20.6.4.808 – 809 that need to be addressed with monitoring or other actions.

##### Long-term Effectiveness and Permanence

Alternative 1 for Stock Tanks is not effective in the long-term or permanent because it is no action and there are some exceedances of NMAC §20.6.4.808 – 809 for copper. However, natural attenuation may be happening, particularly following the 2023 rangeland improvements, but without monitoring to ascertain the rates, this alternative is not protective.

### **Reduction of Toxicity, Mobility, or Volume**

Alternative 1 does not reduce toxicity, mobility or volume because it is the no action alternative and does not provide treatment.

### **Short-term Effectiveness**

Alternative 1 is not effective in the short-term because it is no action and there are some exceedances of NMAC §20.6.4.808 - 809.

### **Implementability**

Alternative 1 is implementable at the site, as no action is required.

### **Cost**

There are no costs associated with this alternative.

## **7.5.1.2 Alternative 2: Monitoring**

### **Protection of Human Health and the Environment**

Alternative 2 does not provide any form of overall protection of human health and the environment because it only involves monitoring. However, current exposure at stock tank 06 likely is low and protective overall, because HQs have oscillated between 1 and 2 since 2007 and improvements were made in 2023 to the approximately 230-acre drainage basin feeding stock tank 06 to provide for stable stormwater conveyance to the tank inlet (see Section 4.1.2). Because of these improvements, it is expected that water quality will continue to improve in stock tank 06 with time.

### **Compliance with ARARs**

The Monitoring alternative is not immediately compliant with ARARs and there are some exceedances of copper based on NMAC §20.6.4.808 – 809. However, monitoring allows for adaptability and may trigger remedial activities over time. In this case, the likelihood of compliance with ARARs is dependent on the management of upstream sources, suspended sediment and dissolve metals concentrations, as well as overall water quality.

### **Long-term Effectiveness and Permanence**

This remedial alternative may be effective in the long-term for reducing concentrations to below aquatic life standards with natural attenuation after the rangeland improvements. This monitoring alternative may be effective enough to address long-term risks because upstream sources of contaminants are concurrently addressed with the rangeland improvements, which were conducted in 2023. Additionally, this alternative affords the opportunity to continue data collection and to take additional action if necessary.

### **Reduction of Toxicity, Mobility, or Volume**

This alternative does not reduce toxicity, mobility or volume because it is the no action alternative and does not provide treatment.

### **Short-term Effectiveness**

This alternative may be effective in the short term, despite some exceedances of NMAC §20.6.4.808 – 809. The rangeland improvements conducted in 2023 may have already improved water quality in the short term, but the monitoring program is needed to establish if this occurred.

### **Implementability**

This remedy is implementable at the site with limited additional effort.

### **Cost**

The present value cost of this alternative is \$258,000 (see Appendix F).

## **7.5.1.3 Alternative 3: Excavation and Monitoring**

### **Protection of Human Health and the Environment**

Removal of historical sediments from stock tanks would address a major contributor to aquatic risk; however, there is still the potential for long-term risk to aquatic receptors due to upstream contribution from either suspended sediments or dissolved metals concentrations in surface water. Improvements were made in 2023 to the approximately 230-acre drainage basin feeding stock tank 06 to provide for stable stormwater conveyance to the tank inlet (see Section 4.1.2). Because of these improvements, it is expected that water quality will continue to improve in stock tank 06 with time with or without active remediation.

### **Compliance with ARARs**

The likelihood of compliance with the ARARs on an ongoing basis is moderate to high with this alternative. There is the potential for reduced long-term effectiveness of this alternative if upstream contributions continue. Additional excavation may be required at a future time. If they do not continue, natural attenuation is expected.

### **Long-term Effectiveness and Permanence**

The effectiveness of this remedial alternative to reduce concentrations to aquatic life standards in the long-term is moderate to high. Once excavation of historical sediments is complete, monitoring will be required to document that water quality criteria have been met. Long-term effectiveness of this alternative may be dependent on addressing upstream contributors (suspended sediments and storm water). This process may already be underway following the rangeland improvements conducted in 2023.

### **Reduction of Toxicity, Mobility, or Volume**

This alternative would not meet the statutory preference for treatment. Alternative 6 does not reduce the toxicity, mobility, or volume because no treatment would be implemented.

### **Short-term Effectiveness**

This alternative is considered effective in the short-term, and water quality may already be improving following the 2023 rangeland improvements. Normal risks associated with nuisance dust, excavation hazards, noise, and traffic. Coordination may need to be made to provide alternative water sources to tanks owners during construction.

### **Implementability**

Based on the location of stock tank 06 and its accessibility, this alternative is implementable at the site. Equipment and materials to complete the job are readily available.

### **Cost**

The present value cost of this alternative is \$325,000 (see Appendix F).

## 7.5.2 Comparative Analyses of Surface Water Alternatives – Stock Tanks

As presented in Section 4, stock tank 06 was determined to have concentrations of potential concern to aquatic life uses, and it is thus appropriate to evaluate remedial alternatives for this tank. The WER-adjusted acute HQ for copper has been less than 2 since 2007 (Table 4-7). Additionally, the approximately 230-acre drainage basin feeding stock tank 06 has been improved to provide for stable stormwater conveyance to the tank inlet (see Section 4.1.2). Because of these improvements, it is expected that water quality will continue to improve in stock tank 06 with time. As a result, the monitoring remedy (Alternative 2) included on Table 7-5 is the most appropriate remedy for this stock tank, as it allows for adaptability during the monitoring process and the potential to act at a later time if HQs do not continue to decline. Alternative 2 is also the greenest remediation alternative of those considered for stock tank 06 behind no action. Total estimated cost is \$258,000.



## 8. Final Acres for Remedial Actions or Monitoring

Figure 8-1 identifies the acres for remedial actions or monitoring recommended in Section 7 when flat rocky soil PEL is set at 4.98. Figure 8-2 identifies the acres for remedial actions or monitoring recommended in Section 7 when flat rocky soil PEL is set at 4.6, which may provide greater ecological benefit than the 4.98 PEL. As discussed in Section 7.3, none of the active remedies achieve short term effectiveness because they destroy plants in order to protect them and it takes decades for the ecosystem to re-establish. Given the uncertainties in the derivation of PELs and the impact of grazing on the reference and site locations, choosing a lower PEL will actually achieve greater ecological benefit. Moreover, the capital expenditure is nearly triple the cost for flat rocky at a PEL of 4.98 with less certain ecological benefit compared to a PEL of 4.6. None of the acres recommended for remedial actions or monitoring overlap the recent TP-7 windblown areas (Figure 8-1, 8-2) and thus are included in this FS.

In summary, for soil copper, 140 acres are recommended for monitoring. For soil pCu, 313 acres in flat rocky soils are recommended for tilling and monitoring when pCu PEL for flat rocky is 4.98, and 113 acres in flat rocky soils are recommended for tilling and monitoring when pCu PEL for flat rocky is 4.6. For surface water drainages, sections with pools that have aquatic habitat in Drainage A, C1 (non-ephemeral section), C2 (ephemeral section), D2, and D3 are recommended for monitoring. For stock tanks, stock tank 6 in the western drainage in Drainage E, which received rangeland improvements in 2023, is recommended for monitoring.

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# Tables

**TABLE 2-1  
CHEMICAL-SPECIFIC POTENTIALLY APPLICABLE STANDARDS FOR THE STSIU**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Regulatory Program/Authority	Citation	Medium of Potential Interest	Notes
Safe Drinking Water Act (SDWA), Federal	40 CFR 141 Subpart F	Groundwater, Surface Water	Establishes primary drinking standards for public water systems.
SDWA, Federal	40 CFR 143, Subpart B	Groundwater, Surface Water	Establishes secondary non-enforceable health goals for public water systems at levels resulting in no known or anticipated adverse health effects.
Clean Air Act, Federal	40 CFR 50	Air	Establishes primary and secondary ambient air quality standards.
Clean Air Act, Federal	40 CFR 60	Air	Establishes (referenced by NMED AQCR 652) performance standards for new sources based on the specific source categories defined in the regulation.
Air Quality Control Act, State	20.2.3 NMAC	Air	Establishes ambient air quality standards.
Air Quality Control Act, State	20.2.78 NMAC	Air	Defines emissions standards for hazardous air pollutants.
New Mexico Water Quality Act	20.6.2.7.VV NMAC	Groundwater, Surface Water	Definition of a toxic pollutant.
New Mexico Water Quality Act	20.6.2.3101 NMAC	Groundwater	Designates groundwater with total dissolved solids <= 10,000 mg/L as potential source of drinking water.
New Mexico Water Quality Act	20.6.4 NMAC	Surface Water	Provides water quality standards for human contact of surface waters. Defines water quality standards for livestock watering. This statute includes an anti-degradation policy, general water quality standards, primary contact standards, and wildlife standards.
New Mexico Water Quality Act	20.6.2.3103(A) NMAC	Groundwater	Establish human health standards for groundwater quality.
New Mexico Water Quality Act	20.6.2.3103(B) NMAC	Groundwater	Establishes additional standards for domestic water supplies.
New Mexico Water Quality Act	20.6.2.3103(C) NMAC	Groundwater	Establishes groundwater quality standards for irrigation use.
Resource Conservation and Recovery Act (RCRA), Federal	40 CFR 261.24	Soil	Regulates the determination of hazardous wastes by defining the maximum concentrations of listed contaminants as measured using the Toxicity Characteristic Leaching Procedure (TCLP).
CERCLA	40 CFR 300 Title 1, Section 101, 111	All Media	References the National Oil and Hazardous Substances Contingency Plan. Establishes funding and provisions for cleanup at hazardous waste sites.



**TABLE 2-2  
ACTION-SPECIFIC POTENTIALLY APPLICABLE STANDARDS FOR THE STSIU**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Regulatory Program/Authority	Citation	Medium of Potential Interest	Description
CERCLA	40 CFR 300 Title 1, Section 101, 111	All Media	References the National Oil and Hazardous Substances Contingency Plan. Establishes funding and provisions for cleanup at hazardous waste sites.
SARA	42 USC 9601	All Media	Establishes clean-up standards and response actions, including the ARAR process (i.e. Applicable Standards).
Clean Water Act - National Pollution Discharge Elimination System (NPDES)	40 CFR 122 CFR 125	Surface Water	Requires permits for discharging pollutants from any point source into waters, lists hazardous substances and water-quality parameters, and defines the criteria and standards for issuances of permits, determining compliance, and granting variances. Establishes Best Management Practices (BMPs) to prevent releases of toxic constituents to surface waters.
Clean Water Act	40 CFR 230 CFR 231	Surface Water	Requires permits for discharging dredged or fill materials into the navigable waters, including wetlands or floodplains. Permits (Sec 404) are issued if the state has authorization, otherwise, Nation Wide Permits (NWP) can be issued by the US Army, Corps of Engineers. Applies to all stream modifications, including underground and surface mining activities.
Rivers and Harbors Act of 1899	33 CFR 320 CFR 330	Surface Water	Regulates disposal/discharge of dredged or fill materials into US waters, including intermittent streams.
RCRA	40 CFR 241	Soil	Specifies performance requirements for land disposal of wastes.
RCRA	40 CFR 261	Soil	Defines criteria for identifying and classifying hazardous wastes.
RCRA	40 CFR 262	Soil	Establishes standards for generators of hazardous wastes, including requirements for waste shipment packaging, labeling, and manifests. Requirements may be applicable if remediation activities are performed at the S/TSIU and waste generated are hazardous.
RCRA	40 CFR 263	Soil	Establishes standards for transporters of hazardous wastes.
RCRA	40 CFR 264	Soil	Establishes standards for owner and operators of facilities for the treatment, storage, and disposal of hazardous wastes.
RCRA	40 CFR 268	All Media	Establishes treatment standards for hazardous constituents, identifies wastes that are restricted from land disposal and defines the limited circumstances under which they may be land disposed.
Department of Transportation (DOT) Regulations	49 CFR 173, 178, 179	Soil	Establishes requirements for packaging and shipment of hazardous waste.
CERCLA Off-Site Response Policy	OSWER 9634.11	All Media	Defines criteria for qualifying an off-site hazardous waste disposal facility.
Clean Air Act	42 USC Sections 7401 et. seq.	Air	Requires formulation of air quality standards and source performance standards.

**TABLE 2-2  
ACTION-SPECIFIC POTENTIALLY APPLICABLE STANDARDS FOR THE STSIU**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Regulatory Program/Authority	Citation	Medium of Potential Interest	Description
New Mexico Hazardous Waste Act (NMHWA) NMED Hazardous Waste Bureau (HWB)	NMSA 1978, Sections 74-4-1 through 74-4-14	Hazardous Waste	Regulates treatment, storage, and disposal of hazardous waste to ensure maintenance to the quality of the state's environment.
NMHWA, NMED HWB	20.4.1.200 NMAC	Hazardous Waste	Defines criteria for identifying and classifying hazardous waste.
NMHWA, NMED HWB	20.4.1.300 NMAC	Hazardous Waste	Defines standards applicable to generators of hazardous wastes for packaging, labeling, and manifesting waste for transport.
NMHWA, NMED HWB	20.4.2.400 NMAC	Hazardous Waste	Defines standards applicable to the transportation of hazardous waste.
NMHWA, NMED HWB	20.4.1.900 NMAC	Hazardous Waste	Identifies hazardous wastes which are restricted from land disposal.
New Mexico Solid Waste Management Regulations	20.9.1 NMAC	Solid Waste	Regulates the permitting, design, location, and operation of solid waste disposal facilities.
New Mexico Water Quality Act (NMWQA)	NMSA 1978, Sections 74-6-1 through 74-6-17	Groundwater, Surface Water	Bans non-permitted discharge of any water contaminant.
NMWQA	20 NMAC 6.2, Section 1-201	Groundwater, Surface Water	Requires that NMED be notified of any discharge which could affect surface water or groundwater quality.
NMWQA	20 NMAC 6.2, Section 3-104	Groundwater	Discharge plan may be required for any discharge affecting groundwater quality.
NMWQA	20 NMAC 6.2, Section 4-103	Groundwater, Surface Water	Abatement standards and requirements for the vadose zone, groundwater and surface water.
Occupational Safety and Health Act (OSHA)	29 CFR 1910, 1926, 1954	All Media	These standards establish safety requirements for hazardous waste operations and sets exposure limits of chemicals.
RCRA	42 USC Sections 8901 et. seq.	Hazardous Waste	Regulates treatment, storage, and disposal of hazardous waste and encourages resource conservation and recycling.

**TABLE 2-3  
LOCATION-SPECIFIC POTENTIALLY APPLICABLE STANDARDS FOR THE STSIU**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Regulatory Program/Authority	Citation	Medium of Potential Interest	Notes
National Historic Preservation Act	36 CFR 63	Historic, Archaeological	Establishes procedures for determining a property's eligibility for inclusion in the National Register of Historic Places.
National Historic Preservation Act	36 CFR 800	Historic, Archaeological	Requires that federal agencies consider the effects of actions on historic properties and archaeological resources.
National Historic Preservation Act of 1979	36 CFR 296 7 43 CFR	Historic, Archaeological	Establishes procedures to be followed by federal land managers in providing protection for archaeological resources.
Standards and Guidelines for Archaeology and Historic Preservation	48 CFR 44716	Archaeological	Provides guidelines for conducting archaeological surveys.
American Indian Religious Freedom Act of 1978	42 USC 1996	Cultural	Requires consultation with local tribes if a project could effect ceremonial, religious, or burial sites.
American Indian Graves Freedom and Reparation Act	25 USC 3001 through 25 USC 3013	Cultural	Requires that project activities cease if Native American graves are discovered.
Migratory Bird Treaty Act	50 CFR 10, 21	Wildlife	Prohibits pursuit, hunting, taking, capture, possession, or killing of all migratory birds or their nests or eggs.
Bald and Golden Eagle Protection Act	50 CFR 10, 22	Wildlife	Prohibits taking or killing of bald and golden eagles.
Endangered Species Act of 1973	40 CFR 17 and 50 CFR 402	Plant, Wildlife	Requires that actions do not jeopardize endangered species or adversely modify their critical habitat, and establishes the process for consulting with the US Fish and Wildlife Service.
Fish and Wildlife Coordination Act	40 CFR 6.302g	Surface Water	Requires that federal agencies be consulted prior to modifying any stream so that wildlife will be protected.
Endangered Species Act	16 USC 1531	Wildlife	Protects endangered species and restricts activities within their habitat.
Resource Conservation and Recovery Act (RCRA)	40 CFR 241.202	All Media	Establishes standards for siting RCRA solid-waste disposal facilities.
Fish and Wildlife Coordination Act	40 CFR 6.302	Rivers	Protects wildlife habitats and prevents the modification of streams or rivers that effect fish or wildlife.

**TABLE 2-3  
LOCATION-SPECIFIC POTENTIALLY APPLICABLE STANDARDS FOR THE STSIU**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Regulatory Program/Authority	Citation	Medium of Potential Interest	Notes
Executive Order, 11990	40 CFR 6 Appendix A	Wetlands	Protects wetlands and regulates activities conducted in a wetland area in order to minimize potential destruction, loss or degradation of the wetlands.
Clean Water Act	40 CFR 230 33 CFR 320-330	Wetlands	Prohibits filling of wetlands and prohibits the discharge dredged or filled material to a wetland without a permit.
Executive Order, 11988	40 CFR 6 Appendix A	Floodplains	Restricts the types of activities that can be conducted within a floodplain to minimize harm and preserve natural values.
New Mexico Cultural Properties Act	NMSA 18.6	Historic, Archaeological	Requires identification of cultural resources, assessment of potential effects, and consultation with the State Historic Preservation Officer.
New Mexico Wildlife Conservation Act, and New Mexico Endangered Plant Act	NMSA 17-2-27 through NMSA 17-2-46	Plant, Wildlife	Establishes the State's authority to conduct an investigation for the purpose of identifying endangered and threatened species and developing (if necessary) an appropriate management plan for ensuring the protection of such species.
New Mexico Prehistoric and Historic Sites and Preservation Act	NMSA 1978, Sections 18-8- 1 through 18-8-8	Historic, Archaeological	Requires identification of historic resources, assessment of potential impacts, and consultation with State Historic Preservation Office.
National Environmental Policy Act	42 USC Section 4331 et. seq.	Ecosystems	Policy to encourage harmony between humans and the environment to minimize environmental damages and support health and welfare. The Act encourages coordination and cooperation between government agencies in planning and conduction of any action that will affect the government.
National Environmental Policy Act	40 CFR Part 6	Ecosystems	Procedures requiring integration of all applicable federal laws and executive orders into the environment review process mandated under the Act.

TABLE 3-1  
SUMMARY OF SURFACE SOIL ARSENIC AND IRON RESULTS

FREEMONT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELT/TAILING SOILS IU FEASIBILITY STUDY

Sample Location: Sample Depth (inch): Date Collected:	S1 0-1 11/06/04	S2 0-1 11/06/04	S3 0-1 11/06/04	S4 0-1 11/06/04	S7 0-1 10/28/04	S8 0-1 11/06/04	S9 0-1 10/29/04	S10 0-1 11/06/04	S11 0-1 10/27/04	S12 0-1 10/28/04	S13 0-1 10/28/04	S14 0-1 10/29/04	S15 0-1 10/29/04	S16 0-1 11/06/04	S17 0-1 10/27/04	S18 0-1 10/27/04	S19 0-1 10/30/04	S20 0-1 10/27/04	
Metals																			
Arsenic	3.70 J	1.80 J	3.50 J	1.60 J		4.30 J	5.00 J	3.00 J	3.40 J	2.50 J	6.40 J	5.60 J	6.80 J	4.50 J	3.60 J	3.60 J	5.00 J	3.50 J	
Iron	13,700	9,630	15,300	45,500		21,200	16,800	13,900	17,100	17,000	21,000	19,100	24,000	17,100	20,100	16,500	19,300	18,000	

Location ID: Sample Depth (inches): Date Collected:	S21 0-1 10/27/04	S22 0-1 10/30/04	S23 0-1 10/28/04	S24 0-1 10/28/04	S28 0-1 10/30/04	S29 0-1 10/31/04	S30 0-1 10/28/04	S31 0-1 10/30/04	S32 0-1 10/29/04	S33 0-1 10/29/04	S34 0-1 10/29/04	S35 0-1 11/07/04	S36 0-1 11/10/04	S37 0-1 11/07/04	S38 0-1 11/07/04	S39 0-1 11/11/04	S40 0-1 11/11/04
Metals																	
Arsenic	3.30 J	6.30 J	2.60 J	2.80 J	2.90 J	1.40 J	1.10 J	2.50 J	2.70 J	2.60 J	2.80 J	4.80 J	4.70 J	5.10 J	5.90 J	6.50 J	9.40 J
Iron	20,500	19,800	17,200	17,100	16,000	18,100	14,500	13,900	15,300	14,500	17,900	18,200	21,100	19,900	19,000	18,900	24,500

Location ID: Sample Depth (inches): Date Collected:	S41 0-1 11/07/04	S42 0-1 11/08/04	S43 0-1 11/11/04	S44 0-1 11/11/04	S48 0-1 11/11/04	S49 0-1 11/07/04	S50 0-1 11/08/04	S51 0-1 11/09/04	S52 0-1 11/11/04	S53 0-1 11/07/04	S54 0-1 11/08/04	S55 0-1 11/09/04	S56 0-1 11/09/04	S57 0-1 11/08/04	S58 0-1 11/08/04	S59 0-1 10/30/04	S60 0-1 10/30/04
Metals																	
Arsenic	6.20 UJ	6.30 J	3.70 J	0.980 J	7.30 J	10.3 J	1.80 J	5.20 J	3.70 J	15.1 J	7.30 J	4.50 J	3.10 J	3.50 J	3.90 J	18.9 J	21.8 J
Iron	33,600	25,100	12,800	7,170	21,000	41,000	13,900	25,500	14,300	45,700	34,800	24,200	18,600	34,600	32,700	62,400	58,600

Location ID: Sample Depth (inches): Date Collected:	S61 0-1 10/31/04	S62 0-1 10/31/04	S63 0-1 10/31/04	SS97 0-1 07/24/06	SS101 0-1 07/12/06	SS102 0-1 07/12/06	SS103 0-1 07/24/06	SS104 0-1 07/22/06	SS105 0-1 07/20/06	SS106 0-1 07/22/06	SS107 0-1 07/19/06	SS108 0-1 07/17/06	SS109 0-1 07/20/06	SS110 0-1 07/22/06	SS111 0-1 07/19/06	SS112 0-1 07/19/06	SS113 0-1 07/17/06
Metals																	
Arsenic	25.4 J	16.0 J	9.40 J	3.90	2.20	2.20	2.40	2.10	2.00	2.10	0.980	1.40	1.90	2.90	2.30	2.20	1.80
Iron	34,600	35,500	24,600	29,400	55,000	34,200	12,300	10,300	30,200	10,500	7,360 J	32,300	23,600	24,700	12,100 J	14,500 J	14,200

Location ID: Sample Depth (inches): Date Collected:	SS114 0-1 07/17/06	SS115 0-1 07/23/06	SS116 0-1 07/22/06	SS117 0-1 07/20/06	SS121 0-1 07/11/06	SS122 0-1 07/18/06	SS123 0-1 07/13/06	SS124S 0-1 07/19/06	SS125S 0-1 07/19/06	SS126 0-1 07/22/06	SS127 0-1 07/20/06	SS128 0-1 07/20/06	SS129S 0-1 07/19/06	SS130 0-1 07/21/06	SS131S 0-1 07/18/06	SS132 0-1 07/24/06	SS133 0-1 07/22/06
Metals																	
Arsenic	1.80	3.10	3.60	5.40	3.20	0.760	2.00	2.70	1.30	2.00	1.90	1.80	2.10	1.50	1.80	2.90	2.50
Iron	31,800	24,200	34,400	31,100 J	21,400 J	10,600 J	14,300	26,700 J	12,300 J	19,000	27,700 J	31,800 J	26,000 J	21,200	17,500 J	26,200	23,300

Location ID: Sample Depth (inches): Date Collected:	SS134 0-1 07/21/06	SS135 0-1 07/21/06	SS136 0-1 07/22/06	SS137 0-1 07/21/06	SS141 0-1 07/23/06	SS142 0-1 07/21/06	SS143 0-1 07/24/06	SS144 0-1 07/22/06	SS145 0-1 07/23/06	SS146 0-1 07/23/06	SS147 0-1 07/22/06	SS148 0-1 07/18/06	SS149 0-1 07/24/06	SS150 0-1 07/24/06	SS151 0-1 07/15/06	SS152 0-1 07/15/06	SS153 0-1 07/15/06
Metals																	
Arsenic	2.20	2.10	2.30	2.30	1.90	3.20	2.20	4.20	3.40	2.70	2.00	3.80	2.70	2.60	2.30	1.20	2.50
Iron	47,200	36,400	25,200	62,000	55,600	53,500	29,200	30,800	141,000	64,300	24,700	123,000 J	46,500	24,600	19,700	24,100	24,200

TABLE 3-1  
SUMMARY OF SURFACE SOIL ARSENIC AND IRON RESULTS

FREPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMEL TAILING SOILS IU FEASIBILITY STUDY

Location ID: Sample Depth(Inches): Date Collected:	Pre-FS RAC	Units	SS154 0 - 1 07/16/06	SS155 0 - 1 07/16/06	SS156 0 - 1 07/13/06	SS157 0 - 1 07/15/06	SS160 0 - 1 07/16/06	SS161D 0 - 1 07/11/06	SS162D 0 - 1 07/11/06	SS163D 0 - 1 07/11/06	SS164D 0 - 1 07/14/06	SS165D 0 - 1 07/14/06
Metals	27	mg/kg	0.0400 U	1.80	2.10	1.70	0.95	2.2	1.9	0.92	1.9	1.8
Arsenic	100,000	mg/kg	24,400	23,300	28,600	20,100	18,400	18,800	16,200	12,700	9,460	10,600
Iron												

The following data were collected for the RI but were not used in the FS because the sampling locations overlapped interim action locations:

Location ID: Sample Depth(Inches): Date Collected:	Pre-FS RAC	Units	S64 0 - 1 11/05/04	S65 0 - 1 11/05/04	S66 0 - 1 11/05/04	S67 0 - 1 11/05/04	S70 0 - 1 11/05/04	S71 0 - 1 11/05/04	S72 0 - 1 11/04/04	S73 0 - 1 11/05/04	S74 0 - 1 11/01/04	S75 0 - 1 11/01/04	S76 0 - 1 11/01/04	S77 0 - 1 11/01/04	S78 0 - 1 11/04/04	S79 0 - 1 11/04/04	S80 0 - 1 11/10/04	S81 0 - 1 11/10/04	S82 0 - 1 11/10/04	S83 0 - 1 11/12/04
Metals	27	mg/kg	4.1 J	2.3	1.2	2.4 J	2.2	3.3 J	2.3	2.1	1.6	1.7	2.6	1.3	1.5	1.4	1.1 J	2.5 J	2.2 J	1.9
Arsenic	100,000	mg/kg	25,700	16,200	20,400	27,400	19,700	15,000	9,670	12,900	9,780	15,800	11,700	9,710	10,900	11,100	17,200	20,300	12,600	11,400
Iron																				

Location ID: Sample Depth(Inches): Date Collected:	Pre-FS RAC	Units	S84 0 - 1 11/10/04	S85 0 - 1 11/10/04	S86 0 - 1 11/17/04	S87 0 - 1 11/12/04	S90 0 - 1 11/17/04	S91 0 - 1 11/17/04	S92 0 - 1 11/17/04	S93 0 - 1 11/17/04	S94 0 - 1 11/12/04	S95 0 - 1 11/12/04	S96 0 - 1 11/17/04
Metals	27	mg/kg	0.45 J	1.4 J	2.4	1.4	2.6	1.8	2.2	2.5	1.1	2.6	2.1
Arsenic	100,000	mg/kg	14,600	18,500	16,200	9,290	14,800	23,646	36,300	15,300	18,500	8,650	15,800
Iron													

Notes:  
Shaded values exceed the Pre-FS RAC  
mg/kg = milligram per kilogram







TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0-6"

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1" soil <sup>1</sup>	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	Copper Concentration <sup>1</sup> To Compare to HH Pre-Remedial Action (250 µm sieve)	For Bird Pre-FS RAC Analysis (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
R-12	U05-3030	ACC Background Report, Chino 1995	1995	0-1"	not sieved		594224.661	2616792.668	32.6330850	-108.1768700	73	66	Y	Removed	Outside vegetation alliance polygon extent
R-14	U06-3037	ACC Background Report, Chino 1995	1995	0-1"	not sieved		592949.320	2624423.600	32.6396480	-108.1833840	43	39	Y	Removed	Outside vegetation alliance polygon extent
Reference Plot #1 (West)															
Reference Plot #2 (North)															
Reference Plot #3 (Northeast)															
Reference Plot #4 (East)															
S78	U04-1113	Five Year pH Monitoring Report, Arcadis 2017a	2009, 2010, 2014	0-6"	2 mm sieve <sup>a</sup>		623657.066	26289231.249	32.1064820	-108.1296710	--	1857	Y	Removed	Outside vegetation alliance polygon extent
S79	U04-1114	Five Year pH Monitoring Report, Arcadis 2017a	2010, 2014	0-6"	2 mm sieve <sup>a</sup>		614951.000	26337640.457	32.1190110	-108.1113300	--	2457	Y	Removed	Outside vegetation alliance polygon extent
S80	U04-1115	Phase I Remedial Investigation Report, SRK 2008	2004	0-6"	2 mm sieve <sup>a</sup>		614751.033	26336845.318	32.6868800	-108.1038550	--	1418	Y	Removed	Outside vegetation alliance polygon extent
SS100	U04-1114	Phase I Remedial Investigation Report, SRK 2008	2004	0-6"	2 mm sieve <sup>a</sup>		590227.316	2632955.068	32.6222900	-108.1273220	207	157	Y	Removed	Outside vegetation alliance polygon extent
SS101	U04-1136	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		590163.969	26337693.377	32.6220920	-108.1108180	157	157	Y	Removed	Outside vegetation alliance polygon extent
SS102	U04-1137	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		641472.113	2651849.359	32.7632140	-108.0650930	234	164	Y	Removed	Outside vegetation alliance polygon extent
SS108	U04-1143	Phase I Remedial Investigation Report, SRK 2009	2006	0-1"	0.25 mm sieve		64421.219	2664449.121	32.7708120	-108.0241220	201	141	Y	Removed	Outside vegetation alliance polygon extent
SS110	U04-1145	Phase I Remedial Investigation Report, SRK 2009	2006	0-1"	0.25 mm sieve		631172.255	2662593.322	32.7439650	-108.0302110	252	176	Y	Removed	Outside vegetation alliance polygon extent
SS113	U04-1148	Phase I Remedial Investigation Report, SRK 2009	2006	0-1"	0.25 mm sieve		627617.986	2632796.780	32.7250040	-108.1268490	692	484	Y	Removed	Outside vegetation alliance polygon extent
SS114	U04-1148	Phase I Remedial Investigation Report, SRK 2009	2006	0-1"	0.25 mm sieve		631963.248	2653021.813	32.7370850	-108.0817240	209	146	Y	Removed	Outside vegetation alliance polygon extent
SS180	U04-1148	Phase I Remedial Investigation Report, SRK 2008	2006	0-6"	2 mm sieve		628954.890	2661385.893	32.7277630	-108.0339910	119	83	Y	Removed	Outside vegetation alliance polygon extent
SS120	U04-1149	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		628607.894	2654304.807	32.7233700	-108.0970040	125	125	Y	Removed	Outside vegetation alliance polygon extent
SS149	U04-1190	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		620848.117	2661067.331	32.7066060	-108.0349800	119	83	Y	Removed	Outside vegetation alliance polygon extent
SS150	U04-1191	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		591115.292	2649489.245	32.6247890	-108.0724030	628	440	Y	Removed	Outside vegetation alliance polygon extent
SS151	U04-1192	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		598309.346	2639041.201	32.6198750	-108.1193070	259	181	Y	Removed	Outside vegetation alliance polygon extent
SS152	U04-1193	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		598903.881	2640169.399	32.6131490	-108.1027050	237	168	Y	Removed	Outside vegetation alliance polygon extent
SS153	U04-1194	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		598306.127	2645235.684	32.6115400	-108.0861820	438	307	Y	Removed	Outside vegetation alliance polygon extent
SS154	U04-1195	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		597900.387	2646237.218	32.6133480	-108.0731910	372	260	Y	Removed	Outside vegetation alliance polygon extent
SS157	U04-1198	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		597030.387	2653503.773	32.6135830	-108.0695350	387	271	Y	Removed	Outside vegetation alliance polygon extent
SS99	U04-1134	Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		581830.815	2653347.367	32.5992920	-108.0725210	141	99	Y	Removed	Outside vegetation alliance polygon extent
West-01		Phase I Remedial Investigation Report, SRK 2008	2006	0-1"	0.25 mm sieve		620676.833	2624471.733	32.7592920	-108.0692420	93	85	Y	Removed	Outside vegetation alliance polygon extent
West Ref-01		Five Year Report Amendment Study, Arcadis 2017b	2008	0-6"	2 mm sieve <sup>a</sup>		620695.653	2629470.743	32.70590	-108.13377	--	805	Y	Removed	Outside vegetation alliance polygon extent
East Ref-01		Five Year Report Amendment Study, Arcadis 2017b	2008	0-6"	2 mm sieve <sup>a</sup>		623252.943	2631938.713	32.71065	-108.1383	--	873	Y	Removed	Outside vegetation alliance polygon extent
North East Ref-01		Five Year Report Amendment Study, Arcadis 2017b	2008	0-6"	2 mm sieve <sup>a</sup>		614754.960	2639621.006	32.6897	-108.1040	--	977	Y	Removed	Outside vegetation alliance polygon extent
STS-SS-2010-016		Five Year Report Amendment Study, Arcadis 2017b	2010	0-6"	2 mm sieve <sup>a</sup>		619650.369	2637650.024	32.7038	-108.1111	--	3065	Y	Removed	Outside vegetation alliance polygon extent
STS-SS-2010-017		Five Year pH Monitoring Report, Arcadis 2017a	2010	0-6"	2 mm sieve <sup>a</sup>		621050.366	2637336.024	32.7069680	-108.1120680	--	1120	Y	Removed	Outside vegetation alliance polygon extent
STS-SS-2010-018		Five Year pH Monitoring Report, Arcadis 2017a	2010	0-6"	2 mm sieve <sup>a</sup>		617464.171	2639694.326	32.6971380	-108.1064720	--	2060	Y	Removed	Outside vegetation alliance polygon extent
I-01	U06-3007	ACC Background Report, Chino 1996	1996	0-1"	not sieved		619916.129	2639851.163	32.7038770	-108.1095900	1300	1729	Y	Removed	Outside vegetation alliance polygon extent
I-03	U06-3022	ACC Background Report, Chino 1996	1996	0-1"	not sieved		597545.877	2645953.788	32.6003000	-108.1046040	554	499	Y	Removed	Borrow Area (WBT)
I-04	U06-3008	ACC Background Report, Chino 1996	1996	0-1"	not sieved		604970.956	2642871.008	32.6629280	-108.0936580	549	744	Y	Removed	Borrow Area (WBT)
I-06	U06-3018	ACC Background Report, Chino 1996	1996	0-1"	not sieved		603681.947	2645936.003	32.6597730	-108.0844200	943	489	Y	Removed	Borrow Area (WBT)
I-08	U06-3012	ACC Background Report, Chino 1996	1996	0-1"	not sieved		593488.853	2645548.099	32.6383020	-108.0852330	647	582	Y	Removed	Operations Reclamation Occurred
I-12	U06-3003	ACC Background Report, Chino 1996	1996	0-1"	not sieved		592389.858	2647114.896	32.6260780	-108.0801240	645	581	Y	Removed	Outside vegetation alliance polygon extent
I-13	U06-3001	ACC Background Report, Chino 1996	1996	0-1"	not sieved		595043.572	2642398.073	32.6080520	-108.0947350	216	194	Y	Removed	Outside vegetation alliance polygon extent
I-19	U04-1091	ACC Background Report, Chino 1996	1996	0-1"	not sieved		602186.191	2633270.742	32.6931340	-108.121810	773	694	Y	Removed	Outside vegetation alliance polygon extent
U04-1092		ACC Background Report, Chino 1996	1996	0-1"	not sieved		618466.844	2637069.898	32.6692350	-108.1083970	609	376	Y	Removed	IRA
U04-1093		ACC Background Report, Chino 1996	1996	0-1"	not sieved		614282.179	2638171.243	32.6683880	-108.1038660	1680	244	Y	Removed	Borrow Area (WBT)

Footnotes:  
 1. For copper for small grassland (excluding bird) applies to 0-6" soil sieved to < 2 mm (200 µm). For FS RAC for human health applies to 0-1" soil sieved to < 0.25 mm (60 µm). The ratio used to convert copper concentration of 0-1" sieved to < 0.25 mm sieved is 0.8. For bird pre-remedial action applies to 0-6" soil sieved to < 2 mm (200 µm). The ratio used to convert copper concentration by 1.5 (birds: withdrawal tailings). If unlisted, multiplied by 0.9 if outside: withdrawal tailings and by 1.3 if inside: withdrawal tailings. The 0.9 and 1.3 are from median ratios of unlisted 0-1" soil in 2009 lab reports (at 25 cm in tailings) to the data set for the study (at 0 to 6 inches).  
 2. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as IRA.  
 3. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as WBT.  
 4. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
 5. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as NIMD/OT.  
 6. Two SW samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two data points.  
 7. Removed samples with the asterisk are set to 32.7 for the Thiessen analysis. All other removed samples are completely voided.  
 8. -- means no data available at the location.



TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0-6"

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1" soil <sup>1</sup>	Northing (State Plane)	Latitude	Longitude	To Compare to HH Pre-FS RAC (250 µm sieve)	For Bird Pre-FS RAC (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
004-1054	S20	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618570.68	32.7028530	-108.1336590	4240	2668	Y	IRA	
004-1055	S21	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618571.288	32.7028603	-108.1336200	6570	4693	Y	Removed*	IRA
004-1056	S22	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618572.325	32.7028740	-108.1336490	5210	3647	Y	Removed*	IRA
004-1057	S23	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618573.378	32.7029055	-108.1336430	3030	2121	Y		
004-1058	S24	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618574.274	32.7001180	-108.1365960	3910	2737	Y		
004-1059	S25	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618573.595	32.7001280	-108.1349770	4630	3241	Y		
004-1060	S26	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617573.221	32.6973620	-108.1368300	2600	1820	Y		
004-1061	S27	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617722.447	32.6973730	-108.1368300	3150	2205	Y		
004-1062	S28	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618573.603	32.6973780	-108.1337560	5940	4088	Y	Removed*	IRA
004-1063	S29	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618574.834	32.6946170	-108.1394170	1690	1183	Y		
004-1064	S30	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618574.834	32.6946170	-108.1394170	1690	1183	Y		
004-1065	S31	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618576.028	32.6946400	-108.1326880	3120	2184	Y	Removed*	IRA
004-1066	S32	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		614738.172	32.6895200	-108.1310400	2440	1708	Y		
004-1067	S33	Phase I Remedial Investigation Report, SRK 2008	2004	0-6"	2 mm sieve		614819.381	32.6862530	-108.1274220	2570	1799	Y		
004-1068	S34	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		614621.475	32.6892800	-108.1242790	4340	3038	Y	Removed*	Borrow Area
004-1069	S35	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		619917.989	32.7043740	-108.1128100	3270	2289	Y		
004-1070	S36	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		620098.555	32.6973930	-108.1101670	3970	2779	Y		
004-1071	S37	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		619153.985	32.6993850	-108.1134070	1900	1330	Y	Removed*	IRA
004-1072	S38	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		619114.575	32.7016700	-108.1098720	4280	2996	Y	Removed*	IRA
004-1073	S39	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		619215.672	32.6994470	-108.1067260	2470	1729	Y		
004-1074	S40	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		619095.517	32.6997930	-108.1035220	3610	2527	Y		
004-1075	S41	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618114.562	32.6969812	-108.1132060	8170	5719	Y	Removed*	IRA
004-1076	S42	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618115.788	32.6972789	-108.1094940	5780	4048	Y	Removed*	IRA
004-1077	S43	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		618116.574	32.6999575	-108.1067330	2230	1551	Y		
004-1078	S44	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617127.889	32.6994456	-108.1026980	908	535	Y		
004-1080	S46	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617127.889	32.6994456	-108.1026980	908	535	Y		
004-1081	S47	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617127.889	32.6994456	-108.1026980	908	535	Y		
004-1082	S48	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617127.889	32.6994456	-108.1026980	908	535	Y		
004-1083	S49	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617032.038	32.6974039	-108.1107300	3000	2085	Y	Removed*	IRA
004-1084	S50	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617032.038	32.6974039	-108.1107300	3000	2085	Y	Removed*	IRA
004-1085	S51	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617032.038	32.6974039	-108.1107300	3000	2085	Y	Removed*	IRA
004-1086	S52	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617032.038	32.6974039	-108.1107300	3000	2085	Y	Removed*	IRA
004-1087	S53	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617032.038	32.6974039	-108.1107300	3000	2085	Y	Removed*	IRA
004-1088	S54	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617032.038	32.6974039	-108.1107300	3000	2085	Y	Removed*	IRA
004-1089	S55	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617032.038	32.6974039	-108.1107300	3000	2085	Y	Removed*	IRA
004-1090	S56	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617032.038	32.6974039	-108.1107300	3000	2085	Y	Removed*	IRA
004-1091	S57	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617032.038	32.6974039	-108.1107300	3000	2085	Y	Removed*	IRA
004-1092	S58	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617032.038	32.6974039	-108.1107300	3000	2085	Y	Removed*	IRA
004-1093	S59	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		616963.075	32.6961331	-108.1248320	14100	9870	Y	Removed*	Borrow Area (WBT)
004-1094	S60	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		619304.400	32.6974787	-108.1246840	18300	12810	Y	Removed*	Borrow Area (WBT)
004-1095	S61	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		619304.400	32.6974787	-108.1246840	18300	12810	Y	Removed*	Borrow Area (WBT)
004-1097	S62	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617369.661	32.6973200	-108.1231390	20100	14070	Y	Removed*	Operations Reclamation Occurred
004-1098	S63	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		617369.661	32.6973200	-108.1231390	20100	14070	Y	Removed*	Operations Reclamation Occurred
004-1099	S64	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		638489.167	32.6931631	-108.1327770	689	482	Y	Removed*	Operations Reclamation Occurred
004-1100	S65	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		634421.253	32.7494030	-108.1314170	660	462	Y	Removed*	IRA
004-1101	S66	Phase I Remedial Investigation Report, SRK 2008	2004	0-1"	0.25 mm sieve		632606.041	32.7397040	-108.1314510	789	552	Y	Removed*	IRA

Footnotes:  
 \*For copper for small grain testing that applies to 0.6" soil sieved to < 0.25 mm (200 µm). For FS RAC for human health applies to 0.6" soil sieved to < 0.25 mm (200 µm). The ratio used to convert copper concentration of 0.6" sieved to < 0.25 mm sieved is 0.6. For windblown tailings, the ratio used to convert copper concentration from 0.6" sieved to < 0.25 mm sieved is 0.7. For windblown tailings, the ratio used to convert copper concentration from 0.6" sieved to < 0.25 mm sieved is 0.7. For windblown tailings, the ratio used to convert copper concentration from 0.6" sieved to < 0.25 mm sieved is 0.7.  
 1. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as IRA.  
 2. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as IRA.  
 3. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as NMDOT.  
 4. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as NMDOT.  
 5. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04-1097 and U04-1098.  
 6. Removed samples with the asterisk are set to 327 for the Theissen's analysis. All other removed samples are completely voided.  
 7. -- means no data available at the location.

TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0.6"

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0.1" soil*	Northing (State Plane)	Latitude	Longitude	Copper Concentration <sup>1</sup> To Compare to HH Pre FS RAC (250 µm sieve)	For Bird Pre-FS RAC (2000 µm sieve)	Pre-Interim Remedial Action (IRA)	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
U04-1102	S67	Phase I Remedial Investigation Report, SRK 2008	2004	0.1"	0.25 mm sieve		639502.857	32.7392950	-108.1308900	889	623	Y	Removed	IRA and outside vegetation polygons
U04-1103	S68	Phase I Remedial Investigation Report, SRK 2008	2004	0.1"	0.25 mm sieve		628776.376	32.7428140	-108.1334220	846	592	Y	Removed	IRA and outside vegetation polygons
U04-1104	S69	Phase I Remedial Investigation Report, SRK 2008	2004	0.1"	0.25 mm sieve		628776.687	32.7228750	-108.1334440	710	497	Y	Removed	IRA
U04-1105	S70	Phase I Remedial Investigation Report, SRK 2008	2004	0.1"	0.25 mm sieve		624446.087	32.93119.891	-108.1362550	2280	1596	Y	Removed	IRA
U04-1106	S71	Phase I Remedial Investigation Report, SRK 2008	2004	0.1"	0.25 mm sieve		622021.900	32.93472.066	-108.1444570	5360	3745	Y	Removed	Borrow Area
U04-1107	S72	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		611902.416	32.93139.171	-108.1284500	1160	1160	Y	Removed	
U04-1108	S73	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		607892.266	32.93276.939	-108.1260240	1290	1290	Y	Removed	
U04-1109	S74	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		604073.324	32.93226.309	-108.1286240	529	529	Y	Removed	
U04-1110	S75	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		600284.039	32.93391.253	-108.1482740	940	940	Y	Removed	
U04-1111	S76	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		598474.705	32.93196.860	-108.1293190	278	278	Y	Removed	
U04-1112	S77	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		594805.310	32.93438.935	-108.1222310	267	267	Y	Removed	
U04-1113	S78	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		611644.482	32.93932.182	-108.1055320	1440	1440	Y	Removed	
U04-1116	S81	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		611644.477	32.93832.465	-108.0992470	875	875	Y	Removed	
U04-1117	S82	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		611644.477	32.93832.465	-108.0992470	455	455	Y	Removed	
U04-1118	S83	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		611644.466	32.93952.841	-108.0811830	358	358	Y	Removed	
U04-1119	S84	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		609070.884	32.93524.941	-108.1049170	362	362	Y	Removed	Borrow Area (WBT)
U04-1120	S85	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		609053.370	32.941276.198	-108.0992260	451	451	Y	Removed	IRA
U04-1121	S86	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		609040.375	32.94354.287	-108.0918220	513	513	Y	Removed	IRA, and Co-located with FID 18"
U04-1122	S87	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		608718.728	32.94702.118	-108.0816590	309	309	Y	Removed	IRA
U04-1123	S88	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		608982.076	32.94974.237	-108.0715320	484	484	Y	Removed	IRA
U04-1124	S89	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		608923.351	32.95238.921	-108.0632720	399	399	Y	Removed	IRA
U04-1125	S90	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		610033.010	32.95419.877	-108.0555680	255	255	Y	Removed	IRA
U04-1126	S91	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		608644.353	32.94432.274	-108.0965990	926	926	Y	Removed	Borrow Area (WBT)
U04-1127	S92	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		608551.137	32.94322.150	-108.0920030	591	591	Y	Removed	IRA
U04-1128	S93	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		605114.831	32.94983.157	-108.0810480	308	308	Y	Removed	IRA
U04-1129	S94	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		605980.775	32.94936.070	-108.0729140	313	313	Y	Removed	IRA
U04-1130	S95	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		605800.724	32.95232.267	-108.0653480	494	494	Y	Removed	IRA
U04-1131	S96	Phase I Remedial Investigation Report, SRK 2008	2004	0.5"	2 mm sieve		605800.724	32.95232.267	-108.0653480	237	237	Y	Removed	IRA
U04-1132	S97	Phase I Remedial Investigation Report, SRK 2008	2004	0.1"	0.25 mm sieve		63472.988	32.93310.020	-108.1471360	412	288	Y	Removed	Outside vegetation alliance polygon extent
U04-1133	S98	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		63974.103	32.93411.123	-108.1249600	475	333	Y	Removed	Outside vegetation alliance polygon extent
U04-1134	SS103	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		63406.149	32.93136.184	-108.1351980	497	248	Y	Removed	Outside vegetation alliance polygon extent
U04-1135	SS104	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		63430.157	32.93136.184	-108.1351980	276	163	Y	Removed	Outside vegetation alliance polygon extent
U04-1141	SS106	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		629459.283	32.93498.330	-108.1246460	534	372	Y	Removed	Outside vegetation alliance polygon extent
U04-1142	SS107	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		630571.157	32.93668.927	-108.1233230	194	136	Y	Removed	Outside vegetation alliance polygon extent
U04-1144	SS109	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		628541.137	32.94344.410	-108.1272640	597	418	Y	Removed	Outside vegetation alliance polygon extent
U04-1145	SS111	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		626557.141	32.94233.451	-108.1025640	551	388	Y	Removed	Outside vegetation alliance polygon extent
U04-1147	SS112	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		624722.087	32.94607.043	-108.1172320	568	381	Y	Removed	Outside vegetation alliance polygon extent
U04-1150	SS115	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		622615.631	32.94702.620	-108.1401570	3800	2660	Y	Removed	Outside vegetation alliance polygon extent
U04-1151	SS116	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		622615.631	32.94702.620	-108.1401570	1460	1022	Y	Removed	Outside vegetation alliance polygon extent
U04-1152	SS117	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		619724.447	32.94687.532	-108.1105660	4450	3115	Y	Removed	Outside vegetation alliance polygon extent
U04-1154	SS118D	Phase I Remedial Investigation Report, SRK 2008	2006	0.5"	2 mm sieve		619356.828	32.94300.467	-108.0936070	259	259	Y	Removed	Outside vegetation alliance polygon extent
U04-1155	SS121	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		616358.262	32.94464.416	-108.0884660	896	627	Y	Removed	Outside vegetation alliance polygon extent
U04-1159	SS122	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		617920.343	32.95110.146	-108.0765520	119	83	Y	Removed	Outside vegetation alliance polygon extent
U04-1160	SS123	Phase I Remedial Investigation Report, SRK 2008	2006	0.1"	0.25 mm sieve		618270.278	32.95847.824	-108.0381800	449	314	Y	Removed	Outside vegetation alliance polygon extent
U04-1162	SS124D	Phase I Remedial Investigation Report, SRK 2008	2006	0.5"	2 mm sieve		613136.715	32.95627.700	-108.0920930	523	523	Y	Removed	Outside vegetation alliance polygon extent
U04-1164	SS125D	Phase I Remedial Investigation Report, SRK 2008	2006	0.5"	2 mm sieve		614418.219	32.96924.166	-108.0809110	166	166	Y	Removed	Outside vegetation alliance polygon extent

Footnotes:  
 \* For copper for small ground testing that applies to 0.6" soil sieved to < 0.2 mm (200 µm), the FS RAC for human health applies to 0.1" soil sieved to < 0.25 mm (60 µm). The ratio used to convert copper concentration of 0.1" sieved to < 0.25 mm sieved is 0.6" sieved to < 2 mm sieved as < 2 mm sieved is from Appendix A (median rate of SS samples of Arcadis (2011)). Multiplied copper concentration by 0.7 if outside windblown tailings (and by 1.5 if inside windblown tailings). If unsited, multiplied by 0.9 if outside windblown tailings and by 1.3 if inside windblown tailings. The 0.9 and 1.3 are from median ratios of unsited/0.1" soil in 2009 lab reports (at 25 µm in tailings) compared to the data set for 0.6" sieved.  
 † Data were collected for ecological endpoint and were not corrected to human health values.  
 ‡ A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as IRA.  
 § A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
 ¶ A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as MMDOT.  
 ¶ A duplicate sample with the same Loc ID has been recorded in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04-1110 and U04-1111.  
 \*\* Removed samples with the asterisk are set to 327 for the Theissen's analyses. All other removed samples are completely voided.  
 -- means no data available at the location.

TABLE 3-2 STSIU Surface Soil Copper Results Adjusted to 0-6"

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Table with columns: Location ID, Alternate ID, Source Report, Year Collected, Original Soil Depth, Sieving Status, Windblown Tailing (WBT) adjustment needed for 0-1" soil, Northing (State Plane), Easting (State Plane), Latitude, Longitude, Copper Concentration (To Compare to HH Pre-Remedial Action Dataset (250 um sieve)), For Bird Pre-Remedial Action Dataset (2000 um sieve), Pre-Interim Remedial Action (IRA) Dataset, Reason Removed (also see this column's footnotes), Operations Reclamation Occurred

Footnotes: 1. For copper for small grain testing this applies to 0-6" soil sieved to < 2 mm (2000 um). For ES-BAC for human health applies to < 0.25 mm (625 um) soil. The ratio used to convert copper concentration of 0.1" sieved at < 0.25 mm sieved to < 2 mm sieved is from Appendix A (median ratio of SSS samples of Arcadis 2011). Multiplied copper concentration by 0.7 if outside windblown tailings (and by 1.5 if inside windblown tailings). If unsieved, multiplied by 0.9 if outside windblown tailings and by 1.3 if inside windblown tailings. The 0.9 and 1.3 are from median ratios of unsieved 0.1" soil in 2009 lab reports (at 25 um) in tailings outside of the ore stockpile (0-6 inches).  
2. For copper for ecological endpoints and were not corrected to human health values.  
3. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as IRA.  
4. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as IRA.  
5. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
6. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arcadis 2009. This sample was marked as NIMD0T.  
7. Two samples were taken from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04-1176 and U04-1177.  
8. Removed samples with the asterisk are set to 327 for the Thiessens analyses. All other removed samples are completely voided.  
-- means no data available at the location.

TABLE 3-2 STSIU Surface Soil Copper Results Adjusted to 0.6"

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Table with columns: Location ID, Alternate ID, Source Report, Year Collected, Original Soil depth, Sieving Status, 1 if Windblown Tailing (WBT) adjustment needed for 0-1" soil, Northing (State Plane), Easting (State Plane), Latitude, Longitude, Copper Concentration (To Compare to HH Pre FS RAC (250 μm sieve)), Pre-Interim Remedial Action (IRA) Dataset, Reason Removed (also see this column's footnotes), and Removed in Post-IRA Dataset.

Footnotes: 1) For copper for small ground testing but applied to 0.6" soil sieved to < 0.25 mm (60 μm). See FS RAC for human health applies to 0.1" soil sieved to < 0.25 mm (60 μm). The ratio used to convert copper concentration of 0.1" sieved at < 0.25 mm sieved to 0.6" sieved at < 2 mm sieved is from Appendix A (median ratio of 0.34). Multiplied copper concentration by 0.7 if outside windblown tailings (and on a transect) compared to sieved 0.6" sieved soil in Drexler lab report in Appendix D or the pH Monitoring Report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (FD 8, 18, 20). In tailings, outside of the windblown tailings, the data is based on 0.6 inch... data were collected for ecological endpoint and were not connected to human health values.

















TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0-6"

STSIU Feasibility Study  
Freepport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1" soil'	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	Copper Concentration <sup>1</sup> To Compare to FS RAC (250 µm sieve)	For Bird Pre-Remedial Action (IRA) Dataset	Pre-Interim Remedial Action (IRA) Dataset	Reason Removed (also see this column's footnotes)
1001-SG		Railroad Interim Removal Action, Collier 2013	2012	0-2"	0.25 mm sieve		615626.416	2630754.467	32.6925780	-108.1334820	2.685	N	Remove <sup>2</sup>	IRA
1002-SG		Railroad Interim Removal Action, Collier 2013	2012	0-2"	0.25 mm sieve		615626.366	2630756.116	32.6924700	-108.1334750	1.508	N	Remove <sup>2</sup>	IRA
1003-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621252.048	2633228.258	32.7075100	-108.1254890	2.35	N		
1004-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621251.513	2633078.059	32.7075070	-108.1254760	1.10	N		
1005-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621251.127	26332928.666	32.7075050	-108.1254630	2.908	N		
1006-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621401.238	26332928.666	32.7075050	-108.1254670	2.00	N		
1007-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621560.711	2633277.514	32.7083280	-108.1265530	83	N		
1008-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621560.366	26332938.948	32.7083280	-108.1264680	743	N		
1009-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621561.027	2633078.061	32.7083200	-108.1265800	863	N		
1010-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621401.516	2633079.292	32.7079200	-108.1265760	1796	N		
1011-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621502.107	2633228.251	32.7079200	-108.1265490	947	N		
1012-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621502.707	2633218.340	32.7081900	-108.1265240	1942	N		
1013-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621250.102	2633278.883	32.7075000	-108.1266510	1467	N		
1014-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621249.637	26332628.854	32.7075000	-108.1274380	697	N		
1015-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621249.553	26332478.688	32.7074980	-108.1274260	1834	N		
1016-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621249.334	26332178.692	32.7074980	-108.1284140	116	N		
1017-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621248.716	26332028.710	32.7074920	-108.1293890	142	N		
1018-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621248.517	26331879.128	32.7074900	-108.1298770	163	N		
1019-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621247.954	26331728.715	32.7074870	-108.1303640	598	N		
1020-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621247.427	26331579.347	32.7074850	-108.1306500	1200	N		
1021-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621400.522	26332628.665	32.7079130	-108.1274400	134	N		
1022-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621389.539	26332478.305	32.7079110	-108.1274400	97	N		
1023-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621389.520	2633238.292	32.7079090	-108.1284170	240	N		
1024-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621569.522	2633277.810	32.7083210	-108.1294190	867	N		
1025-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621569.492	2633277.810	32.7083210	-108.1294190	807	N		
1026-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621569.326	2633267.803	32.7083250	-108.1274320	346	N		
1027-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621569.139	2633277.828	32.7083180	-108.1274440	824	N		
1028-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621568.648	2633207.848	32.7083180	-108.1280600	674	N		
1029-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621568.340	2633197.866	32.7083140	-108.1293840	536	N		
1030-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621567.869	2633177.829	32.7083120	-108.1303700	1278	N		
1031-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621567.501	2633142.819	32.7083090	-108.1313450	626	N		
1032-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621387.161	26331428.341	32.7079090	-108.1313450	629	N		
1033-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621387.143	26331578.415	32.7079080	-108.1314240	1029	N		
1034-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621387.047	26331728.304	32.7079000	-108.1303670	203	N		
1035-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621386.447	26331878.194	32.7079020	-108.1295800	54	N		
1036-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621386.046	26332178.374	32.7079000	-108.1295820	119	N		
1037-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621386.000	26332178.374	32.7079000	-108.1295800	119	N		
1038-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621247.013	2633379.865	32.7074820	-108.1313390	574	N		
1039-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621102.176	2633379.865	32.7074820	-108.1249800	756	N		
1040-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621101.897	2633229.077	32.7070970	-108.1249800	1551	N		
1041-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621101.401	2633079.664	32.7070950	-108.1254850	2438	N		
1042-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621101.052	26332928.866	32.7070950	-108.1254730	927	N		
1043-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621101.052	26332928.866	32.7070950	-108.1264610	3923	N		
1044-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621100.703	26332779.036	32.7070910	-108.1264610	2580	N		
1045-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621100.319	26332628.838	32.7070890	-108.1274370	2074	N		

<sup>1</sup> For copper for small ground testing bird applies to 0.6" soil sieved to < 2 mm (200 µm). For FS RAC: for human health applies to 0.4" soil sieved to < 0.25 mm (60 µm). The ratio used to convert copper concentration of 0.4" sieved to < 0.25 mm sieved is 0.6" sieved to < 0.25 mm sieved is 1.0. For human health: for human health applies to 0.4" soil sieved to < 0.25 mm (60 µm). The ratio used to convert copper concentration of 0.4" sieved to < 0.25 mm sieved is 0.6" sieved to < 0.25 mm sieved is 1.0. For human health: for human health applies to 0.4" soil sieved to < 0.25 mm (60 µm). The ratio used to convert copper concentration by 1.5 if fish: windblown tailings). If uncorrected, multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. This 0.9 and 1.3 are: from median ratios of uncorrected 0.4" soil in 2009 lab reports (at 25 m on transect) compared to sieved 0.4" co-located soil in Dreier lab report in Appendix D or the pH Monitoring Report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (P-D 8, 18, 20). In tailings, the ratio is based on the data for the site (at 0.6 m) and the ratio is based on the data for the site (at 0.6 m).  
<sup>2</sup> Data were collected for ecological endpoint and were not corrected to human health values.  
<sup>3</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>4</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>5</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
<sup>6</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NIMDOT.  
<sup>7</sup> Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
<sup>8</sup> Removed samples with the asterisk are set to 32.7 for the Theissen's analyses. All other removed samples are completely voided.  
<sup>9</sup> -- means no data available at the location.

TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0-6"

STSIU Feasibility Study  
Frapport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1' soil*	Northing (State Plane)	Eastings (State Plane)	Latitude	Longitude	Copper Concentration <sup>1</sup> To Compare to HH Pre-Remedial FS RAC (250 µm sieve)	For Bird Pre-Remedial FS RAC (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
1047-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1048-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1049-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1050-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1051-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1052-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1053-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1054-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1055-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1056-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1057-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1058-SG-A		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1058-SG-B		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1059-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1060-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1061-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1062-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1063-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1064-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1065-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1066-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1067-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1068-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1069-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1070-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1071-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1072-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
1073-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2000-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2001-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2002-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2003-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2004-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2005-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2006-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2007-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2008-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2009-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			
2009-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		2632478.764	327070860	-108.1273250	2294	1606	N			

Footnotes:  
 \* For copper for small ground (soil) applies to 0-6" soil sieved to < 2 mm (2000 µm). For FS RAC for human health applies to 0-1" soil sieved to < 0.25 mm (950 µm). The ratio used to convert copper concentration of 0-1" sieved to < 0.25 mm sieved is 0.67.  
 \*\* For copper for small ground (soil) applies to 0-6" soil sieved to < 2 mm (2000 µm). For FS RAC for human health applies to 0-1" soil sieved to < 0.25 mm (950 µm). The ratio used to convert copper concentration of 0-1" sieved to < 0.25 mm sieved is 0.67.  
 1. 1.5 ft (18 in) windblown tailings. If uncorrected, multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. The 0.9 and 1.3 are: from median ratios of uncorrected 0-1" soil in 2009 lab reports; (at 25 m in transect) compared to sieved 0-6" co-located soil in Drexler lab report in Appendix D or the pH Monitoring Report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (P-D 8, 18, 20). In tailings, the data are the average of 6 samples.  
 2. 1.5 ft (18 in) windblown tailings. If uncorrected, multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. The 0.9 and 1.3 are: from median ratios of uncorrected 0-1" soil in 2009 lab reports; (at 25 m in transect) compared to sieved 0-6" co-located soil in Drexler lab report in Appendix D or the pH Monitoring Report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (P-D 8, 18, 20). In tailings, the data are the average of 6 samples.  
 3. Data were collected for ecological endpoint and were not corrected to human health values.  
 4. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 5. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 6. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
 7. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NIMDOT.  
 8. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
 9. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
 10. Removed samples with the asterisk are set to 32.7 for the Theissen's analyses. All other removed samples are completely voided.  
 -- means no data available at the location.







TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0.6\*

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1' soil <sup>1</sup>	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	To Compare to FS RAC (250 µm sieve)	Copper Concentration <sup>1</sup>	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
5064-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622299.913	2632475.068	32.7103850	-108.1274570	475	333	N		
5065-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622300.267	2632625.082	32.7103870	-108.1274570	141	94	N		
5066-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622300.888	2632775.070	32.7103880	-108.1264810	2047	1433	N		
5067-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622302.109	2632928.033	32.7103910	-108.1264810	1914	1340	N		
5068-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622302.447	2633078.023	32.7103960	-108.1259940	261	183	N		
5069-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622303.521	2633225.062	32.7103980	-108.1259900	1145	802	N		
5070-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622303.979	2633375.024	32.7104020	-108.1254320	1256	879	N		
5071-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622304.366	2633525.074	32.7104030	-108.1245320	1527	932	N		
5072-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622304.718	2633676.103	32.7104030	-108.1244430	1527	1069	N		
5073-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622305.089	2633826.056	32.7104030	-108.1235550	336	235	N		
5074-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622305.447	2633974.881	32.7104070	-108.1230710	777	544	N		
5075-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622306.883	2634126.340	32.7099520	-108.1323310	1185	830	N		
5076-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622307.468	2634276.109	32.7099540	-108.1315560	1700	1190	N		
5077-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622308.147	2634426.081	32.7099540	-108.1308680	916	641	N		
5078-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622309.554	2634576.337	32.7099590	-108.1300800	821	575	N		
5079-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622310.925	2634726.497	32.7099590	-108.1298930	1770	119	N		
5080-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622312.320	2634876.110	32.7099690	-108.1294930	534	374	N		
5081-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622313.743	2635026.044	32.7099690	-108.1284180	330	231	N		
5082-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622315.113	2635176.097	32.7099700	-108.1284180	145	102	N		
5083-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622316.468	2635326.411	32.7099690	-108.1274290	722	505	N		
5084-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622317.847	2635476.081	32.7099700	-108.1274290	125	88	N		
5085-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622319.213	2635625.941	32.7099700	-108.1274570	845	592	N		
5086-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622320.579	2635775.114	32.7099750	-108.1266620	222	155	N		
5087-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622321.947	2635925.075	32.7099840	-108.1259910	432	302	N		
5088-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622323.320	2636076.459	32.7099840	-108.1259910	1034	724	N		
5089-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622324.693	2636226.121	32.7099840	-108.1250160	918	643	N		
5090-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622326.066	2636376.443	32.7099840	-108.1245260	1035	739	N		
5091-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622327.439	2636526.104	32.7099910	-108.1245260	430	101	N		
5092-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622328.812	2636676.023	32.7099910	-108.1235660	206	104	N		
5093-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622330.185	2636826.344	32.7099910	-108.1235660	206	104	N		
5094-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622331.558	2636976.338	32.7099910	-108.1235660	206	104	N		
5095-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622332.931	2637126.322	32.7099910	-108.1235660	206	104	N		
5096-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622334.304	2637276.316	32.7099910	-108.1235660	206	104	N		
5097-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622335.677	2637426.310	32.7099910	-108.1235660	206	104	N		
5098-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622337.050	2637576.304	32.7099910	-108.1235660	206	104	N		
5099-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622338.423	2637726.298	32.7099910	-108.1235660	206	104	N		
5100-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622339.796	2637876.292	32.7099910	-108.1235660	206	104	N		
5101-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622341.169	2638026.286	32.7099910	-108.1235660	206	104	N		
5102-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622342.542	2638176.280	32.7099910	-108.1235660	206	104	N		
5103-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622343.915	2638326.274	32.7099910	-108.1235660	206	104	N		
5104-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622345.288	2638476.268	32.7099910	-108.1235660	206	104	N		
5105-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622346.661	2638626.262	32.7099910	-108.1235660	206	104	N		
5106-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622348.034	2638776.256	32.7099910	-108.1235660	206	104	N		
5107-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622349.407	2638926.250	32.7099910	-108.1235660	206	104	N		
5108-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622350.780	2639076.244	32.7099910	-108.1235660	206	104	N		
5109-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622352.153	2639226.238	32.7099910	-108.1235660	206	104	N		
5110-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622353.526	2639376.232	32.7099910	-108.1235660	206	104	N		
5111-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622354.899	2639526.226	32.7099910	-108.1235660	206	104	N		
5112-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622356.272	2639676.220	32.7099910	-108.1235660	206	104	N		

Footnotes:  
 \* For copper for small grain (sieving) but applies to 0.6" soil sieved to < 0.25 mm (200 µm). For FS RAC: for human health applies to 0.1" soil sieved to < 0.25 mm (60 µm). The ratio used to convert copper concentration of 0.1" sieved to 0.6" sieved is 0.6. For FS RAC: for human health applies to 0.1" soil sieved to < 0.25 mm (60 µm). The ratio used to convert copper concentration of 0.1" sieved to 0.6" sieved is 0.6.  
 1. If field: windblown tailings; if unmined: multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. The 0.9 and 1.3 are: from median ratios of unmined 0.1" soil in 2009 lab reports; (at 25 cm in tailings) and from the data on 0.6" sieved soil.  
 2. Data were collected for ecological endpoint and were not corrected to human health values.  
 3. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 4. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as 21 sieved.  
 5. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
 6. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NIMDOT.  
 7. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used (P-ID 8, 18, 19).  
 8. Removed samples with the asterisk are set to 327 for the Theissen's analyses. All other removed samples are completely voided.  
 9. -- means no data available at the location.

TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0.6"

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1" soil <sup>1</sup>	Northing (State Plane)	Eastings (State Plane)	Latitude	Longitude	To Compare to HH Pre-Remedial Action Dataset (250 µm sieve)	For Bird Pre-Remedial Action Dataset (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
5113-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622002.516	2633225.837	32.7095720	-108.1255040	2387	1671	N	Removed	Co-located samples <sup>2</sup>
5114-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622002.300	2633236.007	32.7095700	-108.1255130	541	379	N		
5115-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622002.869	2633556.626	32.7095760	-108.1242260	2194	1536	N		
5116-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622003.077	2633676.126	32.7095820	-108.1240380	149	104	N		
5117-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622003.033	2633826.355	32.7095800	-108.1235630	1290	903	N		
5118-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622003.893	2633976.114	32.7095820	-108.1228750	134	94	N		
5119-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		622004.419	2634277.594	32.7095860	-108.1220850	313	219	N		
5122-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621846.464	2633097.114	32.7091250	-108.1326130	241	169	N		
5123-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621846.377	2633127.252	32.7091250	-108.1326260	919	643	N		
5125-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621847.113	26331576.939	32.7091330	-108.1306830	654	458	N		
5127-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621846.338	2633176.301	32.7091330	-108.1303760	109	76	N		
5128-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621846.446	2633187.115	32.7091390	-108.1296870	813	569	N		
5129-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621847.856	26332026.354	32.7091390	-108.1294020	952	666	N		
5130-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621849.445	2633217.052	32.7091430	-108.1281200	86	60	N		
5131-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621849.419	2633226.891	32.7091450	-108.1284250	176	123	N		
5132-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621849.674	2633247.156	32.7091480	-108.1278370	637	448	N		
5133-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621850.295	2633267.197	32.7091500	-108.1274490	103	72	N		
5135-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621852.288	2633377.125	32.7091610	-108.1250110	345	242	N		
5136-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621852.686	2633527.041	32.7091630	-108.1245230	724	507	N		
5140-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621853.075	2633677.114	32.7091650	-108.1240360	539	377	N		
5141-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621853.301	2633627.125	32.7091680	-108.1235480	1303	912	N		
5142-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621853.851	2633977.041	32.7091700	-108.1230600	1681	1184	N		
5143-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621854.282	2634127.114	32.7091720	-108.1225730	326	228	N		
5145-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621854.473	2634277.124	32.7091770	-108.1219960	357	250	N		
5146-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621855.174	2634427.319	32.7091770	-108.1215960	1572	1058	N		
5148-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621855.959	2634977.521	32.7091730	-108.1211100	489	328	N		
5149-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621856.374	2635127.532	32.7091750	-108.1206230	1088	748	N		
5150-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621702.259	2633377.853	32.7091490	-108.1250600	1025	478	N		
5152-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621702.693	2633527.853	32.7091510	-108.1245730	926	406	N		
5153-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621703.068	2633677.854	32.7091530	-108.1240860	753	517	N		
5155-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621703.889	2633977.854	32.7091550	-108.1235990	263	188	N		
5156-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621704.279	2634127.935	32.7091580	-108.1231120	463	114	N		
5170-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621705.621	2634277.515	32.7087620	-108.1220820	1511	1054	N		
5171-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621705.663	2634427.695	32.7087640	-108.1216920	865	606	N		
5172-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621705.456	2634577.968	32.7087670	-108.1211070	2160	1512	N		
5173-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621545.984	2633977.968	32.7083010	-108.1326080	2746	1922	N		
5174-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621546.372	2634127.968	32.7083030	-108.1322120	4763	3355	N		
5175-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621546.786	2634277.921	32.7083050	-108.1318330	1603	1122	N		
5189-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621552.293	2633377.899	32.7083370	-108.1250060	691	484	N		
5190-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621552.681	2633527.811	32.7083390	-108.1245180	1277	894	N		
5191-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621553.061	2633677.891	32.7083410	-108.1240300	939	657	N		
5192-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621553.495	2633827.803	32.7083430	-108.1235430	230	161	N		
5193-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621553.885	2633977.814	32.7083450	-108.1230550	92	64	N		
5194-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621554.276	2634127.895	32.7083480	-108.1225670	448	314	N		

Footnotes:  
<sup>1</sup> For copper for small ground testing bird applies to 0.6" soil sieved to < 0.2 mm (2000 µm). For ES BAC: for human health applies to 0.1" soil sieved to < 0.25 mm (650 µm). The ratio used to convert copper concentration of 0.1" sieved to 0.6" sieved is 0.6. For lead and cadmium, the ratio used to convert copper concentration of 0.1" sieved to 0.6" sieved is 0.6. For arsenic, the ratio used to convert copper concentration of 0.1" sieved to 0.6" sieved is 0.6. For nickel, the ratio used to convert copper concentration of 0.1" sieved to 0.6" sieved is 0.6. For zinc, the ratio used to convert copper concentration of 0.1" sieved to 0.6" sieved is 0.6.  
<sup>2</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>3</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>4</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
<sup>5</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NIMDOT.  
<sup>6</sup> Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two identical samples U04.  
<sup>7</sup> Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two identical samples U04.  
<sup>8</sup> Removed samples with the asterisk are set to 32.7 for the Theissen's analyses. All other removed samples are completely voided.  
<sup>9</sup> - means no data available at the location.

TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0.6<sup>1</sup>

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1" soil <sup>1</sup>	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	To Compare to HH Pre-FS RAC (250 µm sieve)	Copper Concentration <sup>1</sup>	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
5195-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621454.668	263427.007	32.7083520	-108.1228800	863	604	N		
5195-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621455.060	263447.887	32.7083520	-108.1219200	1540	1358	N		
5197-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621455.453	263457.769	32.7083540	-108.1211400	670	469	N		
5198-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621386.869	2631128.316	32.7078910	-108.1321800	1723	1206	N		
5200-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621386.733	26311278.266	32.7078920	-108.1313000	2574	1802	N		
5214-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621402.878	2633373.266	32.7079260	-108.1250300	1612	1128	N		
5215-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621403.103	26333678.311	32.7079260	-108.1245150	1667	1167	N		
5217-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621403.493	26333629.392	32.7079310	-108.1234400	404	293	N		
5218-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621403.883	26333578.304	32.7079330	-108.1234500	395	277	N		
5219-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621404.274	2634128.286	32.7079350	-108.1228650	988	692	N		
5220-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621404.665	2634428.102	32.7079370	-108.1220770	1786	1257	N		
5221-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621405.051	2634478.266	32.7079420	-108.1218900	554	388	N		
5222-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621405.451	2634478.266	32.7079440	-108.1210200	1885	1327	N		
5228-SG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621246.780	2631278.106	32.7074810	-108.1318270	1018	713	N		
5242-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621252.712	2633528.686	32.7075140	-108.1245130	226	158	N		
5243-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621253.101	2633978.689	32.7075190	-108.1240250	104	73	N		
5244-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621253.490	2633828.681	32.7075190	-108.1235370	155	109	N		
5245-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621253.880	2633978.684	32.7075210	-108.1234500	713	499	N		
5246-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621254.271	2634128.107	32.7075230	-108.1228650	881	617	N		
5247-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621254.663	2634278.689	32.7075250	-108.1220740	333	233	N		
5248-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621255.055	2634428.102	32.7075270	-108.1218900	763	534	N		
5249-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621255.448	2634578.684	32.7075300	-108.1219900	4388	3051	N	Bedrock dust	
5250-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621255.842	2634728.687	32.7075320	-108.1208110	21824	15277	N	Bedrock dust	
5251-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621103.468	2633629.101	32.7070580	-108.1202400	15804	10923	N	Remover <sup>2</sup>	
5252-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621103.858	2633979.084	32.7070630	-108.1235350	15074	922	N	Remover <sup>2</sup>	
5253-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621104.249	2634279.088	32.7071100	-108.1234470	2426	1698	N		
5254-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621104.640	2634579.092	32.7071150	-108.1228720	464	339	N		
5255-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621105.031	2634879.096	32.7071200	-108.1219300	2931	2052	N		
5256-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621105.422	2635179.100	32.7071250	-108.1210600	942	650	N		
5257-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621105.813	2635479.104	32.7071300	-108.1205900	630	454	N		
5258-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621106.204	2635779.108	32.7071350	-108.1201210	880	630	N		
5259-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621106.595	2636079.112	32.7071400	-108.1196510	1652	1156	N		
5260-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621106.986	2636379.116	32.7071450	-108.1191820	1080	746	N		
5271-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		620954.268	2634128.102	32.7068940	-108.1235370	239	171	N		
5272-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		620954.659	2634428.106	32.7069010	-108.1234500	607	425	N		
5273-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		620955.050	2634728.110	32.7069080	-108.1233630	802	581	N		
5274-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		620955.441	2635028.114	32.7069150	-108.1232760	1288	922	N		
5275-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		620955.832	2635328.118	32.7069220	-108.1231890	1359	951	N		
5276-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		620956.223	2635628.122	32.7069290	-108.1231020	2931	2052	N		
5277-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		620956.614	2635928.126	32.7069360	-108.1196310	3997	2798	N	Remover <sup>2</sup>	
5282-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		620805.613	2634729.077	32.7062950	-108.1206040	1389	972	N		
5374-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621845.828	2630838.654	32.7091230	-108.1332840	1126	788	N		
5375-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621846.219	2631128.266	32.7102890	-108.1269310	1102	771	N		
5376-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621846.610	2631428.270	32.7102960	-108.1268440	137	95	N		
5377-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621847.001	2631728.274	32.7103030	-108.1267570	398	279	N		
5378-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2"	0.25 mm sieve		621847.392	2632028.278	32.7103100	-108.1266700	165	116	N		

Footnotes:  
<sup>1</sup> For copper for small grain (sieved by 0.425 mm sieve) and for human health applies to < 2 mm (2000 µm). For FS RAC for human health applies to < 0.425 mm (950 µm). The ratio used to convert copper concentration of 0.425 mm sieve to 0.6<sup>1</sup> sieve is 0.6. For lead and cadmium, the ratio used to convert concentration of 0.425 mm sieve to 0.6<sup>1</sup> sieve is 0.6. For arsenic, the ratio used to convert concentration of 0.425 mm sieve to 0.6<sup>1</sup> sieve is 0.6. For manganese, the ratio used to convert concentration of 0.425 mm sieve to 0.6<sup>1</sup> sieve is 0.6. For iron, the ratio used to convert concentration of 0.425 mm sieve to 0.6<sup>1</sup> sieve is 0.6.  
<sup>2</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRAs.  
<sup>3</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
<sup>4</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NMDOT.  
<sup>5</sup> Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04-11 and U04-12.  
<sup>6</sup> Removed samples with the asterisk are set to 327 for the Thiessen analyses. All other removed samples are completely voided.  
 -- means no data available at the location.

TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0.4'

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1' soil <sup>1</sup>	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	Copper Concentration <sup>1</sup> To Compare to FS RAC (250 µm sieve)	For Bird Pre-Remedial Action (IRA) Dataset	Pre-Interim Remedial Action (IRA) Dataset	Reasons Removed (also see this column's footnotes)
5389-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621102.027	2635002.623	32.7071130	-108.1497200	1588	1112	N	
5381-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620828.941	26344876.047	32.7063350	-108.1201240	1060	742	N	
6001-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		623550.007	2632772.065	32.1183770	-108.1269910	1334	934	N	
6002-NG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		623350.007	2632773.060	32.1183770	-108.1269880	337	236	N	
6075-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621944.352	2630376.326	32.7069410	-108.1347700	593	415	N	
6097-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621944.352	2630376.326	32.7069410	-108.1347670	84	59	N	
6121-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621844.359	2630377.134	32.7091160	-108.1347640	159	111	N	
6199-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621394.732	2630529.297	32.7078820	-108.1342690	1463	1024	N	
6225-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621244.789	2630529.297	32.7074690	-108.1342660	1377	964	N	
6228-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621244.789	2630529.297	32.7074690	-108.1342630	2000	1400	N	
6227-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621244.005	2630228.706	32.7074650	-108.1352410	2730	1911	N	
6253-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621094.412	2630228.710	32.7070530	-108.1352380	275	193	N	
6254-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621094.412	2630228.710	32.7070530	-108.1347510	820	574	N	
6255-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		621094.412	2630228.710	32.7070530	-108.1347510	2152	1506	N	
6265-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620944.399	2630229.483	32.7068430	-108.1352360	128	90	N	
6266-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620944.399	2630229.483	32.7068430	-108.1347480	2609	1826	N	
6267-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620944.399	2630229.483	32.7068430	-108.1347480	1865	1306	N	
6278-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620793.996	2630229.486	32.7062290	-108.1352330	145	102	N	
6279-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620794.406	2630379.801	32.7062300	-108.1347450	926	648	N	
6280-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620794.406	2630379.801	32.7062300	-108.1347450	1720	1204	N	
6283-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620643.984	2630250.289	32.7058180	-108.1352300	132	92	N	
6284-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620644.404	2630380.275	32.7058180	-108.1347430	546	382	N	
6285-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620644.404	2630380.275	32.7058180	-108.1347430	2065	1506	N	
6287-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620494.027	2630250.291	32.7058200	-108.1352280	535	375	N	
6289-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620494.027	2630250.291	32.7058200	-108.1347400	1124	787	N	
6290-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620494.027	2630250.291	32.7058200	-108.1347400	145	102	N	
6291-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620494.027	2630250.291	32.7058200	-108.1347400	623	436	N	
6294-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620344.399	2630381.083	32.7049930	-108.1347370	116	81	N	
6296-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620344.399	2630381.083	32.7049930	-108.1347370	160	112	N	
6301-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620184.398	2630351.459	32.7045930	-108.1347500	276	193	N	
6302-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620184.398	2630351.459	32.7045930	-108.1347500	276	193	N	
6303-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620184.398	2630351.459	32.7045930	-108.1347500	276	193	N	
6304-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620184.398	2630351.459	32.7045930	-108.1347500	276	193	N	
6305-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		620184.398	2630351.459	32.7045930	-108.1347500	276	193	N	
6312-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619844.390	2630352.253	32.7037590	-108.1347290	1474	1032	N	
6313-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619844.390	2630352.253	32.7037590	-108.1347290	2317	1622	N	
6314-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619844.390	2630352.253	32.7037590	-108.1347290	2521	1765	N	
6320-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619744.224	2630352.269	32.7033470	-108.1347270	536	375	N	
6321-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619744.224	2630352.269	32.7033470	-108.1347270	2441	1709	N	
6322-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619744.224	2630352.269	32.7033470	-108.1347270	1475	1033	N	
6328-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619594.421	2630363.042	32.7029320	-108.1347240	2160	1572	N	
6329-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619594.421	2630363.042	32.7029320	-108.1347240	1847	1293	N	
6330-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619594.421	2630363.042	32.7029320	-108.1347240	581	407	N	
6336-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619444.419	2630393.447	32.7025200	-108.1347210	489	342	N	
6337-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619444.419	2630393.447	32.7025200	-108.1347210	1239	867	N	
6338-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619444.419	2630393.447	32.7025200	-108.1347210	2443	1710	N	
6344-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619284.416	2630383.851	32.7021070	-108.1337460	319	223	N	
6345-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619284.416	2630383.851	32.7021070	-108.1337460	1708	1196	N	
6346-EG		Golf Course Interim Removal Action, Arcadis 2009	2008	0-2'	0.25 mm sieve		619284.416	2630383.851	32.7021070	-108.1337460	1442	1009	N	

Footnotes:  
 \* For copper for small ground testing bird applies to 0.4' soil sieved to < 2 mm (200 µm). For FS RAC for human health applies to 0.4' soil sieved to < 0.25 mm (60 µm). The ratio used to convert copper concentration of 0.4' sieved to < 0.25 mm sieved is 0.4' sieved to 0.25 mm sieved is 1.5. If this ratio is not available, it is assumed to be 1.5. If this ratio is not available, it is assumed to be 1.5. If this ratio is not available, it is assumed to be 1.5.  
 \* A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
 \* A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NIMDOT.  
 \* Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
 \* Removed samples with the asterisk are set to 32.7 for the Thiessen analyses. All other removed samples are completely voided.  
 \* - means no data available at the location.



TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0-6"

STSIU Feasibility Study  
Freepport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1' soil <sup>1</sup>	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	Copper Concentration <sup>2</sup> To Compare to HH Pre-FS RAC (250 µm sieve)	For Bird Pre-FS RAC (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
B-025		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620192.866	2634366.663	32.704670	-108.121730	268	215	N		
B-026		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620192.864	2634412.003	32.704680	-108.121730	244	164	N		
B-027		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620192.863	2634459.251	32.704680	-108.121780	459	384	N		
B-028		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620239.171	2634134.859	32.704730	-108.122330	69	49	N		
B-029		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620239.190	2634161.306	32.704730	-108.122380	1529	1411	N		
B-030		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620239.172	2634227.754	32.704730	-108.122310	145	111	N		
B-031		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620239.156	2634273.894	32.704730	-108.122080	3197	3133	N		
B-032		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620239.175	2634320.341	32.704730	-108.121900	1037	928	N		
B-033		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620239.156	2634366.481	32.704730	-108.121800	195	153	N		
B-034		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620239.176	2634412.929	32.704730	-108.121620	1835	1719	N		
B-035		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620239.161	2634449.809	32.704730	-108.121790	1645	1527	N		
B-036		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620239.181	2634505.516	32.704730	-108.121280	651	561	N		
B-037		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620285.405	2634088.545	32.704860	-108.122830	1561	1443	N		
B-038		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620285.464	2634134.865	32.704860	-108.122530	1410	1293	N		
B-039		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620285.467	2634181.432	32.704860	-108.122320	185	144	N		
B-040		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620285.486	2634227.572	32.704860	-108.122320	138	105	N		
B-041		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620285.489	2634320.160	32.704860	-108.121930	3475	3429	N		
B-042		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620285.472	2634366.607	32.704860	-108.121760	3047	2874	N		
B-043		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620285.462	2634412.747	32.704860	-108.121500	2188	2080	N		
B-044		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620285.475	2634459.195	32.704860	-108.121790	2142	2032	N		
B-045		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620285.495	2634505.335	32.704860	-108.121290	1374	1257	N		
B-046		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620331.773	2633718.314	32.704990	-108.123680	1355	1238	N		
B-047		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620331.779	2634088.664	32.704990	-108.122540	1936	1821	N		
B-048		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620331.798	2634135.111	32.704990	-108.122330	3608	3570	N		
B-049		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620331.721	2634181.251	32.704990	-108.122930	3046	2974	N		
B-050		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620331.800	2634227.899	32.704990	-108.122320	1056	955	N		
B-051		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620331.783	2634273.838	32.704990	-108.122620	1036	927	N		
B-052		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620331.768	2634320.288	32.704990	-108.121930	659	593	N		
B-053		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620331.807	2633718.134	32.705110	-108.123680	1651	1535	N		
B-054		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620331.789	2633764.363	32.705110	-108.123460	679	614	N		
B-055		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620378.024	2632052.343	32.705110	-108.125840	1052	942	N		
B-056		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620378.027	2632098.390	32.705110	-108.125840	2151	2082	N		
B-057		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620378.023	2634273.864	32.705110	-108.122820	1632	1716	N		
B-058		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620424.002	2633381.044	32.705240	-108.121930	3160	3094	N		
B-059		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620424.003	2633903.455	32.705240	-108.123680	2406	2304	N		
B-060		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620424.005	2633949.882	32.705240	-108.123360	3633	3597	N		
B-061		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620424.005	2633996.022	32.705240	-108.123190	2206	2098	N		
B-062		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620424.011	2634273.783	32.705240	-108.122860	549	467	N		
B-063		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620424.011	2634320.230	32.705240	-108.122830	3151	3063	N		
B-064		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620424.004	2634320.230	32.705240	-108.121930	423	351	N		
B-065		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620424.013	2634366.678	32.705240	-108.121780	3045	2972	N		
B-066		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620424.037	2634412.817	32.705240	-108.121630	3239	3177	N		
B-067		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620470.897	2633764.527	32.705360	-108.123790	3567	3547	N		
B-068		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620470.915	2633381.075	32.705360	-108.123680	86	65	N		
B-069		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620470.888	2633867.114	32.705360	-108.123480	141	107	N		
B-070		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620470.916	2633903.851	32.705360	-108.123270	627	538	N		

<sup>1</sup> Data were collected for ecological endpoint and were not connected to human health values.  
<sup>2</sup> Data were collected for ecological endpoint and were not connected to human health values.  
<sup>3</sup> Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>4</sup> Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>5</sup> Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>6</sup> Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>7</sup> Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>8</sup> Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>9</sup> Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>10</sup> Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>11</sup> Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two parent samples U04.  
<sup>12</sup> Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two parent samples U04.  
<sup>13</sup> Removed samples with the asterisk are set to 32.7 for the Thiessen analyses. All other removed samples are completely voided.  
<sup>14</sup> -- means no data available at the location.

TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0.6<sup>1</sup>

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1' soil <sup>1</sup>	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	To Compare to FS RAC (250 µm sieve)	Copper Concentration <sup>1</sup> For Bird Pre-Remedial FS RAC Analysis (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
B-090		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620470.707	2633949.001	32.7053680	-108.1213370	914	609	N		
B-091		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620470.707	2634330.566	32.7053710	-108.1213320	1621	1504	N		
B-092		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620470.707	2634366.966	32.7053710	-108.1213320	40	28	N		
B-093		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620470.710	2634412.943	32.7053720	-108.1213310	60	42	N		
B-094		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620470.694	2634449.063	32.7053720	-108.1213320	54	38	N		
B-097		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620470.717	2634588.117	32.7053720	-108.1213320	2612	2518	N		
B-098		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620517.097	2633567.890	32.7053680	-108.1244000	2342	2238	N		
B-105		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620517.022	2634320.175	32.7054980	-108.1213330	538	456	N		
B-106		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620517.024	2634366.622	32.7054980	-108.1213320	517	437	N		
B-107		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620517.024	2634412.762	32.7054980	-108.1213320	180	140	N		
B-108		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620517.007	2634449.209	32.7054980	-108.1214810	2130	2020	N		
B-109		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620516.951	2634505.448	32.7055000	-108.1213310	1888	1773	N		
B-110		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620517.011	2634551.795	32.7055000	-108.1218000	397	329	N		
B-115		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620517.020	2635477.971	32.7055070	-108.1181890	723	628	N		
B-116		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620516.990	2635524.110	32.7055070	-108.1180190	383	316	N		
B-117		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620516.990	2635570.957	32.7055070	-108.1176880	506	427	N		
B-118		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.324	2633671.887	32.7056200	-108.1240410	3427	3377	N		
B-119		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.305	2633718.834	32.7056210	-108.1238900	4208	4217	N		
B-123		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.308	2633903.607	32.7056220	-108.1238880	3485	3439	N		
B-124		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.316	2634273.854	32.7056250	-108.1226840	704	610	N		
B-125		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.329	2634300.301	32.7056250	-108.1213330	2850	2767	N		
B-126		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.319	2634366.440	32.7056280	-108.1217630	77	56	N		
B-127		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.301	2634472.897	32.7056280	-108.1216320	433	360	N		
B-128		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.322	2634493.027	32.7056280	-108.1216320	141	107	N		
B-129		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.305	2634505.474	32.7056270	-108.1213310	105	78	N		
B-130		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.325	2634551.674	32.7056270	-108.1218100	2962	2884	N		
B-132		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.301	2634596.475	32.7056320	-108.1182220	2509	2411	N		
B-133		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.323	2634596.475	32.7056330	-108.1187240	1723	1606	N		
B-134		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.304	2634596.475	32.7056330	-108.1187240	3309	3039	N		
B-137		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.313	2634596.475	32.7056330	-108.1184710	3451	3368	N		
B-138		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.308	2634541.649	32.7056330	-108.1182000	795	646	N		
B-139		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620563.313	2634554.234	32.7056340	-108.1181700	735	640	N		
B-140		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620569.011	2633672.013	32.7057470	-108.1241920	1860	1744	N		
B-141		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.020	2633718.153	32.7057480	-108.1240410	3351	3328	N		
B-142		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.020	2633764.600	32.7057480	-108.1237400	2209	1676	N		
B-145		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.023	2633903.226	32.7057480	-108.1238900	2053	1941	N		
B-146		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.029	2634273.980	32.7057520	-108.1220840	570	468	N		
B-147		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.013	2634320.119	32.7057530	-108.1210340	2441	2340	N		
B-148		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.032	2634366.566	32.7057530	-108.1217830	2165	2056	N		
B-149		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.016	2634412.706	32.7057530	-108.1213330	104	78	N		
B-150		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.035	2634459.153	32.7057540	-108.1214820	1741	1624	N		
B-151		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.018	2634505.600	32.7057540	-108.1213310	1388	1282	N		Bedrock dust
B-152		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.021	2634548.186	32.7057550	-108.1218100	2122	2012	N		
B-153		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620609.021	2634588.186	32.7057550	-108.1218100	2235	2127	N		

Footnotes:  
<sup>1</sup> For copper for small ground (sieving bird) applies to 0.6<sup>1</sup> soil sieved to < 2 mm (2000 µm). For FS RAC for human health applies to 0.1<sup>1</sup> soil sieved to < 0.25 mm (950 µm). The ratio used to convert copper concentration of 0.1<sup>1</sup> sieved to 0.6<sup>1</sup> sieved is < 2 mm sieved to 0.6<sup>1</sup> sieved as < 2 mm sieved is from Appendix A (median ratio of SSS samples of Arcadis 2011). Multiplied copper concentration by 0.7 if outside windblown tailings (and by 1.5 if inside; windblown tailings). If unlisted, multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. The 0.9 and 1.3 are from median ratios of unlisted 0.1<sup>1</sup> soil in 2009 lab reports (at 25 µm in tailings) and 0.6<sup>1</sup> sieved to 0.1<sup>1</sup> sieved (at 25 µm in tailings).  
<sup>2</sup> For lead and cadmium for small ground (sieving bird) applies to 0.6<sup>1</sup> soil sieved to < 2 mm (2000 µm). For FS RAC for human health applies to 0.1<sup>1</sup> soil sieved to < 0.25 mm (950 µm). The ratio used to convert copper concentration of 0.1<sup>1</sup> sieved to 0.6<sup>1</sup> sieved is < 2 mm sieved to 0.6<sup>1</sup> sieved as < 2 mm sieved is from Appendix A (median ratio of SSS samples of Arcadis 2011). Multiplied copper concentration by 0.7 if outside windblown tailings (and by 1.5 if inside; windblown tailings). If unlisted, multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. The 0.9 and 1.3 are from median ratios of unlisted 0.1<sup>1</sup> soil in 2009 lab reports (at 25 µm in tailings) and 0.6<sup>1</sup> sieved to 0.1<sup>1</sup> sieved (at 25 µm in tailings).  
<sup>3</sup> Data were collected for ecological endpoint and were not corrected to human health values.  
<sup>4</sup> Duplicate sample with the same Loc ID has been recorded in Gulf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>5</sup> Duplicate sample with the same Loc ID has been recorded in Gulf Course Interim Removal Action, Arcadis 2009. This sample was marked as MNDOT.  
<sup>6</sup> Duplicate sample with the same Loc ID has been recorded in Gulf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
<sup>7</sup> Duplicate sample with the same Loc ID has been recorded in Gulf Course Interim Removal Action, Arcadis 2009. This sample was marked as MNDOT.  
<sup>8</sup> Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
<sup>9</sup> Removed samples with the asterisk are set to 32.7 for the Theissens analyses. All other removed samples are completely voided.  
 -- means no data available at the location.







TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0.6"

STSIU Feasibility Study  
Freepport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1" soil*	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	To Compare To FS RAC (250 µm sieve)	Copper Concentration <sup>1</sup> For Bird Pre-Analysis (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
B-261		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.628	263374.623	32.7062570	-108.1238920	1880	1765	N		
B-262		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.657	263374.623	-108.1237420	1809	1693	N			
B-263		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.659	263338.039	32.7062580	-108.1234110	1445	1328	N		
B-264		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.658	263367.078	32.7062580	-108.1234110	1678	1561	N		
B-265		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.641	263390.325	32.7062580	-108.1234000	1922	1807	N		
B-266		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.624	263394.664	32.7062590	-108.1231400	3169	3135	N		
B-267		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.625	263402.750	32.7062590	-108.1228900	113	85	N		
B-268		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.625	263402.750	32.7062600	-108.1228900	218	172	N		
B-269		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.644	263408.897	32.7062600	-108.1228880	840	739	N		
B-270		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.631	263430.316	32.7062620	-108.1219350	1289	1174	N		
B-271		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.651	263436.655	32.7062620	-108.1218500	2343	2239	N		
B-272		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.624	263442.902	32.7062620	-108.1216340	978	871	N		
B-273		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.654	263449.841	32.7062630	-108.1214840	1152	1039	N		
B-274		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.657	263450.588	32.7062630	-108.1214330	3489	3443	N		
B-275		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.657	263451.627	32.7062640	-108.1211830	3413	3362	N		
B-276		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.641	263458.074	32.7062640	-108.1210320	2757	2670	N		
B-277		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.656	263468.418	32.7062660	-108.1198280	2196	2087	N		
B-278		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.640	263501.454	32.7062670	-108.1198770	1189	1075	N		
B-279		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.625	263506.103	32.7062670	-108.1195270	1914	1799	N		
B-280		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.645	263510.160	32.7062680	-108.1193760	2227	2119	N		
B-281		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.630	263513.589	32.7062680	-108.1192260	1284	1169	N		
B-282		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.650	263520.058	32.7062680	-108.1190750	2506	2626	N		
B-283		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620794.624	263526.483	32.7062680	-108.1189240	2396	2294	N		
B-284		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.133	263579.447	32.7063850	-108.1245440	1839	1723	N	Removed	Bedrock dust
B-285		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.132	263582.586	32.7063850	-108.1241940	1022	913	N		
B-286		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.132	263587.033	32.7063850	-108.1241940	507	428	N		
B-287		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.134	263578.172	32.7063850	-108.1238920	1010	902	N		
B-288		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.134	263574.619	32.7063850	-108.1237420	990	932	N		
B-289		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.133	263581.078	32.7063850	-108.1235920	892	788	N		
B-290		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.133	263581.078	32.7063850	-108.1235920	691	587	N		
B-291		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.133	263581.078	32.7063850	-108.1235920	451	347	N		
B-292		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.137	263594.931	32.7063850	-108.1231400	1690	1672	N		
B-293		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.136	263596.327	32.7063870	-108.1229880	1273	1149	N		
B-294		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.139	263402.371	32.7063870	-108.1226380	1231	1117	N		
B-295		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.137	263408.823	32.7063870	-108.1226880	2403	2301	N		
B-296		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.141	263414.862	32.7063870	-108.1223870	912	807	N		
B-297		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.161	263411.069	32.7063880	-108.1223870	1224	1109	N		
B-298		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.146	263420.134	32.7063880	-108.1219350	562	497	N		
B-299		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.128	263436.581	32.7063880	-108.1218500	990	882	N		
B-300		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.148	263442.720	32.7063900	-108.1216350	934	828	N		
B-301		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.131	263449.167	32.7063900	-108.1214840	944	838	N		
B-302		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.131	263459.306	32.7063900	-108.1213340	1179	1065	N		
B-303		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.134	263455.753	32.7063910	-108.1211830	370	305	N		
B-304		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.134	263468.543	32.7063940	-108.1196280	228	180	N		
B-305		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.154	263501.682	32.7063940	-108.1196270	2436	2335	N		
B-306		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.138	263506.128	32.7063940	-108.1196270	2948	2880	N		
B-308		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.143	263513.714	32.7063950	-108.1192260	1741	1624	N		

Footnotes:  
 \* For copper for small ground (testing) bird applies to 0.6" soil sieved to < 2 mm (2000 µm). For FS RAC for human health applies to 0.1" soil sieved to < 0.25 mm (950 µm). The ratio used to convert copper concentration of 0.1" sieved to 0.6" sieved is 0.6. For lead and cadmium for human health applies to 0.1" soil sieved to < 2 mm (2000 µm). The ratio used to convert lead concentration from 0.1" sieved to 0.6" sieved is 0.6. For lead and cadmium for human health applies to 0.1" soil sieved to < 2 mm (2000 µm). The ratio used to convert lead concentration from 0.1" sieved to 0.6" sieved is 0.6.  
 1. If field; windblown tailings) If unsorted, multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. This: 0.9 and 1.3 are: from median ratios of unsorted 0.1" soil in 2009 lab reports (at 25 cm in tailings) and the data for the 2013 monitoring report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (P-D 8, 18, 20).  
 2. If field; windblown tailings) If unsorted, multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. This: 0.9 and 1.3 are: from median ratios of unsorted 0.1" soil in 2009 lab reports (at 25 cm in tailings) and the data for the 2013 monitoring report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (P-D 8, 18, 20).  
 3. Data were collected for ecological endpoint and were not connected to human health values.  
 4. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 5. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 6. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 7. Duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NIMDOT.  
 8. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two parent samples U04.  
 9. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two parent samples U04.  
 10. Removed samples with the asterisk are set to 32.7 for the Theissen's analyses. All other removed samples are completely voided.  
 -- means no data available at the location.

TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0-6"

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1' soil*	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	Copper Concentration <sup>1</sup> To Compare to HH Pre-Remedial FS RAC (250 µm sieve)	For Bird Pre-Remedial FS RAC Analysis (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
B-309		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.663	2635200.161	32.7063560	-108.1192750	1690	1573	N		
B-310		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620841.148	2635246.300	32.7063380	-108.1182920	3304	3342	N		
B-311		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.447	2633579.267	32.7065110	-108.1241940	1588	1481	N		
B-312		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.465	2633625.713	32.7065110	-108.1241940	1588	1481	N		
B-313		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.466	2633671.852	32.7065120	-108.1241940	1665	1667	N		
B-314		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.466	2633718.259	32.7065120	-108.1241940	2172	2062	N		
B-315		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.467	2633764.438	32.7065120	-108.1241940	766	669	N		
B-316		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.467	2633810.895	32.7065120	-108.1241940	2350	2247	N		
B-317		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.468	2633857.024	32.7065130	-108.1241940	1110	999	N		
B-318		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.468	2633903.471	32.7065130	-108.1241940	2684	2593	N		
B-319		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.469	2633949.817	32.7065140	-108.1241940	1753	1636	N		
B-320		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.470	2633996.056	32.7065140	-108.1241940	2828	2743	N		
B-321		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.471	2634042.503	32.7065140	-108.1241940	3123	3065	N		
B-322		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.472	2634088.842	32.7065140	-108.1241940	3127	3059	N		
B-323		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.473	2634135.089	32.7065150	-108.1241940	336	275	N		
B-324		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.474	2634178.228	32.7065150	-108.1241940	3046	2974	N		
B-325		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.475	2634224.574	32.7065160	-108.1241940	3371	3318	N		
B-326		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.476	2634270.920	32.7065160	-108.1241940	1323	1207	N		
B-327		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.477	2634316.266	32.7065160	-108.1241940	1962	1848	N		
B-328		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.478	2634362.612	32.7065170	-108.1241940	1962	1848	N		
B-329		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.479	2634408.958	32.7065170	-108.1241940	2944	2865	N		
B-330		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.480	2634455.304	32.7065170	-108.1241940	148	113	N		
B-331		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.481	2634501.650	32.7065170	-108.1241940	2242	2134	N		
B-332		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620887.482	2634547.996	32.7065170	-108.1241940	1882	1778	N		
B-333		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620933.721	2633671.979	32.7063380	-108.1241940	1140	1028	N		
B-334		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620933.722	2633718.118	32.7063380	-108.1241940	1784	1668	N		
B-335		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620933.723	2633764.257	32.7063380	-108.1241940	799	701	N		
B-336		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620933.724	2633810.396	32.7063380	-108.1241940	658	597	N		
B-337		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620933.725	2633856.535	32.7063380	-108.1241940	405	338	N		
B-338		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620933.726	2633902.674	32.7063380	-108.1241940	730	635	N		
B-339		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620933.727	2633948.813	32.7063380	-108.1241940	3245	2922	N		
B-340		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620933.728	2633994.952	32.7063380	-108.1241940	2659	2302	N		
B-341		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620933.729	2634041.091	32.7063380	-108.1241940	2658	2568	N		
B-342		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620933.730	2634087.230	32.7063380	-108.1241940	3356	3301	N		
B-343		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.370	263352.864	32.7067770	-108.1492770	344	281	N		
B-344		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.371	2633579.003	32.7067770	-108.1492770	929	824	N		
B-345		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.372	2633635.142	32.7067770	-108.1492770	929	824	N		
B-346		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.373	2633691.281	32.7067770	-108.1492770	1571	1453	N		
B-347		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.374	2633747.420	32.7067770	-108.1492770	1245	1131	N		
B-348		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.375	2633803.559	32.7067770	-108.1492770	1869	1784	N		
B-349		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.376	2633859.698	32.7067770	-108.1492770	743	647	N		
B-350		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.377	2633915.837	32.7067770	-108.1492770	2215	2107	N		
B-351		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.378	2633971.976	32.7067770	-108.1492770	313	254	N		
B-352		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.379	2634028.115	32.7067770	-108.1492770	2450	2350	N		
B-353		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.380	2634084.254	32.7067770	-108.1492770	2949	2871	N		
B-354		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.381	2634140.393	32.7067770	-108.1492770	1874	1758	N		
B-355		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621026.382	2634196.532	32.7067770	-108.1492770	746	649	N		

Footnotes:  
 \* For copper for small ground (testing) bird applies to 0-6" soil sieved to < 2 mm (2000 µm). For FS RAC for human health applies to 0-1' soil sieved to < 0.25 mm (620 µm). The ratio used to convert copper concentration of 0-1' sieved to 0-6" sieved is 0.6.  
 1. If fish: windblown tailings) If unmined, multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. The 0.9 and 1.3 are from median ratios of unmined 0-1' soil in 2009 lab reports (at 25 µm in tailings) compared to sieved 0-6" co-located soil in Drexler lab report in Appendix D or the pH Monitoring Report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (P-D 8, 18, 20). In tailings, the data are based on 0.6 inch sieves.  
 2. Data were collected for ecological endpoint and were not corrected to human health values.  
 3. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 4. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as MMDOT.  
 5. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
 6. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as MMDOT.  
 7. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
 8. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as MMDOT.  
 9. Removed samples with the asterisk are set to 327 for the Theissens analyses. All other removed samples are completely voided.  
 -- means no data available at the location.

TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0.6'

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1' soil <sup>1</sup>	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	Copper Concentration <sup>1</sup> To Compare to HH Pre-Action FS RAC (250 µm sieve)	For Bird Pre-Action FS RAC (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
E-356		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621072.671	2635061.138	32.70703310	-108.1952930	1066	1072	N	Removed	Bedrock dust
E-357		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621072.672	2635071.277	32.70703310	-108.1952930	1186	985	N		
E-358		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621118.997	2633579.287	32.7071470	-108.1244970	808	708	N		
E-359		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621118.998	2633579.287	32.7071470	-108.1244970	171	132	N		
E-360		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621118.999	2633579.287	32.7071470	-108.1244970	3146	3079	N		
E-361		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621118.999	2633671.872	32.7071480	-108.1244660	2029	1916	N		
E-362		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621118.984	2635054.817	32.7071580	-108.1968290	447	373	N	Removed	Bedrock dust
E-363		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621118.984	2635054.817	32.7071580	-108.1968290	1095	984	N		
E-364		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621118.969	2635107.402	32.7071580	-108.1957790	1395	1278	N		
E-365		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621165.278	2635054.442	32.7072850	-108.1968800	1540	1422	N	Removed	Bedrock dust
E-366		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621165.289	2635061.081	32.7072850	-108.1968300	1575	1458	N		
E-367		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621165.282	2635107.827	32.7072860	-108.1957790	1142	1030	N		
E-368		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621211.607	2634968.821	32.7074120	-108.1968310	869	767	N	Removed	Bedrock dust
E-369		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621211.602	2635054.759	32.7074120	-108.1968310	728	632	N		
E-370		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621211.576	2634922.306	32.7074130	-108.1968300	476	399	N		
E-371		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621257.001	2634922.306	32.7075390	-108.1968300	2916	2836	N	Removed	Bedrock dust
E-372		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621257.005	2634968.438	32.7075390	-108.1968320	3287	3228	N	Removed	Bedrock dust
E-373		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621257.005	2635014.684	32.7075400	-108.1968310	391	323	N		
E-374		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621257.005	2635061.023	32.7075400	-108.1968310	350	287	N		
E-375		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621304.215	2634922.117	32.7076680	-108.1968300	2670	2579	N	Removed	Bedrock dust
E-376		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621304.199	2634968.563	32.7076670	-108.1968320	1352	1236	N		
E-377		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621304.220	2635054.702	32.7076670	-108.1968300	1386	1240	N		
E-378		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621350.829	2634875.797	32.7077930	-108.1201340	779	881	N	Removed	Bedrock dust
E-379		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621350.829	2634922.542	32.7077930	-108.1968300	1567	1449	N		
E-380		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621350.813	2634968.381	32.7077930	-108.1968300	2808	2723	N	Removed	Bedrock dust
E-381		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621386.829	2634829.476	32.7079200	-108.1201650	2079	1968	N		
E-382		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621386.829	2634873.822	32.7079200	-108.1201340	2403	2301	N		
E-383		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621386.807	2634922.069	32.7079210	-108.1958440	730	635	N		
E-384		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621443.132	2634973.051	32.7080470	-108.1204360	2482	2392	N	Removed	Bedrock dust
E-385		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621443.132	2634973.051	32.7080470	-108.1204360	2482	2392	N		
E-386		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621443.132	2634973.051	32.7080470	-108.1204360	2482	2392	N		
E-387		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621443.132	2634973.051	32.7080470	-108.1204360	2482	2392	N		
E-388		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621443.132	2634973.051	32.7080470	-108.1204360	2482	2392	N		
E-389		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621443.132	2634973.051	32.7080470	-108.1204360	2482	2392	N		
E-390		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621443.132	2634973.051	32.7080470	-108.1204360	2482	2392	N		
E-391		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621322.823	2634931.241	32.7077120	-108.1201030	6292	6514	N	Removed	Bedrock dust
E-392		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621322.823	2634931.241	32.7077120	-108.1201030	6292	6514	N		
E-393		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621322.823	2634931.241	32.7077120	-108.1201030	6292	6514	N		
E-394		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621322.823	2634931.241	32.7077120	-108.1201030	6292	6514	N		
E-395		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621322.823	2634931.241	32.7077120	-108.1201030	6292	6514	N		
E-396		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621322.823	2634931.241	32.7077120	-108.1201030	6292	6514	N		
E-397		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620336.825	2633650.134	32.7049980	-108.1241100	1365	1249	N		
E-398		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620336.825	2633650.134	32.7049980	-108.1241100	1365	1249	N		
E-399		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620336.825	2633650.134	32.7049980	-108.1241100	1365	1249	N		
E-400		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620437.849	2633663.305	32.7052750	-108.1241330	1043	933	N		
E-401		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620437.849	2633663.305	32.7052750	-108.1241330	1043	933	N		
E-402		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620539.809	2633596.889	32.7054140	-108.1242220	270	216	N		
E-403		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620539.809	2633596.889	32.7054140	-108.1242220	270	216	N		
E-403		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620635.805	2633602.472	32.7057000	-108.1242880	771	673	N		
E-403		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620635.805	2633602.472	32.7057000	-108.1242880	1029	920	N		

**Footnotes:**  
<sup>1</sup> For copper for small ground (sieving) bird applies to 0.6' soil sieved to < 2 mm (2000 µm). For FS RAC for human health applies to 0.1' soil sieved to < 0.25 mm (625 µm). The ratio used to convert copper concentration of 0.1' sieved to 0.6' sieved is < 2 mm sieved is 0.6' sieved is 0.6'.  
<sup>2</sup> If unlisted, multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. The 0.9 and 1.3 are from median ratios of unlisted 0.1' soil in 2009 lab reports (at 25 µm in tailings) compared to sieved 0.6' co-located soil in Dreier lab report in Appendix D or the pH Monitoring Report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (P-D 8, 18, 20).  
<sup>3</sup> Data were collected for ecological endpoint and were not connected to human health values.  
<sup>4</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>5</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
<sup>6</sup> A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NMDOT.  
<sup>7</sup> Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
<sup>8</sup> Removed samples with the asterisk are set to 32.7 for the Thiessens analyses. All other removed samples are completely voided.  
<sup>9</sup> - means no data available at the location.

TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0-6"

STSIU Feasibility Study  
Freepport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1' soil'	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	Copper Concentration <sup>1</sup> To Compare to HH Pre-Remedial Action (IRA) Dataset	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
B-404		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620667.958	2633561.688	32.7059620	-108.1242700	284	228		
B-405		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620669.037	2633563.066	32.7059680	-108.1242950	2846	2763		
B-406		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620732.255	2633553.743	32.7060850	-108.1244320	2060	1973		
B-407		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620784.220	2633553.883	32.7062260	-108.1244270	600	513		
B-408		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620832.200	2633555.554	32.7062860	-108.1244200	3576	3536		
B-409		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620879.238	2633551.316	32.7064890	-108.1244360	1913	1798		
B-410		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620930.245	2633504.818	32.7066300	-108.1248770	759	662		
B-411		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620980.276	2633502.856	32.7067650	-108.1249570	1023	914		
B-412		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621032.868	2633505.013	32.7069090	-108.1248880	915	810		
B-413		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621084.867	2633756.541	32.7051310	-108.1237640	249	199		
B-414		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621130.284	2633460.188	32.7071770	-108.1247340	155	119		
B-415		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620881.286	2634282.379	32.7064990	-108.1220590	1841	1725		
B-416		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620833.710	2634285.079	32.7063690	-108.1220500	106	79		
B-417		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620784.444	2634286.852	32.7062300	-108.1220440	1928	1813		
B-418		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620384.474	2634006.857	32.7051330	-108.1229510	2599	2504		
BX-038		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620238.544	2634008.037	32.7047310	-108.1226530	436	364		
BX-063		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620553.764	2634639.229	32.7055490	-108.1208960	1249	1134		
BX-065		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620594.664	2634685.323	32.7056870	-108.1207460	2288	2182		
BX-066		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620594.652	2634763.723	32.7056980	-108.1204270	1441	1324		
BX-070		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620634.053	2634980.802	32.7058250	-108.1197670	1731	1614		
BX-071		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620634.066	2635029.718	32.7058250	-108.1196270	2845	2761		
BX-077		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620535.045	2635275.714	32.7058580	-108.1186270	500	421		
BX-079		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620582.350	2635274.920	32.7058270	-108.1186680	50	35		
BX-087		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620582.350	2635475.925	32.7058880	-108.1181770	1109	897		
BX-088		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621222.361	2635131.239	32.7071680	-108.1193020	1721	1604		
BX-111		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621222.361	2635493.446	32.7074420	-108.1198420	2352	2233		
BX-116		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621074.630	2635032.866	32.7070370	-108.1196710	2364	2276		
BX-120		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620978.807	2635073.869	32.7067720	-108.1196710	2236	2158		
BX-121		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620927.343	2635023.031	32.7066320	-108.1196600	3507	3463		
BX-142		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620928.659	2635128.030	32.7066320	-108.1196600	4126	4019		
BX-144		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		620928.659	2635128.030	32.7066320	-108.1196600	246	207		
BX-169		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621155.230	2633455.236	32.7073070	-108.1245150	146	106		
BX-171		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621224.663	2633455.236	32.7073070	-108.1245150	2548	2451		
BX-172		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621273.841	2633409.107	32.7075710	-108.1245110	1916	1802		
BX-175		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621320.865	2633409.221	32.7077010	-108.1245020	233	185		
BX-188		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621421.267	2633336.930	32.7079780	-108.1257330	1132	1039		
BX-195		B Ranch Interim Removal Action, Arcadis 2020	2019	shallow	2 mm sieve		621468.555	2633360.044	32.7081060	-108.1256330	862	759		
RR001		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617981.038	2630116.784	32.6984950	-108.1355750	2,282	1597		
RR004		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618115.538	2630117.547	32.6988650	-108.1355740	2,366	1656		
RR005		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618112.973	2629982.856	32.6985770	-108.1360110	2,076	1453		
RR006		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618246.990	2630000.103	32.6992250	-108.1356570	2,156	1509		
RR007		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618246.990	2630114.506	32.6992250	-108.1356830	1,078	755		
RR008		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618245.817	2630364.198	32.6992250	-108.1347730	3,329	2330		
RR009		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618377.344	2630368.266	32.6995870	-108.1347930	3,777	2644		
RR010		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618378.035	2630115.557	32.6995870	-108.1358200	1,741	1219		
RR011		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618509.225	2630116.177	32.6998470	-108.1358280	1,531	1077		
RR012		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618511.207	2630378.672	32.6999550	-108.1347280	2,839	1962		

Footnotes:  
 1. For copper for small ground testing (not applied to 0.6" soil sieved to < 2 mm (200 µm)), the FS RAC for human health applies to 0.1" soil sieved to < 0.25 mm (60 µm). The ratio used to convert copper concentration of 0.1" sieved to 0.6" sieved is < 2 mm sieved is < 2 mm sieved is from Appendix A (median ratio of SSS samples of Arcadis 2011). Multiplied copper concentration by 0.7 if outside windblown tailings (and by 1.5 if inside windblown tailings). If unlisted, multiplied by 0.9 if outside windblown tailings and by 1.3 if inside windblown tailings. The 0.9 and 1.3 are from median ratios of unlisted 0.1" soil in 2009 lab reports (at 25 m on transect) compared to sieved 0.6" co-located soil in Dreier lab report in Appendix D or the pH Monitoring Report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (P-D 8, 18, 20). In tailings, the data are the same as for the 0.6" sieved soil.  
 2. Data were collected for ecological endpoint and were not corrected to human health values.  
 3. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 4. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 5. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
 6. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NMDOT.  
 7. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
 8. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
 9. Removed samples with the asterisk are set to 32.7 for the Theissen's analyses. All other removed samples are completely voided.  
 - means no data available at the location.

TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0.6'

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1' soil'	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	Copper Concentration <sup>1</sup> To Compare to HH Pre-Remediation FS RAC (250 µm sieve)	For Bird Pre-Remediation FS RAC (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
RR-012		Razorback Ridge Interim Removal Action, Golder 2015	2014	0-1"	2 mm sieve		619442.061	2637954.668	32.7025700	-108.1101100	176	158	N		
RR-013		Razorback Ridge Interim Removal Action, Golder 2015	2014	0-1"	2 mm sieve		619441.768	2638074.655	32.7025700	-108.1097200	50	45	N		
RR-014		Railroad Interim Removal Action, Golder 2015	2012	0-2"	0.25 mm sieve		618640.416	2638116.828	32.7025700	-108.1356800	1,476	1,033	N		
RR-015		Razorback Ridge Interim Removal Action, Golder 2015	2014	0-1"	2 mm sieve		619441.656	2638194.021	32.7025700	-108.1093300	1,235	1,112	N		
RR-016		Razorback Ridge Interim Removal Action, Golder 2015	2014	0-1"	2 mm sieve		619323.594	2637353.944	32.7022400	-108.1126800	123	111	N		
RR-017		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619323.279	2637473.910	32.7022400	-108.1167000	141	127	N		
RR-018		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617717.894	2639099.329	32.6977180	-108.1329970	544	381	N		
RR-019		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617718.638	2639774.865	32.6977180	-108.1334300	1,407	985	N		
RR-020		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619322.328	2637836.855	32.7022400	-108.1104900	390	351	N		
RR-021		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617849.828	2639775.533	32.6984400	-108.1334300	1,008	706	N		
RR-022		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619322.015	2637966.623	32.7022400	-108.1101000	989	890	N		
RR-023		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619321.702	2638006.618	32.7022400	-108.1097100	61	55	N		
RR-024		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617941.018	2639776.181	32.6985000	-108.1334310	140	98	N		
RR-025		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619202.888	2637596.638	32.7019100	-108.1112700	303	273	N		
RR-026		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618115.519	2639776.600	32.6988700	-108.1334300	354	255	N		
RR-027		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619202.584	2637716.605	32.7019100	-108.1108800	1,578	1,420	N		
RR-028		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618114.657	2639098.698	32.6988690	-108.1330400	387	271	N		
RR-029		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618246.709	2639777.548	32.6992310	-108.1334290	220	154	N		
RR-030		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619201.957	2637956.538	32.7019100	-108.1101000	1,285	1,139	N		
RR-031		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618377.878	2639676.697	32.6989500	-108.1337590	237	166	N		
RR-032		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619201.644	2638076.995	32.7019100	-108.1097700	465	419	N		
RR-033		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619508.269	2639684.072	32.6999500	-108.1338010	311	218	N		
RR-034		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619630.729	2637716.291	32.7003130	-108.1336200	470	329	N		
RR-035		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619322.525	2637716.291	32.7015800	-108.1108800	225	21	N		
RR-036		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618775.326	2639645.729	32.7006830	-108.1336830	1,101	771	N		
RR-037		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619322.272	2637636.258	32.7015800	-108.1109600	973	916	N		
RR-038		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618985.978	2639642.698	32.7010170	-108.1337300	272	190	N		
RR-039		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619401.869	2639777.229	32.7019100	-108.1101000	1,189	848	N		
RR-040		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619401.869	2639777.229	32.7019100	-108.1101000	1,211	848	N		
RR-041		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618112.947	2638076.193	32.7015800	-108.1334500	262	246	N		
RR-042		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		619401.123	2638106.126	32.6988650	-108.1324980	377	264	N		
RR-043		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618244.458	2631072.331	32.6992270	-108.1324700	991	1181	N		
RR-044		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618379.322	2631060.347	32.6995270	-108.1324700	5,623	3836	N		RR Depot
RR-045		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618962.467	2637715.676	32.7012500	-108.1106900	697	627	N		Removed*
RR-046		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618962.153	2637635.844	32.7012500	-108.1106900	1,08	87	N		
RR-047		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618510.452	2631172.536	32.6995950	-108.1324480	1,148	804	N		
RR-048		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618961.540	2637955.912	32.7012500	-108.1101000	851	766	N		
RR-049		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618641.842	2631173.183	32.7003190	-108.1324770	4,338	3,037	N		
RR-050		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618961.527	2638075.879	32.7012500	-108.1097100	590	531	N		
RR-051		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618772.951	2631170.538	32.7006800	-108.1324560	1,616	1,131	N		
RR-052		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618775.483	2631305.995	32.7006800	-108.1317190	6,041	4,229	N		Removed*
RR-053		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618842.009	2637715.662	32.7009200	-108.1317180	464	436	N		
RR-054		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618906.873	2631305.741	32.7010490	-108.1317180	3,290	2,303	N		
RR-055		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		618842.005	2637635.630	32.7009200	-108.1104900	79	71	N		

Footnotes:  
 1. For copper for small ground (sieving) bird applies to 0.6' soil sieved to < 2 mm (2000 µm). For FS RAC for human health applies to 0.1' soil sieved to < 0.25 mm (950 µm). The ratio used to convert copper concentration of 0.1' sieved to 0.6' sieved is 0.6.  
 2. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 3. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as RR Depot.  
 4. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
 5. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NMDOT.  
 6. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
 7. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
 8. Removed samples with the asterisk are set to 32.7 for the Theissen's analyses. All other removed samples are completely voided.  
 9. - means no data available at the location.







TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0-6"

STSIU Feasibility Study  
Freeport-McMoran Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1" soil'	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	Copper Concentration <sup>1</sup> To Compare to HH Pre-Action FS RAC (250 µm sieve)	For Bird Pre-Action FS RAC (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Removed in Post-IRA Dataset	Reason Removed (also see this column's footnotes)
RR-109		Razorback Ridge Interim Removal Action, Golder 2015	2014	0-1"	2 mm sieve		617280.91	26381.94	32.9566300	-108.1031100	302	272	N		
RR-110		Razorback Ridge Interim Removal Action, Golder 2015	2014	0-1"	2 mm sieve		617283.17	26383.14	32.9566400	-108.1032000	311	280	N		
RR-111		Razorback Ridge Interim Removal Action, Golder 2015	2014	0-1"	2 mm sieve		617283.06	26384.34	32.9566400	-108.1033000	538	484	N		
RR-112		Razorback Ridge Interim Removal Action, Golder 2015	2014	0-1"	2 mm sieve		620491.235	262995.54	32.9566400	-108.1081400	49	44	N		
RR-113		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617160.333	262915.237	32.9563000	-108.1362600	3,557	2,490	N		
RR-114		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		620754.400	262965.154	32.7061160	-108.1364610	1,804	1,263	N		
RR-115		Razorback Ridge Interim Removal Action, Golder 2015	2014	0-1"	2 mm sieve		620755.030	262970.266	32.7061170	-108.1365880	2,490	1,743	N		
RR-116		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617163.347	262943.455	32.9563100	-108.1083300	986	887	N		
RR-117		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		620886.201	262971.388	32.7064770	-108.1368870	1,752	1,226	N		
RR-118		Razorback Ridge Interim Removal Action, Golder 2015	2014	0-1"	2 mm sieve		617163.036	262954.169	32.9563100	-108.1081400	36	32	N		
RR-119		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		620886.069	262956.840	32.7064790	-108.1373240	3,156	2,209	N		
RR-120		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621281.570	262987.355	32.7075660	-108.1363940	359	251	N		
RR-121		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617040.274	262913.925	32.9569700	-108.1093100	219	197	N		
RR-122		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621413.316	262997.627	32.7079280	-108.1368200	3,410	2,387	N		
RR-123		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617043.001	262983.133	32.9569900	-108.1089200	146	131	N		
RR-124		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621413.745	262985.193	32.7079290	-108.1364670	2,626	1,638	N		
RR-125		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617043.289	262984.533	32.9569900	-108.1095300	577	519	N		
RR-126		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621544.306	262996.629	32.7082880	-108.1366400	1,845	1,292	N		
RR-127		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		617042.978	262953.559	32.9569900	-108.1091400	277	249	N		
RR-128		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621414.411	262959.592	32.7079290	-108.1368940	2,263	1,594	N		
RR-129		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616920.268	262919.689	32.9564400	-108.1372270	1,520	1,064	N		
RR-130		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621490.164	262940.149	32.7087360	-108.1376440	1,845	1,292	N		
RR-131		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616923.534	262933.673	32.9565500	-108.1085100	167	168	N		
RR-132		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621565.221	262931.355	32.7084900	-108.1347480	973	881	N		
RR-133		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616923.223	262943.049	32.9565500	-108.1085200	426	383	N		
RR-134		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621590.022	262951.382	32.7084200	-108.1341900	1,235	895	N		
RR-135		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621672.046	262956.250	32.9565500	-108.1091900	1,033	760	N		
RR-136		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621619.897	262956.563	32.7086620	-108.1344200	676	472	N		
RR-137		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616800.190	262916.376	32.9563100	-108.1093000	227	204	N		
RR-138		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621678.124	262977.540	32.7086620	-108.1346900	1,297	908	N		
RR-139		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616803.476	262933.166	32.9563200	-108.1081000	460	414	N		
RR-140		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621809.314	262975.788	32.7080200	-108.1334680	1,927	1,349	N		
RR-141		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616803.165	262946.337	32.9563200	-108.1085200	497	447	N		
RR-142		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621809.658	262954.416	32.7080200	-108.1338620	191	134	N		
RR-143		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616802.654	262956.311	32.9563200	-108.1081300	134	121	N		
RR-144		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621810.604	262955.391	32.7090200	-108.1342100	529	370	N		
RR-145		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616802.643	262967.288	32.9563200	-108.1077400	539	485	N		
RR-146		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621941.913	262950.749	32.7093850	-108.1343100	381	297	N		
RR-147		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616663.418	262933.649	32.6949900	-108.1089100	38	34	N		
RR-148		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		621941.168	262956.223	32.7093840	-108.1338930	117	82	N		
RR-149		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616663.061	262943.025	32.6949900	-108.1085200	334	301	N		
RR-150		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		622072.477	262964.821	32.7097450	-108.1339030	168	118	N		
RR-151		Railroad Interim Removal Action, Golder 2013	2012	0-2"	0.25 mm sieve		616662.705	262956.027	32.6949900	-108.1081300	558	502	N		

Footnotes:  
 1. For copper for small ground testing bird applies to 0-6" soil sieved to < 2 mm (2000 µm). For FS RAC for human health applies to 0-1" soil sieved to < 0.25 mm (625 µm). The ratio used to convert copper concentration of 0-1" sieved to 0-6" sieved is 1.5. For lead and zinc, the ratio used to convert copper concentration of 0-1" sieved to 0-6" sieved is 1.5. For lead and zinc, the ratio used to convert copper concentration by 1.5 if inside windblown tailings and by 1.3 if outside windblown tailings. The 0.9 and 1.3 are from median ratios of unexcavated 0-1" soil in 2009 lab reports (at 25 m on transect) compared to sieved 0-6" co-located soil in Dreher lab report in Appendix D or the pH Monitoring Report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (P-D 8, 18, 20). In tailings, the data are from the 0-6 inch soil.  
 2. Data were collected for ecological endpoint and were not corrected to human health values.  
 3. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as IRA.  
 4. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as P-160.  
 5. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as P-161.  
 6. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as Volcanic Rock.  
 7. A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Removal Action, Arcadis 2009. This sample was marked as NMDOT.  
 8. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
 9. Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two located samples U04.  
 10. Removed samples with the asterisk are set to 32.7 for the Theissen's analyses. All other removed samples are completely voided.  
 11. -- means no data available at the location.











TABLE 3-2  
STSIU Surface Soil Copper Results Adjusted to 0-6"

STSIU Feasibility Study  
Freeport-McMoRan Chino Mines Company  
Vanadium, New Mexico

Location ID	Alternate ID	Source Report	Year Collected	Original Soil Depth	Sieving Status	1 if Windblown Tailing (WBT) adjustment needed for 0-1" soil'	Northing (State Plane)	Easting (State Plane)	Latitude	Longitude	To Compare To HH Pre FS RAC (250 µm sieve)	Copper Concentration <sup>1</sup> For Bird Pre-Remedial Action (IRA) Analysis (2000 µm sieve)	Pre-Interim Remedial Action (IRA) Dataset	Reason Removed (also see this column's footnotes)
STSIU-2011-7		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		617485.31	2639213.323	32.6872	-108.1060	--	529	N	
STSIU-2011-8		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		622319.895	2642948.007	32.7103	-108.0939	--	287	N	
STSIU-2011-9		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		617007.041	2641057.872	32.6959	-108.1000	--	560	N	
STSIU-2011-10	STS-PT-2013-29	FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		609319.048	2645961.777	32.6748	-108.0940	--	96	N	July 2013. Alternative Oct. 2013 sample = 234.
STSIU-2011-11		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		609298.573	2643500.091	32.6747	-108.0920	--	216	N	
STSIU-2011-12		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		605464.845	2645029.462	32.6642	-108.0870	--	316	N	
STSIU-2011-13		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		610054.863	2642886.566	32.6768	-108.0940	--	305	N	July 2013. Alternative Oct. 2013 sample = 152.
STSIU-2011-14		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		621457.331	2638455.245	32.7081	-108.1150	--	1640	N	July 2013. Alternative Oct. 2013 sample = 153.
STSIU-2011-15		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		621860.589	2635533.562	32.7092	-108.1180	--	1640	N	
STSIU-2011-16		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		620234.14	2645679.97	32.7048	-108.0950	--	395	N	
STSIU-2011-17		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		609837.313	2642270.639	32.6762	-108.0960	--	654	N	
STSIU-2011-18		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		612649.306	2650699.742	32.6940	-108.0677	--	213	N	
STSIU-2012-B1		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		609299.201	2653329.641	32.6748	-108.0601	--	288	N	July 2013. Alternative Oct. 2013 sample = 164.
STSIU-2012-B2		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		608053.458	2658515.91	32.6744	-108.0445	--	182	N	July 2013. Alternative Oct. 2013 sample = 248.
STSIU-2012-B3		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		608040.451	2658905.776	32.6744	-108.0423	--	344	N	July 2013. Alternative Oct. 2013 sample = 248.
STSIU-2012-B3		FS Data Collection, submitted herein	2013	0-6"	2 mm sieve (g)		608925.052	2657963.466	32.6738	-108.0449	--	161	N	July 2013. Alternative Oct. 2013 sample = 253.

**References:**

- Arctadis, 2010. AOC Terrestrial Invertebrate Copper Bioaccumulation and Bioavailability Study, Smelter/Tailing Soils Investigation Unit. Prepared for Freeport-McMoRan Chino Mines Company, Vanadium, New Mexico. (composite of 15 samples in 100-m radius plot).
- Arctadis, 2011. Administrative Order on Consent Feasibility Study Proposal, Smelter Tailings Soils Investigation Unit. Prepared for Chino Mines Company, Hurley, New Mexico, October.
- Chino, 1995. Administrative Order on Consent, Remedial Investigation Background Report, Chino Mines Investigation Area. Prepared by Chino Mines Company, Hurley, New Mexico. Department received 28 October 1995. (one grab sample).
- Newfields, 2005. Chino Mines Administrative Order on Consent Site-wide Ecological Risk Assessment. Prepared for Chino Mines Company, November. (average of 3 samples on 50-m transect).
- SRK, Inc., 2008. Chino Mines Administrative Order on Consent, Revised Remedial Investigation Report, Smelter/Tailing Soils Investigation Unit - Revision 2. February 6. (method was composite of 6 random subsamples in 100' x 100' area)
- Arctadis, 2009. STSIU Interim Remedial Action, Completion Report (aka Golf Course IRA), March 10.
- Golder Associates, 2013. Supplemental Completion Report, Interim Remedial Action, STSIU (aka Railroad IRA), May 28.
- Arctadis, 2017. Year 5 Monitoring Report for Smelter/Tailing Soils Investigation Unit Amendment Study Plots. November.
- Arctadis, 2017. Year 5 Report on pH Monitoring to Evaluate the Effect of the White Rain on the Smelter Tailing Soils Investigation Unit. September.
- Golder Associates, 2015. Supplemental Completion Report, Razorback Ridge Area, Interim Remedial Action, STSIU, October 1.
- Arctadis, 2020. Draft Completion Report, B Ranch Area, Interim Remedial Action, STSIU, August.

**Footnotes:**  
 - Data for copper for small ground leaching bird applies to 0-6" soil sieved to < 2 mm (2000 µm). For FS RAC for human health applies to 0-1" soil sieved to < 0.25 mm (650 µm). The ratio used to convert copper concentration of 0-1" sieved to < 0.25 mm sieved to < 2 mm sieved is 0.67. Data for lead and zinc for small ground leaching bird applies to 0-6" soil sieved to < 2 mm (2000 µm). The ratio used to convert lead and zinc concentration of 0-1" sieved to < 0.25 mm sieved to < 2 mm sieved is 0.77. Multiplied copper concentration by 0.77 to adjust windblown tailings (and by 1.5 if finick; windblown tailings) if unsorted; multiplied by 0.9 if outside; windblown tailings and by 1.3 if inside; windblown tailings. The 0.9 and 1.3 are from median ratios of unsorted 0-1" soil in 2009 lab reports; (at 25 m on transect) compared to sieved 0-6" co-located soil in Drexler lab report in Appendix D or the pH Monitoring Report (Appendix B to the FS Report) for soils in and outside of windblown tailings, respectively (P-D 8, 18, 20). In tailings area, the data set for 0-6" soil was 0.67 and for 0-1" soil was 0.77.  
 - Data were collected for ecological endpoint and were not corrected to human health values.  
 - A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arctadis 2009. This sample was marked as IRA.  
 - A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arctadis 2009. This sample was marked as 21.  
 - A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arctadis 2009. This sample was marked as 21.  
 - A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arctadis 2009. This sample was marked as Volcanic Rock.  
 - A duplicate sample with the same Loc ID has been recorded in Golf Course Interim Remedial Action, Arctadis 2009. This sample was marked as NMDOT.  
 - Two samples were collected from the same location in the same sampling year (with identical sample location coordinates), an average was used in place of the two parent samples. For the co-located samples U04-11 and U04-12, the average was used in place of the two parent samples.  
 - Removed samples with the asterisk are set to 327 for the Theissens analyses. All other removed samples are completely voided.  
 - means no data available at the location.



**TABLE 3-3  
STSIU SOIL COPPER 95 UCLs**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Habitat Alliance Polygon ID	Number of Samples	Acres	Copper (mg/kg)			Ratio to 1600 Pre-FS RAC
			Minimum	Maximum	Spatially Weighted 95 UCL	
1-3	22	23	127	3204	1653	1.0
1-4	63	360	63	2527	537	0.3
1-5	2	53	1022	1120	1071 (a)	0.7
10-11	58	1785	96	1107	449	0.3
10-14	343	286	34	3463	1061	0.7
10-16	10	71	438	1640	1328	0.8
10-17	10	43	327	1560	708	0.4
10-8	55	873	160	2653	726	0.5
42-15	9	348	285	1611	503 (b)	0.3
88-10	15	107	259	1140	713	0.4
88-15	123	46	72	3463	1278	0.8
88-17	7	15	426	1560	998 (b)	0.6
88-19	82	879	34	3204	992	0.6
88-22	8	39	372	1611	705 (b)	0.4
137-7	358	116	58	3561	827	0.5
137-8	341	4127	34	3492	593	0.4
Mine Facilities/Urban	229	277	78	3565	455	0.3
Mine Facilities/Other (29-1)	253	324	14	3342	627	0.4

**Notes:**

mg/kg = milligram per kilogram

(a) Polygon has two samples and does not meet minimum sample count to generate bootstrap datasets and calculate a 95UCL; therefore, the average concentration was estimated.

(b) As described in Section 9.2 of the text, 95UCLs are based on the arithmetic mean (if the distribution of 95UCLs was symmetric) or geometric mean (if the distribution was asymmetric) of 500 iterations of bootstrapped datasets, which is just below the limit imposed by ProUCL 5.2 (column limit of 512). For three alliance polygons, ProUCL did not succeed on all 500 iterations of the bootstrapped datasets; Polygon 42-15 had 354 iterations, Polygon 88-17 had 149 iterations, and Polygon 88-22 had 224 iterations.

Gray shading indicates Ratio > 1 and exceeds the corresponding RAC indicated in the column heading.

TABLE 3-4  
STSIU SOIL pCu VALUES

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Location	Copper mg/kg	pH SU	pCu (calculated <sup>1</sup> )	Latitude	Longitude	Source
ERA02	1000	6.00	4.98	32.6890	-108.1064	Appendix B
ERA03	652	5.90	5.37	32.6850	-108.1047	Appendix B
ERA04	562	5.40	5.08	32.6891	-108.0922	Appendix B
ERA05	238	6.40	7.00	32.6892	-108.0911	Appendix B
ERA06	622	6.30	5.80	32.6990	-108.0958	Appendix B
ERA07	758	6.70	5.95	32.6758	-108.0998	Appendix B
ERA08	643	7.00	6.41	32.6749	-108.1031	Appendix B
ERA09	291	4.60	5.09	32.6864	-108.0729	Appendix B
ERA10	310	5.30	5.67	32.6837	-108.0666	Appendix B
ERA11	277	7.00	7.38	32.6567	-108.0610	Appendix B
ERA12	215	7.80	8.42	32.6459	-108.0636	Appendix B
ERA13	125	5.50	6.90	32.6825	-108.0489	Appendix B
ERA14	129	7.50	8.73	32.6430	-108.1187	Appendix B
ERA15	529	7.80	7.40	32.7062	-108.1406	Appendix B
FID 10	2550	4.70	2.69	32.7056	-108.1135	Appendix B
FID 101	285	3.90	4.47	32.6732	-108.0910	Appendix B
FID 102	282	3.50	4.11	32.6624	-108.0887	Appendix B
FID 103	443	4.00	4.05	32.6553	-108.0873	Appendix B
FID 104	459	3.83	3.83	32.6646	-108.0943	Appendix B
FID 105	834	4.50	3.79	32.6873	-108.1033	Appendix B
FID 106	516	4.40	4.25	32.6728	-108.0629	Appendix B
FID 15	1850	5.60	3.90	32.6978	-108.1069	Appendix B
FID 16	1440	4.80	3.44	32.6960	-108.1071	Appendix B
FID 17	4550	6.18	2.87	32.6978	-108.1135	Appendix B
FID 18	310	4.20	4.65	32.6740	-108.0918	Appendix B
FID 22	378	6.90	6.93	32.6669	-108.0780	Appendix B
FID 23	202	4.36	4.28	32.6568	-108.0841	Appendix B
FID 28	423	7.30	7.17	32.6700	-108.0511	Appendix B
FID 37	708	4.70	4.16	32.7064	-108.0954	Appendix B
FID 43	636	6.50	5.96	32.6589	-108.0884	Appendix B
FID 7	491	5.10	4.96	32.6786	-108.0675	Appendix B
FID 8	473	4.60	4.54	32.6668	-108.0920	Appendix B
Reference Plot #1	882	8.00	6.98	32.7065	-108.1383	Appendix B
Reference Plot #2	760	6.20	5.48	32.7130	-108.1297	Appendix B
Reference Plot #3	1540	5.40	3.92	32.7038	-108.1111	Appendix B
Reference Plot #4	1020	4.90	3.93	32.6897	-108.1040	Appendix B
STS-PCUG-2011-1	263	5.60	6.14	32.6327	-108.0785	FS workplan
STS-PCUG-2011-10	324	7.4	7.57	32.6382	-108.0664	FS workplan
STS-PCUG-2011-11	254	4.60	5.25	32.6774	-108.0802	FS workplan
STS-PCUG-2011-12	536	6.70	6.34	32.6984	-108.0822	FS workplan
STS-PCUG-2011-13	602	5.10	4.72	32.7129	-108.0871	FS workplan
STS-PCUG-2011-14	354	5.90	6.08	32.6453	-108.0831	FS workplan
STS-PCUG-2011-15	357	4.30	4.58	32.6966	-108.0735	FS workplan

TABLE 3-4  
STSIU SOIL pCu VALUES

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Location	Copper mg/kg	pH SU	pCu (calculated <sup>1</sup> )	Latitude	Longitude	Source
STS-PCUG-2011-16	864	5.2	4.40	32.7158	-108.1146	FS workplan
STS-PCUG-2011-17	994	5.10	4.15	32.7046	-108.1025	FS workplan
STS-PCUG-2011-18	1540	5.3	3.83	32.7122	-108.1140	FS workplan
STS-PCUG-2011-19	1210	3.90	2.80	32.6925	-108.1046	FS workplan
STS-PCUG-2011-2	876	6.50	5.59	32.7022	-108.0842	FS workplan
STS-PCUG-2011-20	520	4.10	3.96	32.6784	-108.0920	FS workplan
STS-PCUG-2011-21	558	4.80	4.53	32.7079	-108.0973	FS workplan
STS-PCUG-2011-22	976	7.4	6.31	32.7124	-108.1416	FS workplan
STS-PCUG-2011-23	551	5.80	5.48	32.7173	-108.1003	FS workplan
STS-PCUG-2011-24	1000	4.2	3.30	32.7102	-108.1057	FS workplan
STS-PCUG-2011-25	706	4.6	4.07	32.7077	-108.1046	FS workplan
STS-PCUG-2011-27	438	6.9	6.76	32.7170	-108.1461	FS workplan
STS-PCUG-2011-28	959	7.5	6.42	32.7162	-108.1389	FS workplan
STS-PCUG-2011-29	671	5.2	4.69	32.7185	-108.1033	FS workplan
STS-PCUG-2011-3	587	4.80	4.47	32.7013	-108.0765	FS workplan
STS-PCUG-2011-30	1500	7.4	5.81	32.7145	-108.1375	FS workplan
STS-PCUG-2011-31	304	4.30	4.76	32.6517	-108.0845	FS workplan
STS-PCUG-2011-32	420	3.80	3.93	32.6499	-108.0784	FS workplan
STS-PCUG-2011-33	273	6.7	7.12	32.6474	-108.0717	FS workplan
STS-PCUG-2011-34	364	4.00	4.28	32.6541	-108.0766	FS workplan
STS-PCUG-2011-35	287	5.50	5.95	32.6723	-108.0848	FS workplan
STS-PCUG-2011-36	270	5.60	6.11	32.6723	-108.0703	FS workplan
STS-PCUG-2011-37	244	6.90	7.44	32.6612	-108.0815	FS workplan
STS-PCUG-2011-38	350	3.90	4.23	32.6613	-108.0740	FS workplan
STS-PCUG-2011-39	360	4.70	4.94	32.6637	-108.0668	FS workplan
STS-PCUG-2011-4	794	4.60	3.94	32.7150	-108.0941	FS workplan
STS-PCUG-2011-40	312	3.80	4.27	32.6532	-108.0693	FS workplan
STS-PCUG-2011-41	587	3.30	3.08	32.6377	-108.0806	FS workplan
STS-PCUG-2011-5	458	5.8	5.69	32.7215	-108.1193	FS workplan
STS-PCUG-2011-6	290	5.00	5.47	32.6900	-108.0677	FS workplan
STS-PCUG-2011-7	387	7.7	7.65	32.7153	-108.1545	FS workplan
STS-PCUG-2011-8	449	5.8	5.71	32.7204	-108.1148	FS workplan
STS-PCUG-2011-9	246	4.80	5.47	32.6858	-108.0816	FS workplan
STS-RWU-2011-1	338	5.20	5.48	32.7124	-108.1083	Appendix C
STS-RWU-2011-10	96	4.6	6.37	32.6748	-108.0840	Appendix C
STS-RWU-2011-11	216	4.3	5.16	32.6747	-108.0920	Appendix C
STS-RWU-2011-12	316	3.9	4.35	32.6642	-108.0870	Appendix C
STS-RWU-2011-13	305	5.6	5.97	32.6768	-108.0940	Appendix C
STS-RWU-2011-14	1640	5.3	3.76	32.7081	-108.1150	Appendix C
STS-RWU-2011-15	1640	5.7	4.13	32.7092	-108.1180	Appendix C
STS-RWU-2011-16	395	4.9	5.02	32.7048	-108.0850	Appendix C
STS-RWU-2011-17	654	4.6	4.16	32.6762	-108.0960	Appendix C
STS-RWU-2011-2	381	4.10	4.32	32.7045	-108.1050	Appendix C
STS-RWU-2011-3	998	5.10	4.14	32.7076	-108.1070	Appendix C
STS-RWU-2011-4	427	7.20	7.07	32.7123	-108.1430	Appendix C
STS-RWU-2011-5	779	4.60	3.96	32.7067	-108.0950	Appendix C
STS-RWU-2011-6	1300	7.30	5.88	32.7085	-108.1209	Appendix C
STS-RWU-2011-7	529	4.90	4.69	32.6972	-108.1060	Appendix C
STS-RWU-2011-8	287	5.60	6.04	32.7103	-108.0939	Appendix C
STS-RWU-2011-9	560	4.4	4.15	32.6959	-108.1000	Appendix C
STS-RWU-2012-B1	182	4.6	5.63	32.6714	-108.0445	Appendix C
STS-RWU-2012-B2	344	4.7	4.99	32.6714	-108.0423	Appendix C
STS-RWU-2012-B3	161	4.7	5.87	32.6738	-108.0449	Appendix C
STS-SS-2010-016	1211	4.9	3.73	32.7070	-108.1121	FS workplan

TABLE 3-4  
STSIU SOIL pCu VALUES

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Location	Copper mg/kg	pH SU	pCu (calculated <sup>1</sup> )	Latitude	Longitude	Source
STS-SS-2010-017	2227	6	4.06	32.6971	-108.1080	FS workplan
STS-SS-2010-018	1162	6	4.80	32.7039	-108.1063	FS workplan
West Amendment Plot	1767	7.68	5.96	32.7059	-108.1378	Appendix A
Wildlife Reference Plot North	213	5.90	6.66	32.6840	-108.0677	Appendix C
Wildlife Reference Plot South	288	4.60	5.11	32.6748	-108.0601	Appendix C
ERA16	77	6.10	8.02	32.6108	-108.1969	Appendix B
ERA162	153	6.49	7.59	32.7654	-108.1026	Appendix B
ERA163	146	6.95	8.08	32.7691	-108.1083	Appendix B
ERA164	95	5.62	7.33	32.7435	-108.0096	Appendix B
ERA165	124	6.90	8.21	32.7491	-108.0258	Appendix B
ERA17	57	5.63	7.93	32.6131	-108.1921	Appendix B
ERA18	73	6.07	8.04	32.6141	-108.1868	Appendix B
ERA19	62	6.68	8.80	32.6093	-108.1838	Appendix B
ERA20	45	7.73	10.15	32.6089	-108.1854	Appendix B
ERA21	48	6.20	8.65	32.6080	-108.1875	Appendix B
ERA25	70	7.73	9.64	32.6729	-108.0461	Appendix B
ERA27	328	5.77	6.04	32.6183	-108.0861	Appendix B
ERA28	1060	7.53	6.34	32.7472	-108.1298	Appendix B
ERA31	78	7.73	9.53	32.5797	-108.0458	Appendix B
ERA33	176	6.60	7.53	32.7655	-108.1161	Appendix B
S64	482	7.49	7.20	32.7494	-108.1323	Appendix B
S65	462	4.78	4.73	32.7437	-108.1310	Appendix B
S66	552	6.61	6.23	32.7387	-108.1315	Appendix B
S78	207	7.79	8.45	32.6222	-108.1271	Appendix B
S79	157	7.95	8.92	32.6221	-108.1108	Appendix B
SS100	164	6.17	7.21	32.7632	-108.0651	Appendix B
SS101	144	6.95	8.09	32.7608	-108.0469	Appendix B
SS103	348	4.52	4.81	32.7443	-108.1389	Appendix B
SS104	285	5.70	6.14	32.7511	-108.1130	Appendix B
SS107	136	5.53	6.84	32.7332	-108.0734	Appendix B
SS108	176	6.62	7.55	32.7350	-108.0302	Appendix B
SS113	146	6.93	8.05	32.7371	-108.0612	Appendix B
SS114	83	6.92	8.69	32.7278	-108.0340	Appendix B
SS119D	125	6.10	7.46	32.7224	-108.0570	Appendix B
SS120	83	7.18	8.93	32.7066	-108.0350	Appendix B
SS123	314	5.74	6.07	32.6995	-108.0382	Appendix B
SS148	442	3.96	4.02	32.6256	-108.0813	Appendix B
SS149	440	8.16	7.93	32.6248	-108.0724	Appendix B
SS150	424	8.06	7.86	32.6248	-108.0590	Appendix B
SS151	181	7.97	8.77	32.6170	-108.1193	Appendix B
SS152	166	8.24	9.13	32.6131	-108.1027	Appendix B
SS153	307	6.78	7.06	32.6115	-108.0862	Appendix B
SS154	260	7.75	8.15	32.6133	-108.0732	Appendix B
SS155	271	8.10	8.43	32.6136	-108.0593	Appendix B
SS156	137	5.06	6.39	32.5993	-108.0860	Appendix B
SS157	99	8.09	9.58	32.5990	-108.0725	Appendix B
SS158	173	7.98	8.84	32.5993	-108.0582	Appendix B
SS97	288	5.97	6.38	32.7519	-108.1471	Appendix B
SS98	333	5.97	6.21	32.7474	-108.1250	Appendix B
SS99	65	5.77	7.91	32.7586	-108.0924	Appendix B
T-09	581	7.14	6.66	32.6283	-108.0801	Appendix B
T-12	194	7.78	8.52	32.6081	-108.0947	Appendix B
U04-1019	221	5.94	6.66	32.6957	-108.0365	Appendix B
U04-1020	392	4.63	4.78	32.6977	-108.0457	Appendix B

TABLE 3-4  
STSIU SOIL pCu VALUES

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Location	Copper mg/kg	pH SU	pCu (calculated <sup>1</sup> )	Latitude	Longitude	Source
U04-1021	252	5.76	6.34	32.6886	-108.0381	Appendix B
U04-1032	936	7.92	6.84	32.7048	-108.1562	Appendix B
U04-1033	506	8.78	8.34	32.6976	-108.1560	Appendix B
U04-1034	382	8.65	8.55	32.6892	-108.1566	Appendix B

**Notes:**

mg/kg = milligram per kilogram

SU = standard unit

<sup>1</sup> used upland pCu equation from NewFields 2005, applied all points in table to pCu contours developed with natural neighbor method.

All samples taken from the 0-6" bgs interval or adjusted to 0-6" interval with sieving at < 2 mm (see Table E-1 in Appendix B)

**TABLE 3-5  
ERROR MATRICES**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

<b>Rangeland Condition Classification<sup>1</sup></b>				
<b>Classified</b>	<b>Observed</b>		<b>User's Accuracy</b>	<b>Error of Commission</b>
	<b>Unacceptable</b>	<b>Acceptable</b>		
<b>Unacceptable</b>	2	0	100%	0%
<b>Acceptable</b>	1	5	83%	17%
<b>Producer's Accuracy</b>	67%	100%		
<b>Error of Omission</b>	33%	0%		

**Overall Accuracy: 88%**

1. Based on a comparison of 2011 ground-verified OAT scores on 200-m transect that were independent of training dataset (1/4 of total dataset was not used for training) compared to majority classified rangeland class within the 200 m x 200 m area. Training dataset had 74% overall accuracy (18% error of commission), and all data produced 74% accuracy.

<b>Vegetation Cover Classification<sup>2</sup></b>				
<b>Classified</b>	<b>Observed</b>		<b>User's Accuracy</b>	<b>Error of Commission</b>
	<b>Unacceptable</b>	<b>Acceptable</b>		
<b>Unacceptable</b>	9	5	64%	36%
<b>Acceptable</b>	3	14	82%	18%
<b>Producer's Accuracy</b>	75%	74%		
<b>Error of Omission</b>	25%	26%		

**Overall Accuracy: 74%**

2. Based on a comparison of 2011 ground visit assessed vegetative cover independent of process of developing the mapped rangeland polygons of cover.

<b>Vegetation Richness Classification<sup>3</sup></b>				
<b>Classified</b>	<b>Observed</b>		<b>User's Accuracy</b>	<b>Error of Commission</b>
	<b>Unacceptable</b>	<b>Acceptable</b>		
<b>Unacceptable</b>	16	5	76%	24%
<b>Acceptable</b>	4	6	60%	40%
<b>Producer's Accuracy</b>	80%	55%		
<b>Error of Omission</b>	20%	45%		

**Overall Accuracy: 71%**

3. Based on a comparison of 2011 ground visit assessed vegetative richness independent of process of developing mapped rangeland polygons of richness.

TABLE 3-6  
VALIDATION SITE RESULTS

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Location	Soil Category Observed	Soil Category Classified	Rangeland Condition (OAT)		Vegetation Cover	
			Observed	Classified	Observed	Classified
STS-PT-2013-1	Flat Rocky	Flat Rocky	Unacceptable	Unacceptable	Acceptable	Acceptable
STS-PT-2013-2	Flat Rocky	Slope > 13%	Unacceptable	Unacceptable	Acceptable	Acceptable
<b>STS-PT-2013-5</b>	Slope > 13%	Slope > 13%	Acceptable	Acceptable	Acceptable	Acceptable
STS-PT-2013-9	Bedrock	Bedrock	Unacceptable	Unacceptable	Unacceptable	Unacceptable
STS-PT-2013-12	Bedrock	Bedrock	Acceptable	Acceptable	Acceptable	Acceptable
STS-PT-2013-17	Flat Rocky	Flat Rocky	Unacceptable	Unacceptable	Acceptable	Unacceptable
STS-PT-2013-19	Flat Rocky	Flat Rocky	Unacceptable	Unacceptable	Unacceptable	Unacceptable
STS-PT-2013-20	Flat Granular	Flat Granular	Acceptable	Acceptable	Acceptable	Acceptable
STS-RWU-2012-B1	Bedrock	Flat Rocky	Acceptable	Unacceptable	Acceptable	Unacceptable
STS-RWU-2012-B2	Bedrock	Flat Rocky	Unacceptable	Unacceptable	Unacceptable	Unacceptable
STS-RWU-2012-B3	Bedrock	Slope > 13%	Acceptable	Unacceptable	Unacceptable	Unacceptable
STS-RWU-2011-1	Bedrock	Bedrock	Unacceptable	Unacceptable	Unacceptable	Acceptable
STS-RWU-2011-2	Bedrock	Bedrock	Unacceptable	Unacceptable	Acceptable	Acceptable
STS-RWU-2011-3	Slope > 13%	Flat Granular	Acceptable	Acceptable	Acceptable	Acceptable
STS-RWU-2011-4	Flat Granular	Flat Granular	Acceptable	Acceptable	Acceptable	Acceptable
STS-RWU-2011-5	Flat Granular	Flat Rocky	Acceptable	Unacceptable	Acceptable	Unacceptable
STS-RWU-2011-6	Slope > 13%	Slope > 13%	Unacceptable	Acceptable	Unacceptable	Acceptable
STS-RWU-2011-7	Flat Rocky	Bedrock	Unacceptable	Unacceptable	Unacceptable	Unacceptable
<b>STS-RWU-2011-8</b>	Slope > 13%	Slope > 13%	Acceptable	Acceptable	Acceptable	Acceptable
STS-RWU-2011-9	Bedrock	Bedrock	Unacceptable	Unacceptable	Unacceptable	Unacceptable
<b>STS-RWU-2011-10</b>	Flat Granular	Flat Rocky	Unacceptable	Unacceptable	Acceptable	Unacceptable
STS-RWU-2011-11	Bedrock	Bedrock	Unacceptable	Unacceptable	Unacceptable	Unacceptable
STS-RWU-2011-12	Flat Rocky	Flat Rocky	Unacceptable	Unacceptable	Unacceptable	Unacceptable
<b>STS-RWU-2011-13</b>	Flat Granular	Flat Rocky	Unacceptable	Unacceptable	Acceptable	Unacceptable
<b>STS-RWU-2011-14</b>	Slope > 13%	Slope > 13%	Acceptable	Acceptable	Unacceptable	Acceptable
STS-RWU-2011-15	Flat Granular	Flat Rocky	Unacceptable	Unacceptable	Acceptable	Acceptable
<b>STS-RWU-2011-16</b>	Flat Granular	Bedrock	Acceptable	Acceptable	Acceptable	Acceptable
<b>STS-RWU-2011-17</b>	Flat Rocky	Flat Granular	Unacceptable	Acceptable	Acceptable	Acceptable
Overgrazed Reference	Flat Rocky	Flat Rocky	Unacceptable	Unacceptable	Unacceptable	Unacceptable
<b>Wildlife reference plot S</b>	Flat Granular	Flat Rocky	Acceptable	Acceptable	Acceptable	Acceptable
Wildlife reference plot N	Flat Granular	Flat Rocky	Acceptable	Acceptable	Acceptable	Acceptable

**Bolded** are independent data used to assess OAT score accuracy.

**TABLE 3-7**  
**EVALUATION OF SOIL CATEGORIES BASED ON PROBABLE EFFECT LEVELS (PELS)**  
**FREEMPORT-MCMORAN CHINO MINES COMPANY**  
**VANADIUM, NEW MEXICO**  
**SMELTER/TAILING SOILS IU FEASIBILITY STUDY**

Soil category	PEL	Minimum pCu of retained polygons	Below PEL and include for remediation?
Flat Rocky	4.98	3.65	<b>Yes, include</b>
Slope > 13%	4.11	3.84	<b>Yes, include</b>
Bedrock	4.4	3.86	<b>Yes, include</b>
Flat Granular	4.11	NA	No, exclude



TABLE 4-1  
 SURFACE WATER CRITERIA EVALUATION FOR METALS - EPHEMERAL WATERSHEDS  
 FREEPORT-MCMORAN CHINO MINES COMPANY  
 MANDALAY, NEW MEXICO  
 SMEI TERTIARY SOILS TO FEASIBILITY STUDY

Stream Location	Year	Hardness (mg/L)	pH	Aluminum		Cadmium		Lead		Nickel		Silver		Zinc	
				Disolved (ug/L)	Chronic (ug/L)	Disolved (ug/L)	Chronic (ug/L)	Disolved (ug/L)	Chronic (ug/L)	Disolved (ug/L)	Chronic (ug/L)	Disolved (ug/L)	Chronic (ug/L)	Disolved (ug/L)	Chronic (ug/L)
				Q	Q	Q	Q	Q	Q	Q	Q	Q	Q	Q	Q
Lambright Draw Watershed															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4
Sub-Watershed A															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4
Sub-Watershed B															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4
Sub-Watershed C															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4
Sub-Watershed D															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4
Sub-Watershed E															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4
Sub-Watershed F															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4
Sub-Watershed G															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4
Sub-Watershed H															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4
Sub-Watershed I															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4
Sub-Watershed J															
ERA 25	2000	187	--	--	3230	0.5	1.1	1.1	1.1	5.0	0.16	5.0	0.16	5.0	0.16
ERA 45	2000	202	--	--	1.5	3.2	0.4	1.2	1.2	192.7	0.16	0.14	5.4	5.4	5.4

Notes:  
 Acute & Chronic Criteria Calculated with equations 1 (a-c) or 2 (a-c) of New Mexico Administrative Code (NMAC) 20.6.6.8.9(b). Effective February 8, 2020. However, for Aluminum, hardness-adjusted total aluminum criteria are applied to water of pH 6.5 to 9. Only Acute criteria are applicable to these ephemeral drainages.  
 \* = pH outside of 4-9 range and therefore Aluminum is reported as dissolved. If pH unknown, assumed to be average for reach when selecting criteria for comparison.  
 Q = qualifier, if non-detect (<), method detection limit shown  
 Bold = HQ > 1. Italicized bold = gray, the HQ is based on non-detect value and is not reliable.  
 Calculation equations to adjust for hardness are shown below.

Metal	Acute (ug/L)		Conversion factor (CF)
	m <sub>a</sub>	b <sub>a</sub>	
Aluminum (Al)	1.37	1.6308	1.183575 (h)
Cadmium (Cd)	0.979	3.866	hardness(0.04) (1838)
Chromium (Cr) III	0.819	3.7256	0.316
Copper (Cu)	0.842	1.17	1.40202 (h)
Lead (Pb)	1.273	1.46	hardness(0.14)
Manganese (Mn)	0.333	6.4676	57121
Nickel (Ni)	0.846	2.255	0.988
Silver (Ag)	1.72	4.659	0.95
Zinc (Zn)	10.626	0.976	0.9285
Metal	Chronic		Conversion factor (CF)
	m <sub>c</sub>	b <sub>c</sub>	
Aluminum (Al)	1.37	0.5462	0.3985
Chromium (Cr) III	0.819	0.6843	0.86
Copper (Cu)	0.855	1.702	0.98
Manganese (Mn)	0.333	5.6743	0.997
Nickel (Ni)	0.846	0.988	0.988
Zinc (Zn)	10.626	0.9235	0.966
Criteria (ug/L) = exp(m <sub>a</sub> (lnhardness) + b <sub>a</sub> )(CF)			





**TABLE 4-4  
WATER-EFFECT RATIOS (WER) USED TO EVALUATE THE NATURE AND EXTENT OF COPPER CONTAMINATION**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

<b>Sample ID</b>	<b>Sub-Watershed</b>	<b>Water effect ratio (WER)</b>
<b>Round 1 Samples</b>		
WER 1-1	Lucky Bill	6.651
WER 1-2	Lucky Bill	5.334
WER 1-5 <sup>a</sup>	Lower C-1 Drainage	--
WER 1-6	Upper C-1 Drainage	14.407
WER 1-7	B Drainage	4.717
WER 1-9	Martin Canyon	2.207
WER 1-10	Martin Canyon	2.804
WER 1-11	G-drainage	5.956
WER 1-12	Rustler Canyon	0.989
WER 1-RCS	Rustler Canyon	3.273
WER D1-2	D-1 Drainage	13.104
WER D2-1b	D-2 Drainage	8.027
<b>Round 2 Samples</b>		
WER 2-1	Lucky Bill	4.046
WER 2-6	Upper C-1 Drainage	6.151
WER 2-9	Martin Canyon	11.530
WER 2-11	G-drainage	6.889
WER 2-12	Rustler Canyon	2.251
WER 2-D12	D-1 Drainage	5.724

**Geometric mean all STSIU WERs = 5.003**

**Notes:**

a. WER not derived based on results of toxicity test and WER guidance.

**TABLE 4-5  
SURFACE WATER CRITERIA EVALUATION FOR DISSOLVED COPPER - EPHEMERAL WATERSHEDS**

**FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY**

Stream Location	Year	Hardness (mg/L)	Dissolved Cu (µg/L)	Acute Hardness-based Criterion (µg/L)	Acute Hardness-based HQ	Watershed-specific WER-based Acute HQ <sup>1</sup>	IU-based WER-based Acute HQ <sup>2</sup>
<b>Lampbright Draw Watershed</b>							
<b>Lampbright Draw</b>							
ERA 25	2000	187	15.0	24.2	0.6	WBT) adjustm	0.1
ERA 45	2000	202	10.8	26.1	0.4	0.1	0.1
SW-10-2006	2006	89	4.1	12.0	0.34		0.1
<b>Sub-Watershed A</b>							
<b>A-Drainage</b>							
STS-WS-2010-A2	2010	32	73.9	4.6	16		3
STS-WS-2010-A4	2010	30	25.0	4.3	6		1
<b>Sub-Watershed B</b>							
<b>B-Drainage</b>							
WER-1-7	2011	108	43.0	14.4	3		0.6
STS-WS-2010-B-3	2010	111	35.8	14.8	2		0.5
2013-SW-B-AS	2013	233	44.7	29.8	1		0.3
2013-SW-WER-7	2013	78	40.0	10.6	4		0.8
<b>Sub-Watershed C (non-headwaters)</b>							
<b>C-1 Drainage (downgradient of Bolton Canyon Confluence)</b>							
STS-WS-2010-C-1	2010	62	34.8	8.6	4		0.8
STS-WS-2010-C-2	2010	53	36.6	7.4	5		1.0
WER-1-5	2011	64	32.3	8.8	4		0.7
BD4W-1-2004	2004	68	207.0	9.3	22		4
BD4W-1-2007	2007	70	80.0	9.6	8		2
2013-SW-WER-5	2013	63	33.6	8.7	4		0.8
<b>C-2 Drainage (downgradient from confluence with CLF critical habitat)</b>							
ERA37	2000	103	64.8	13.8	5		0.9
ERA38	2000	94	150.0	12.7	12		2
WER-1-BD	2011	66	94.1	9.1	10		2
2013-SW-WER-BD	2013	35	80.6	5.0	16		3
2013-SW-C2-Lower	2013	45	38.3	6.3	6		1
<b>Sub-Watershed D</b>							
<b>D-1 Drainage</b>							
WER-1-D1-2	2011	52	32.3	7.3	4		0.9
WER-1-D1	2011	27	21.1	3.9	5		1
WER-2-D1-2	2011	57	17.9	7.9	2		0.5
STS-WS-2010-D1-2	2010	56	25.0	7.8	3		0.6
2013-SW-WERD1-2	2013	44	41.0	6.2	7		1
<b>D-2 Drainage</b>							
WER-1-D2-1	2011	44	32.8	6.2	5		1
WER-D2-2	2011	55	18.8	7.7	2		0.5
STS-WS-2010-D2-1	2010	47	20.0	6.6	3		0.6
2013-SW-D2	2013	22	60.8	3.2	19		4
<b>D-3 Drainage</b>							
CDW-1-2004	2004	53	327.0	7.4	44		9
CDW-1-2007	2007	35	221.0	5.0	44		9
CDW-1-2013	2013	37	208.1	5.3	40		8
<b>Sub-Watershed G</b>							
<b>G-Drainage</b>							
WER-1-11	2011	164	4.3	21.4	0.2		0.0
WER-2-11	2011	119	7.9	15.8	0.5		0.1

**Notes:**

1. HQs adjusted using the WER determined in the sample's watershed, or nearest watershed (e.g., Subwatershed A used Subwatershed where multiple WERs were determined in a single watershed, the geometric mean of these WERs was calculated and used as a watershed summary of watershed-specific WERs is provided below:

Sub-watershed B	4.717
Sub-watershed C	9.414
Sub-watershed D-1	8.661
Sub-watershed D-2	8.027
Sub-watershed G	6.406

2. HQs adjusted using the STSIU geometric mean WER (5.003) presented in Table 4-4.

**TABLE 4-6  
SURFACE WATER CRITERIA EVALUATION FOR DISSOLVED COPPER - NON-EPHEMERAL WATERSHEDS**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Stream Location	Year	Hardness (mg/L)	Dissolved Cu (µg/L)	Acute Hardness-based Criterion (µg/L)	Acute Hardness-based HQ	Watershed-specific WER-based Chronic HQ <sup>1</sup>	IU-based WER-based Chronic HQ <sup>2</sup>
<b>Lucky Bill Canyon</b>							
WER-1-2	2011	81	6.5	11.01	0.59	0.17	0.17
WER-1-1	2011	91	5.9	12.29	0.48	0.14	0.14
WER-2-1	2011	97	3.4	13.03	0.26	WBT) adjustm	0.08
STS-CA-2010-004	2010	93	3.3	12.55	0.26	0.07	0.08
SW01-2007	2007	46	19.3	6.47	3		0.84
SW02-2007	2007	63	13.9	8.70	2		0.46
SW03-2007	2007	95	14.5	12.81	1		0.34
SW01-2006	2006	69	15.3	9.52	2		0.47
SW02-2006	2006	255	5	32.50	0.15		0.05
SW03-2006	2006	223	20.9	28.57	0.73		0.24
<b>Martin Canyon</b>							
STS-CA-2010-008	2010	139	13.1	18.33	0.71		0.22
STS-WS-2010-MC	2010	141	14.1	18.58	0.76		0.23
SW06A-2007	2007	25	30.1	3.64	8		2
SW07-2006	2006	90	15.1	12.20	1		0.37
SW07-2007	2007	97	8.1	13.06	0.62		0.19
SW08-2006	2006	152	5	19.95	0.25		0.08
SW08-2007	2007	109	4.7	14.58	0.32		0.10
WER-1-10	2011	258	5.4	32.79	0.16		0.05
WER-1-9	2011	86	7.1	11.67	0.61		0.18
WER-2-9	2011	86	13.7	11.60	1		0.35
WER-MC-1	2011	105	8.1	14.11	0.57		0.17
<b>Rustler Canyon</b>							
STS-CA-2010-001	2010	73	2.7	9.99	0.27		0.08
SW05-2006	2006	345	55	43.21	1		0.43
SW05-2007	2007	365	1220	45.52	27		9
SW06-2006	2006	41	19.7	5.76	3		0.95
SW09-2006	2006	130	9.1	17.16	0.53		0.16
SW09-2007	2007	119	11	15.83	0.69		0.21
SW10-2007	2007	74	2.6	10.12	0.26		0.08
SW-210-2006	2006	89	2.5	12.09	0.21		0.06
WER-1-12	2011	74	2.1	10.12	0.21		0.06
WER-1-RCS1	2011	45	5	6.27	0.80		0.22
WER-1-RCS2	2011	40	5.3	5.70	0.93		0.26
WER-1-RCS3	2011	67	2.2	9.21	0.24		0.07
WER-2-12	2011	69	3.6	9.42	0.38		0.11
<b>Headwaters of Sub-Watershed C in Drainage C-1 (upgradient of Bolton Canyon confluence)</b>							
STS-WS-2010-C-3	2010	57	23.4	7.9	3		0.8
WER-1-6	2011	54	57.4	7.5	8		2
WER-2-6	2011	47	30.2	6.6	5		1
2013-SW-WER-6	2013	30	72.3	4.3	17		5
2013-SW-C-BS	2013	47	48.2	6.6	7		2
2013-SW-C-BS-D	2013	41	61.8	5.8	11		3
2013-SW-C-BC	2013	29	28.3	4.2	7		2
<b>C-2 Drainage of Sub-Watershed C (upstream from confluence and in CLF critical habitat)</b>							
2013-SW-C2-Lower	2013	45	38.3	6.3	6		2
2013-SW-C2-Upper	2013	48	33.6	6.7	5		1

**Notes:**

1. HQs adjusted using the WER determined in the sample's watershed, or nearest watershed. In cases where multiple local WERs were determined the geometric mean of these WERs was calculated and used as a watershed-specific WER. A summary of watershed-specific WERs is provided below.

Lucky Bill Canyon	5.236
Martin Canyon	4.148
Rustler Canyon	1.939
Sub-watershed C	9.414

2. HQs adjusted using the STSIU geometric mean WER (5.003) presented in Table 4-4.

**TABLE 4-7  
STOCK TANK SURFACE WATER CRITERIA EVALUATION FOR DISSOLVED COPPER**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Stock Tank Location	Stock Tank ID #	Year	Hardness (mg/L)	Dissolved Cu (µg/L)	Acute Hardness-based Criteria (µg/L)	Chronic Hardness-based HQ	Watershed-specific WER-Adjusted Acute HQ <sup>1</sup>	IU-based WER-Adjusted Acute HQ <sup>2</sup>
<b>Sub Drainage A</b>								
STS-CA-2010-006	T15 <sup>3</sup>	2010	37	47.0	5.27	12.27	2	2
							2	2
SW-4-2004	T16	2004	46	37.1	6.48	WBT) adjustm	1	1
SW-13-2006	T16	2006	48	49.5	6.79	10	2	1
SW13-2007	T16	2007	74	45.6	10.12		0.96	0.90
STS-WS-2010-A1	T16	2010	114	33.1	15.20		0.46	0.44
T16 Stock Tank	T16	2021	98	37.5	13.19		0.60	0.57
							0.95	0.82
Stock Tank ERA41	T25	2000	43	28.6	6.07		1.00	0.94
SW-1-2004	T25	2004	70	43.6	9.62		0.96	0.91
SW-12-2006	T25	2006	37	51.4	5.28		2	2
SW12-2007	T25	2007	54	45.8	7.52		1	1
STS-WS-2010-A3	T25	2010	78	40.8	10.63		0.81	0.77
							1	1
Stock Tank ERA40	T26 <sup>3</sup>	2000	45	25.6	6.27		0.87	0.82
SW-2-2004	T26 <sup>3</sup>	2004	62	51.4	8.59		1	1
SW-11-2006	T26 <sup>3</sup>	2006	41	48.7	5.86		2	2
SW11-2007	T26 <sup>3</sup>	2007	42	39.7	5.93		1	1
							1	1
SW-3-2004	T60 <sup>3</sup>	2004	25	38.0	3.65		2	2
SW-14-2006	T60 <sup>3</sup>	2006	27	51.8	3.94		3	3
SW14-2007	T60 <sup>3</sup>	2007	26	48.7	3.78		3	3
							3	2
<b>Sub Drainage B</b>								
Stock Tank ERA39	T34	2000	192	28.0	24.85		0.24	0.23
SW-5-2004	T34	2004	47	60.6	6.54		2	2
SW15-2006	T34	2006	46	72.1	6.52		2	2
SW15-2007	T34	2007	57	52.2	7.91		1	1
STS-WS-2010-B-2	T34	2010	145	73.9	19.07		0.82	0.77
							1	0.99
<b>Sub Drainage C</b>								
STS-WS-2010-C-4	T29 <sup>3</sup>	2010	21	53.1	3.09		2	3
							2	3
<b>Sub Drainage D</b>								
STS-WS-2010-D1-3	T53	2010	61	33.6	8.44		0.46	0.80
ST-D12-2	T53	2011	56	44.1	7.78		0.65	1
							0.55	0.95
STS-WS-2010-D1-1	T54	2010	76	13.0	10.38		0.14	0.25
							0.14	0.25
<b>Sub Drainage G</b>								
ST-G8	T36	2011	94	--	12.68		--	--
ST-G8-2	T36	2011	122	2.3	16.21		0.02	0.03
							0.02	0.03
<b>Sub Drainage E</b>								
SW-6-2004	T06	2004	96	95	12.89		--	1
SW04-2006	T06	2006	54	220	7.51		--	6
SW04-2007	T06	2007	83	129	11.28		--	2
ST-E-2	T06	2011	55	79	7.60		--	2
T06 Stock Tank	T06	2021	139	113	18.33		--	1
T06 Stock Tank	T06	2022	90	142	12.17		--	2
							--	2

**Notes:**

- HQs adjusted using the WER determined in the sample's watershed, or nearest watershed. In cases where multiple WERs were determined in a watershed, the geometric mean of these WERs was calculated and used as a watershed-specific WER. A summary of watershed-specific WERs is provided in Table 4-4.
- HQs adjusted using the STSIU geometric mean WER (5.003) presented in Table 4-4.

Sub-watershed B 4.717  
 Sub-watershed C 9.414  
 Sub-watershed D-1 8.661  
 Sub-watershed D-2 8.027  
 Sub-watershed G 6.406

- Stock tank was removed between 2013 and 2022 as part of ongoing rangeland improvements.

TABLE 5-1  
SOIL REMEDIAL TECHNOLOGIES  
FREETPORT-MCDORAN CHINO MINES COMPANY  
VANADUM, NEW MEXICO  
SMELT TAILINGS SOILS FEASIBILITY STUDY

NO.	REMEDIAL TECHNOLOGY	DESCRIPTION	PRELIMINARY SCREENING		CONCLUSION
			EFFECTIVENESS	IMPLEMENTABILITY	
1	No Action	No further active response actions.	Contaminants will naturally attenuate over time. Does not provide additional mechanisms to prevent contaminant exposure to site receptors.	Is considered implementable.	Being retained for baseline comparison of remedial technologies.
2	Monitoring	No further active response actions. Monitoring will be conducted to prove the occurrence of natural remediation.	Contaminants will naturally attenuate over time. Does not provide additional mechanisms to prevent contaminant exposure to site receptors.	Is considered implementable.	Being retained for evaluation and as part of a remedial alternative.
3	Excavation and Reuse	Excavation and onsite management includes removal of contaminated soils with heavy construction equipment and hand removal techniques. Excavated soils will be managed onsite at the waste rock stockpiles or reused in the adjacent operational areas as cover for the tailings ponds.	Excavation	Is considered highly effective at reducing the presence of site contaminants.	Costs include equipment use and maintenance, material handling and transport, and backfill materials and activities. Costs are considered less than Excavation and Disposal option because offsite transportation and disposal is not needed. Long term O&M costs are considered to be low.
3a	Excavation and Disposal	Excavation and disposal includes removal of contaminated soils with heavy construction equipment and hand removal techniques. Excavated soils will be characterized to determine final offsite disposition.	Excavation	Excavation is considered to be generally technically implementable with the exception of certain areas of the site that are more difficult to access with equipment and personnel due to terrain conditions.	Costs include equipment use and maintenance, material handling and transport, backfill materials and activities, characterization, and offsite transportation and disposal. Long term O&M costs are considered to be low.
4	Limestone and Organic Matter	5-Year amendment study tested combinations of lime slurry and organic matter, and tilling applications. Plant coverage, pH, and soil chemistry were monitored. Plant coverage, pH, and soil chemistry would be monitored post-amendment operations.	Soil Amendments	This technology has been demonstrated to be implementable via the ongoing amendment study. This technology may be less implementable at certain areas of the site that are not readily accessible by equipment.	Costs are moderate and include purchase and transport of amendments to the site, and equipment for application. Minimal soil handling is required as soils generally remain in place, minimizing transportation and disposal costs. Long term O&M costs are considered low to moderate.
4b	Tilling or Ripping	5-Year amendment study tested combinations of lime slurry and organic matter, and tilling applications. Plant coverage, pH, and soil chemistry were monitored. Tilling de-compacts soil and provides additional dilution of metals and has potential to raise alkaline pH conditions to more neutral pH conditions pending existing pH levels within the soil treatment area being tilled. Plant coverage, pH, and soil chemistry would be monitored post-tilling operations.	Soil Amendments	This technology has been demonstrated to be implementable via the 5-year amendment study. This technology may be less implementable at certain areas of the site that are not readily accessible by equipment for tilling (i.e., areas with too steep of a slope, > 13%).	Tilling includes increased efforts and costs as compared to adding lime and organic matter without tilling. However, minimal soil handling is required as soils generally remain in place, minimizing transportation and disposal costs compared to excavation and soil covers. Long term O&M costs are considered low to moderate.
4c	Ferricyanide	Ferricyanide is mixed into the impacted soils to reduce the copper bioavailability to site receptors.	Soil Amendments	This technology is considered to be implementable, based on the ongoing soil amendment study which has demonstrated that soil mixing is feasible. A pilot treatability study would be required to determine the most effective soil mixing technique.	Cost will depend on soil mixing technique determined and market value of the amendment materials. The overall technology is considered to have moderate costs and is comparable to the costs of the lime and O&M amendments being tested as part of the Amendment Study. Long term O&M are considered low to moderate.
4d1	Use of Chelating Agent: Soil Washing (Ex-Situ)	Ex-situ soil washing includes removal of contaminants from soil using separation methodologies, including "washing" the soils with a detergent and/or chelating agent solution. The resulting clean portions of soil can be returned to the site for reuse and the resulting wash mixture and contaminated soil fines would be characterized for final disposition.	Soil Amendments	A pilot treatability study would be required to determine implementability, including a consideration of the accessibility of soil washing materials (e.g., water source) and equipment to areas of remediation.	Ex-Situ soil washing is labor intensive and requires a high level of soil handling (excavation, washing, soil replacement). Requires disposal of used wash solution and potentially portions of soil. Cost savings include utilizing remediated soil for onsite reuse, reducing imported backfill required.
4d2	Use of Chelating Agent: Soil Washing (In-Situ)	In-situ soil washing includes introducing a chelating agent into the soil. The objective of the chelating agent is to mobilize the copper within the soil column. The copper becomes soluble within the groundwater and the groundwater is subsequently extracted for treatment and/or disposal with a groundwater extraction system.	Soil Amendments	Due to the large areas needing remediation and lack of current infrastructure to support a groundwater extraction system, this is considered to have low implementability. However, a pilot treatability study would be required to determine implementability.	In-Situ soil washing is labor intensive and requires the introduction of the chelating agent into the soil. Installation of groundwater extraction system, and extracted groundwater treatment and/or disposal. Costs for construction and future O&M activities are considered very high.



TABLE 5-1  
SOIL REMEDIATION TECHNOLOGIES

FREEMONT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELT TAILING SOILS FEASIBILITY STUDY

NO.	REMEDIAL TECHNOLOGY	DESCRIPTION	PRELIMINARY SCREENING		CONCLUSION	
			EFFECTIVENESS	IMPLEMENTABILITY		
4d3	Use of Chelating Agent: Phytoremediation	Phytoremediation includes the planting of specified seeds and plants to uptake the contaminants through the root system into the plant. The plant is then harvested and disposed of, thereby removing the contaminants from the site. A chelating agent can be applied to the soil to increase the rate of contaminant uptake into the plants.	is potentially effective at removing copper from soil. Phytoremediation of contaminants in seeds and vegetation has potential to restrict future intrusive activities. Contaminant removal is directly dependent on the success of the plants and the rate of contaminant uptake.	A pilot treatability study would be required to determine implementation factors to evaluate implementability of phytoremediation. Factors to be considered include vegetation, harvesting and replanting schedule, chelating agent, and rate of contaminant uptake.	Initial costs are considered low to moderate and include costs of seeds/plants, chelating agent, and planting. O&M is considered high due to the long maintenance of the plants, and disposal of the plant waste. Costs are considered moderate to high as compared to other technologies.	Not being retained for further evaluation.
5	Soil Cover	A soil cover involves isolating impacted soils from potential site receptors. The soil cover would consist of clean soil and would be constructed over soils with copper levels above the site RAC to serve as a physical barrier between contaminated soils and site receptors. In addition, the cover could be constructed in areas of exposed bedrock to create pre-existing habitat for site receptors.	Containment  Is effective at protecting site receptors from underlying contaminants. Contaminants will remain onsite requiring institutional controls to restrict future intrusive activity practices. Maintenance of BMPs is considered critical to maintain the integrity and effectiveness of the soil cover.	Is considered to be generally technically implementable with the exception of certain areas of the site that are more difficult to access with equipment and personnel due to terrain conditions.	Costs are considered moderate as compared to soil excavation and the soil amendment technologies. There are higher initial costs associated with soil handling and placement of the cover. However, there are no costs associated with excavation or soil amending. O&M costs related to maintaining the cover are considered moderate.	Being retained for evaluation as a part of a remedial alternative.
5b	Impermeable Cover	An impermeable cover, such as asphalt or concrete, placed over targeted impacted soils would greatly minimize or eliminate the potential for direct contact of site receptors with impacted site soils.	An impermeable cover would be effective at reducing exposure of site receptors to impacted soils. However, an impermeable cover would eliminate the vegetative landscape, potential for future grazing, and would impact surface water infiltration and stormwater runoff patterns.	Installation of an impermeable cover is generally considered implementable with exception of areas that are too steep and/or areas with varying topography that can not be graded to conditions that could accommodate the cover.	Costs include surface preparation and grading, and installation of the impermeable cover, including base materials. Costs are considered to be moderate as compared to other soil remedial technologies. Long term O&M costs are considered low.	Not being retained for further evaluation.
6	Surface Soil Controls: Phytostabilization	Surface soil controls includes stabilizing surface soils to prevent or cover an area of impacted soils and provide erosion control. Phytostabilization consists of revegetating the impacted areas with plant species targeted at increasing long term stabilization as compared to existing vegetative conditions.	Phytostabilization has the potential to be effective at stabilizing the soils and reducing transport of impacted soils. However, this technology used alone would not be effective at reducing site contaminant levels and/or exposure to site receptors. This could be effectively be used in conjunction with other remedial technologies.	Revegetating soils with species that are targeted at soil stabilization is generally considered implementable at the site although it will take time to reduce species that do not support the phytoremediating species (due to slopes, percent soil coverage, and soil conditions). Plant species and potential locations would be determined during the remedial design process.	Costs include the costs of the individual plants, planting, and O&M. O&M is considered moderate to high. Site preparation and O&M is considered to be low once the vegetation has been established. Costs are considered moderate compared to other technologies.	Being retained for evaluation as a part of a remedial alternative.
7	Phytoremediation	Phytoremediation consists of planting vegetation that can uptake the contaminants located in the soil and subsequently remove them from the site. Phytoremediation technologies that are able to bioaccumulate and/or degrade the site contaminants.	Phytoremediation is potentially effective at remediating site contaminants. Contaminant reduction would take several years to achieve and would not immediately reduce potential exposure to site receptors. A preliminary remedial design evaluation would be required to determine if the naturally existing conditions at the site (e.g., humidity, access to groundwater, soil conditions) could support phytoremediating plant species.	Assuming that the naturally occurring site conditions can support phytoremediating plant species, phytoremediation is generally considered implementable at the site although it will take time to reduce species that do not support the phytoremediating species (due to slopes, percent soil coverage, access to groundwater, and soil conditions) where it is not implementable.	Costs include the cost of the individual plants/trees, planting, and O&M. O&M is considered moderate to high, as remediation of the site contaminants is directly dependent on the success of the plants/trees to thrive over an extended period of time. Costs are considered moderate to high as compared to other technologies.	Not being retained for further evaluation
8	Electrokinetic Remediation	Electrokinetics is based on the principle that when direct current (DC) is passed through contaminated soil, certain (negatively charged) types of contaminants will migrate through the soil pore water to a place where they can be removed. This alternative uses electrode assemblies that are installed in the ground in a square array and connected to a DC voltage power supply. Each electrode assembly contains water, a pump, an electrode, and various controllers and sensors. The outer casing of the electrodes is made of porous ceramic through which the electrical current and contaminants pass. A vacuum is applied to the casings which keeps the water inside the assembly from flowing out and saturating adjacent soils.	Overall effectiveness of using electrokinetic remediation would need to be determined by conducting a pilot and potentially a bench scale treatability study. Factors that would be evaluated during the treatability studies would include, but would not be limited to, voltage rates, effectiveness in soils with low moisture content, and the increased acidity of the treated soil.	Electrokinetic remediation has not been tested at the site, so currently is considered low for implementability. A pilot treatability study would need to be conducted to determine if the low soil moisture content in the soil would support the removal of copper and if the increased soil acidity would impact vegetative species.	The cost associated with electrokinetic remediation is assumed to be moderate to high. Factors such as the spacing and number of electrodes would be determined during a pilot treatability study and could greatly impact the costs for this technology. Therefore, without the results of a pilot study the overall cost of implementation cannot be determined.	Not being retained for further evaluation

TABLE 5-2  
SURFACE WATER REMEDIAL TECHNOLOGIES

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

NO.	REMEDIAL TECHNOLOGY	DESCRIPTION	PRELIMINARY SCREENING		COST	CONCLUSION
			EFFECTIVENESS	IMPLEMENTABILITY		
1	No Action	No further active response actions for surface water.	Contaminants will naturally attenuate over time. Does not provide additional mechanisms to prevent contaminant exposure to site receptors.	Is considered implementable.	There are no costs associated with no action.	Being retained for baseline comparison of remedial technologies.
2	Monitoring	No further active response actions for surface water. Monitoring to be conducted to prove the occurrence of natural attenuation.	Contaminants will naturally attenuate over time. Does not provide additional mechanisms to prevent or reduce contaminant exposure to site receptors.	Is considered implementable.	Costs are associated with types of monitoring (quantitative and/or qualitative) and monitoring durations selected.	Being retained for evaluation as a part of a remedial alternative.
3	Excavation (In-Drainage, Upland, or Stock Ponds)	Excavation includes removing impacted in-drainage sediments and/or upland soils determined to be contributing to surface water impacts.	Is considered highly effective at reducing the presence of site contaminants.	Is considered to be generally technically implementable with the exception of certain areas of the site that are more difficult to access with equipment and personnel due to terrain conditions.	Costs include equipment use and maintenance, material handling and transport, and characterization and final disposition. Long term O&M costs related to BMPs are considered to be low to moderate.	Being retained for evaluation as a part of a remedial alternative.
4	In-Stream Removal of Suspended Sediments	Consists of construction of settling basins within the stream drainage areas to allow for sediments to descend to bottom of the pool and accumulate. Accumulated impacted sediments would be removed by mechanical methods.	Is considered to be effective at capturing contaminated sediments. Settling pools currently exist on site and have demonstrated effectiveness in capturing contaminated sediments	Some portions of drainage areas may be restricted from construction of pools due to equipment accessibility restrictions.	Costs are considered to be moderate during construction of the settling pools. Long term O&M costs are considered to be moderate to high as compared to excavation and surface water treatment.	Being retained for evaluation as a part of a remedial alternative.
5	Limestone Treatment	Consists of installation of features, such as limestone, within the surface water drainage pathway. As the surface water passes over the feature (e.g. limestone) the pH is elevated, subsequently making the contaminants (copper) less bioavailable to site receptors.	Is considered effective at raising surface water pH as long as surface water makes contact with the limestone features.	Is considered to be technically implementable. Some portions of drainage areas may be restricted from installation of limestone features due to equipment accessibility restrictions.	Costs are considered to be high during construction of the limestone features. Long term O&M costs are considered to be low compared to excavation and in-stream removal of suspended sediments.	Being retained for evaluation as a part of a remedial alternative.
6	Alkaline Washing	Consists of insertion of alkaline fluid into the sediments and banks of the drainage. The alkaline wash will raise pH and sequester metals in the sediments, subsequently reducing mobilization of the contaminants (copper) to surface water.	Is considered effective at raising pH and lowering metals concentrations in surface water as long as the majority of metals loading is coming from sediments.	Is considered to have low implementability. There are large infrastructure requirements and some portions of the drainage area may be restricted due to equipment accessibility restrictions.	Costs are considered to be high both during the construction and O&M phases when compared to other remedial options.	Not being retained for evaluation as a part of a remedial alternative.
7	Sediment and Erosion Control	Consists of construction of ditches, berms, and other controls (e.g., sedimentation controls) to prevent surface water run-on to the impacted areas and runoff from the impacted areas.	Is considered effective at minimizing the transport of contaminants to and from the drainage areas and surface water.	Is considered to be technically implementable in the drainage area and upland portions of the Site with exception to areas that may be inaccessible to equipment and personnel.	Construction costs are considered to be low to moderate. Long term O&M costs associated with maintenance, repair, and replacement of controls are considered to be low to moderate compared to other surface water remedial technologies.	Being retained as a component of remedial alternatives.

**TABLE 5-3  
SUMMARY OF RETAINED REMEDIAL TECHNOLOGIES**

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS IU FEASIBILITY STUDY

Soils	Surface Water
No Action	No Action
Monitoring	Monitoring
Excavation and Reuse	Excavation (In-drainage, Upland, or Stock Ponds)
Excavation and Disposal	In-stream Removal of Suspended Sediments
Soil Amendments (Limestone, Organic Material)	Limestone Treatment
Soil Amendments - Tilling	Sediment and Erosion Control
Soil Amendments (Ferrihydrite)	
Soil Cover	
Surface Soil Controls - Phytostabilization	

TABLE 6-1  
 REMEDIAL ALTERNATIVES FOR CONSIDERATION - SOILS (TOTAL METALS)

FREEPORT-MCMORAN CHINO MINES COMPANY  
 VANADIUM, NEW MEXICO  
 SMELTER/TAILING SOILS IU FEASIBILITY STUDY

MEDIA	TECHNOLOGY TYPE	ALTERNATIVES				
		1 - NO ACTION	2 - MONITORING	3 - EXCAVATION AND MONITORING	4 - SOIL COVER AND MONITORING	5 - SOIL AMENDMENTS/TILLING AND MONITORING
Upland Soils (Total Metals)	No Action	X				
	Monitoring		X	X	X	X
	Lime and/or Organic Material					X
	Tilling					X
	Soil Cover				X	

TABLE 6-2  
 REMEDIAL ALTERNATIVES FOR CONSIDERATION - SOILS (pCu)

FREEPORT-MCMORAN CHINO MINES COMPANY  
 VANADIUM, NEW MEXICO  
 SMELTER/TAILING SOILS IU FEASIBILITY STUDY

MEDIA	TECHNOLOGY TYPE	ALTERNATIVES										
		1 - NO ACTION	2 - MONITORING	3 - EXCAVATION (REUSE OR DISPOSAL) AND MONITORING	4 - SOIL AMENDMENTS (LIME), TILLING AND MONITORING	5 - TILLING AND MONITORING	6 - SOIL AMENDMENTS (FERRIHYDRITE), TILLING AND MONITORING	7 - SOIL COVER AND MONITORING	8 - SURFACE SOIL CONTROLS - PHYTOSTABILIZATION			
Upland Soils (pCu)	No Action	X										
	Monitoring		X		X	X				X		X
	Excavation and Reuse or Disposal			X								
	Soil Amendments				X				X			
	Tilling				X			X				
	Soil Cover											
	Phytostabilization										X	

TABLE 6-3  
 REMEDIAL ALTERNATIVES FOR CONSIDERATION - SURFACE WATER DRAINAGES

FREEPORT-MCMORAN CHINO MINES COMPANY  
 VANADIUM, NEW MEXICO  
 SMELTER/TAILING SOILS IU FEASIBILITY STUDY

MEDIA	TECHNOLOGY TYPE	ALTERNATIVES					
		1 - NO ACTION	2 - MONITORING	3 - EXCAVATION AND MONITORING	4 - EXCAVATION, SEDIMENT CONTROL, LIMESTONE TREATMENT, AND MONITORING	5 - LIMESTONE TREATMENT AND MONITORING	6 - SEDIMENT CONTROL, EROSION CONTROL, AND MONITORING
SURFACE WATER	No Action	X					
	Monitoring		X	X	X	X	X
	Excavation (In-drainage, Upland, or Stock Ponds)			X	X		
	In-stream Removal of Suspended Sediments				X		
	Limestone Treatment				X	X	
	Sediment and Erosion Control						X

TABLE 6-4  
 REMEDIAL ALTERNATIVES FOR CONSIDERATION - SURFACE WATER STOCK TANKS

FREEPORT-MCMORAN CHINO MINES COMPANY  
 VANADIUM, NEW MEXICO  
 SMELTER/TAILING SOILS IU FEASIBILITY STUDY

MEDIA	TECHNOLOGY TYPE	ALTERNATIVES		
		1 - NO ACTION	2 - MONITORING	3 - EXCAVATION AND MONITORING
STOCK TANKS	No Action	X		
	Monitoring		X	X
	Excavation (In-drainage, Upland, or Stock Ponds)			X
	Removal of Suspended Sediments			
	Limestone Treatment			
	Sediment and Erosion Control			

TABLE 7-1  
DETAILED EVALUATION OF THE REMEDIAL ALTERNATIVES - SOILS (TOTAL METALS)

FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELT/TAILING SOILS IU FEASIBILITY STUDY PROPOSAL

ALT DESCRIPTION	OVERALL PROTECTION	COMPLIANCE WITH ARARs	LONG-TERM EFFECTIVENESS AND PERMANENCE	REDUCTION OF TOXICITY, MOBILITY OR VOLUME	SHORT-TERM EFFECTIVENESS	IMPLEMENTABILITY	COST	GREEN REMEDIATION
1 No Action	Does not provide any form of additional overall protection of human health and the environment. However, no risk currently exists to human health and thus no action is protective. Does not meet the RAO for wildlife because does not include monitoring to ensure conditions continue to be protective.	There are no chemical-specific ARARs for surface soil except pre-FS RAC, which would not be met for wildlife because some areas occur with copper > 1100 mg/kg, and pre-FS RAC states they require monitoring. However, the portion of the pre-FS RAC that requires active remediation when copper > 1600 mg/kg is met, with no action because no exposure units exceeded 1600 mg/kg.	Not effective in the long-term or permanent if changes occur in exposure	Does not reduce toxicity, mobility or volume.	Is effective in the short-term because studies show current conditions are protective.	Is implementable at the site, as no action is required.	There are no costs associated with this alternative.	No additional resources would be required to implement this alternative.
2 Monitoring	Does not provide any additional form of overall protection of human health and the environment beyond current conditions. However, the ability of the Monitoring alternative to satisfy this threshold criteria is high and achievable because it meets the requirement of the protective pre-FS RAC for small ground feeding bird that the acres with copper > 1100 mg/kg will be monitored, and the requirement that no acres should be below 1600 mg/kg copper, and none are. The monitoring ensures results for soils between 1100 mg/kg and 1600 mg/kg copper do not change and adversely affect the small ground feeding bird. Monitoring could trigger a remedial alternative over time.	There are no chemical-specific ARARs for surface soil except for the pre-FS RAC, which would be met. This alternative would meet the threshold criterion of soils between 1100 and 1600 mg/kg being monitored. It also meets all part of the pre-FS RAC that requires all soils to be ≤ 1600 mg/kg because no exposure units exceed 1600 mg/kg.	Not effective in the long-term or permanent if exposure increases. The magnitude of risk would be consistent with current conditions unless monitoring results in the implementation of additional remedial actions. However, natural attenuation may lower copper in soils due to erosion and clean windblown deposition over time, and monitoring would increase confidence in permanence of the protection. This alternative affords the opportunity to continue data collection and to take additional action if necessary.	Does not reduce toxicity, mobility, or volume.	Is effective in the short-term because studies show current conditions are protective.	This remedy is implementable at the site with limited additional effort.	Capital Costs: \$148K O&M (NPV, 30 yr): \$893K TIC: \$1.04MM	Emissions and fuel use would be limited to those associated with light vehicles and sample shipping/analysis. This alternative would save existing vegetation from destruction.



TABLE 7-2  
 DETAILED EVALUATION OF THE REMEDIAL ALTERNATIVES - SOILS (pCu)  
 FREEPORT-MCMORAN CHINO MINES COMPANY  
 VANADUM, NEW MEXICO  
 SMELT TAILING SOILS III FEASIBILITY STUDY PROPOSAL

ALTERNATIVE	DESCRIPTION	OVERALL PROTECTION	COMPLIANCE WITH ARARs	LONG-TERM EFFECTIVENESS AND PERMANENCE	REDUCTION OF TOXICITY, MOBILITY OR VOLUME	SHORT-TERM EFFECTIVENESS	IMPLEMENTABILITY	COST	GREEN REMEDIATION
1	No Action	Does not provide any form of overall protection for ecological receptors beyond current conditions. Some acres were carried forward for consideration for remediation in areas where the estimated mean pCu is below the soil-specific PEL, therefore the No Action alternative does not meet the RCO. The No Action alternative does not meet the RCO, leading to increased protection. However, the No Action alternative may be more protective than active remedies for the vegetation community in areas with a pCu below soil-specific PELs that fall within a pCu range of uncertain benefit. The uncertain range is between the selected PELs and an alternative PEL derived by averaging rangeland condition with wildlife habitat endpoints, which is between 4.98 and 4.6 for flat rocky soil, 4.4 and 3.98 for bedrock, and 4.11 and 2.97 for slope soils. The lower end of the range may reduce the overall net ecological risk, which is the tradeoff between the benefit of reducing pCu exposure for plants and the harm from active remedies that temporarily destroy the vegetation community. The PEL used may be biased high (see Section 7.3.1 of the text for detailed explanation), which would mean the No Action alternative could be more protective than remediation in some of the acres carried forward for consideration for remediation.	There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Because there are no action-specific or location-specific ARARs for this alternative. Compliance with the ARARs is low because some acres do not meet the pre-FS RAC.	The magnitude of residual risk (i.e., potential for unacceptable risk to vegetation over the long term) would be consistent with current conditions. There could be changes to pCu in surface soil due to changes in pH over time, either improving or not improving with no action. Therefore, it is uncertain if this alternative is acceptable for long term effectiveness or permanence.	Does not reduce toxicity, mobility or volume because no treatment would be implemented.	More effective in the short term at protecting the plant community than active remedies, which temporarily destroy or damage the plant community. Less effective at protecting the community in long-term where pCu exposure is the main risk, depending on the clean-up level selected. Because no actions would be implemented, there would be no risks to the plant community or impacts to the environment during implementation.	Is implementable at the site, as no action is required.	There are no costs associated with this alternative.	No additional resources would be required to implement this alternative.
2	Monitoring	Does not provide any form of overall protection for ecological receptors beyond current conditions, and does not meet the RAOs. However, monitoring provides more data to evaluate need for remediation and may trigger a remedial alternative over time. However, as with the No Action alternative, Monitoring may be more protective for the vegetation community in areas with pCu below soil-specific PELs that fall within a pCu range of uncertain benefit. The uncertain range is between the selected PELs and an alternative PEL derived by averaging rangeland condition with wildlife habitat endpoints, which is between 4.98 and 4.6 for flat rocky soil, 4.4 and 3.98 for bedrock, and 4.11 and 2.97 for slope soils. The lower end of the range may reduce the overall net ecological risk, which is the tradeoff between reducing pCu exposure for plants and active remedies that destroy the vegetation community during implementation. The PEL used may be biased high (see Section 7.3.1 of the text for detailed explanation). Monitoring could be more protective than remediation in some of the acres carried forward for consideration for remediation.	There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Because no actions would be taken there are no action-specific or location-specific ARARs for this alternative. The likelihood of compliance with the ARARs on an ongoing basis is low.	The magnitude of residual risk (i.e., potential for unacceptable risk to vegetation over the long term) would be consistent with current conditions unless additional remedial actions. Natural attenuation may lower copper in soils due to erosion and availability to wildlife due to clean soil/dust deposition over time. It may not reduce pCu below vegetation risk values; therefore, it is uncertain if this alternative is acceptable for long term effectiveness or permanence. However, this alternative affords the opportunity to continue data collection and to take additional action if necessary.	Does not reduce toxicity, mobility or volume.	The alternative is more effective in the short term than active remedies because active remedies, if implemented, do not destroy the plant community being protected. No actions would be implemented so there would be no risks to the community or workers or impacts to the environment during implementation.	This remedy is implementable at the site with limited additional effort. It would not limit further actions at STS/U; however, a monitoring plan would need to be coordinated with NMED.	Capital Costs: \$280K O&M (NPV, 30 yr): \$1.6MM TIC: \$1.9MM	Emissions and fuel use would be limited to those associated with light vehicles and sample shipping/analysis. This alternative would save existing vegetation from temporary destruction.

TABLE 7-2  
**DETAILED EVALUATION OF THE REMEDIAL ALTERNATIVES - SOILS (pCu)**  
 FREEPORT-MCMORAN CHINO MINES COMPANY  
 VANADUM, NEW MEXICO  
 SMELT TAILING SOILS IN FEASIBILITY STUDY PROPOSAL

ALTERNATIVE	DESCRIPTION	OVERALL PROTECTION	COMPLIANCE WITH ARARs	LONG-TERM EFFECTIVENESS AND PERMANENCE	REDUCTION OF TOXICITY, MOBILITY OR VOLUME	SHORT-TERM EFFECTIVENESS	IMPLEMENTABILITY	COST	GREEN REMEDIATION
3	Excavation (Reuse or Disposal) and Monitoring	This alternative would remove soils with copper and acid concentrations at levels that could pose a potential risk to vegetation from pCu. Once the ecosystem re-establishes, eventually maturing to its previous successional stage, Alternative 3 would meet the threshold criterion of overall protection of human health and the environment. As discussed in Section 7.3.1.1, the PEL for flat rocky soils may be biased high. The beneficial effects of the Alternative on the vegetation community at pCu between 4.6 and the PEL (4.98) in flat rocky soils and in areas of bedrock (pCu between 4.4 and 3.83) and slope soil (pCu between 4.11 and 2.97) is uncertain. Since the Alternative will temporarily destroy the entire community, even the more sensitive species being protected may not return for an undetermined period of time. For the areas of bedrock and slope soil, the Alternative is likely to provide the highest level of overall protection for the vegetation community in the STSU. No areas of bedrock soils with pCu < 3.83 or slope soils with pCu < 2.97, areas with more certain benefits from this Alternative, were identified.	There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Action-specific and location-specific ARARs would be met by remedial design. Alternative 3 is expected to meet the threshold criterion of compliance with ARARs if careful attention is given to restoring the soil productivity and plant community after soil removal.	This alternative would provide a high level of long-term effectiveness and permanence for soil categories that can be excavated (i.e., flat rocky soils) after species return after an undetermined but potentially long time frame. Removal is a proven and reliable method of preventing the potential for contact with materials of concern. Monitoring of erosion from removal areas and soil sampling would be required to ensure that recontamination did not occur. Excavation will temporarily destroy the entire vegetation community. Because the beneficial effects of excavation on the vegetation community at pCu between 4.6 and the PEL (4.98) for flat rocky soils and in areas of bedrock excavation in those areas may be lower than in the areas with pCu < 4.6 in flat rocky soils.	This alternative would not meet the criteria of the statutory preference for treatment. Alternative 3 does not reduce the toxicity, mobility, or volume because no treatment would be implemented; only contaminated soil will be removed.	Alternative 3 would be ineffective at achieving the goal of protecting the vegetation community, wildlife habitat, and rangeland in the short term. Excavation would destroy the entire community, including both less sensitive and more sensitive species, and many species may not return for an extended period of time. When active remediation occurs at < PELs, the entire community is affected by the remediation. While the Alternative will provide better conditions for more sensitive species, the short term effectiveness may be limited because all species are simultaneously destroyed during implementation. Also, Alternative 3 would entail significant earthwork which would result in a risk to workers due to physical hazards, particularly working on steep and potentially unstable slopes. Placing heavy equipment on slopes to degrade them could increase the risk of slope failure. Site preparation, soil excavation, transport and disposal onsite are significant construction activities that will be taken to the operational area for subsequent reuse or disposal. In addition, the use of heavy equipment for grading and earthwork would release diesel emissions into the environment.	This alternative is considered moderately implementable. Excavation would not be implementable in areas of bedrock outcrops or steep slopes. Of the 377 acres brought forward for consideration of remediation, 47 are on bedrock outcrops or steep slopes. The presence of these two physical factors limits the implementability of this alternative in these areas. Soil excavation, transportation and onsite disposal can be achieved in flat rocky soils with standard equipment and materials. Site preparation will include building new roads for haulage. Excavated soil, greater than the PELs, would either be taken to the operational area for subsequent reuse or disposal in West Stockpile for recycling (disposal option).	Capital Costs: \$8.9MM O&M (NPV, 30 yr): \$2.2MM TIC: \$11.1MM	Additional emissions and fuel use would be required to excavate the soils. The excavated soils would be reused or recycled. This alternative would result in the temporary destruction of existing vegetation. While the duration of the disturbance to vegetation is unknown, it could last for an extended period of time.
4	Soil Amendments (Lime), Tilling and Monitoring	The ability of this alternative to satisfy this criterion is high on relatively level habitats, particularly if on poor rangeland. Tilling has been shown to be effective on the site (e.g. for rippled roads, see Appendix A). Copper present in surface soils would be tilled and mixed deeper into the soil along with bringing higher pH soils to the surface. Tilling also provides the benefit of breaking up the soil to allow plant growth. Additionally, tilling would likely increase the rate of natural attenuation, though natural attenuation rates for pCu have not been determined. The addition of lime would also increase pCu in soils and reduce the bioavailability of any remaining copper, although the application of lime alone (where tilling is impractical) would have a low to moderate ability to satisfy this criterion. As discussed in Section 7.3.1.1, the PEL for flat rocky soils may be biased high. The beneficial effects of the Alternative on the vegetation community at pCu between 4.6 and the PEL (4.98) in flat rocky soils and in areas of bedrock (pCu between 4.4 and 3.83) and slope soil (pCu between 4.11 and 2.97) is uncertain. Since the Alternative will temporarily destroy the entire community, even the more sensitive species being protected may not return for an undetermined period of time. For the areas of bedrock and slope soil, the Alternative is likely to provide the highest level of overall protection for the vegetation community in the STSU. No areas of bedrock soils with pCu < 3.83 or slope soils with pCu < 2.97, areas with more certain benefits from this Alternative, were identified.	There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Action-specific and location-specific ARARs would be met by remedial design. Alternative 4 would likely meet the threshold criterion of compliance with ARARs if tilling is included. It is less certain on steeper slopes where lime has runoff and has been less effective.	The long-term effectiveness of this alternative is considered moderate for amendments for improving pCu but high for tilling for improving the plant community in flat rocky soils. Lime (calcium carbonate) has been shown to increase pCu values at the site but does not always increase pCu values at the site to pre-FS RAC levels. Tilling of flat rocky soils < PELs would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species would not return for an undetermined but potentially long time frame. Alternative 4 would provide a high level of long-term effectiveness and permanence only after species return. Its effectiveness in bedrock and on steep slopes is low, given tilling can not be done in bedrock and filling equipment is difficult to use on steep slopes. Tilling would be effective on slope soils but the effectiveness is uncertain because lime is easily washed off with rainfall in these areas. Additional applications of soil amendments may be required in bedrock outcrops or on steep slopes due to transportation of material from storm water flows.	This alternative would mix the upper soil layers with underlying soils, increasing pCu in combination with adding lime, tilling would reduce toxicity and amount of acid in the soils by reducing the bioavailability and mobility and thus toxicity of the copper.	This alternative would not be considered moderately effective in the short-term. Mixing of surface soils with deeper soil layers resulting from copper, and improve conditions for plant growth. However, it may take several years for the soil amendments to increase the pCu values. Also, tilling would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species would not return for an undetermined but potentially long time frame. When active remediation occurs on areas with pCu < PEL levels for the specific soil type, the whole community is affected by the remedy. While there is some gain to improve soil conditions for more sensitive species, there is limited short term effectiveness because all species are simultaneously destroyed. If tilling is included, lime is only applied (no tilling) as it was on steeper slopes in the amendment study (only up to 13% slope). It may still adversely impact the plant community in the short term (observed during amendment study), and only slightly increase the pCu.	The implementability of soil amendments is considered high compared to the other alternatives. If tilling is included, it is an alternative that could be applied to areas with steep slopes and bedrock outcrop areas, which cannot be addressed by physical means. However, additional applications of soil amendments may be required in the future in bedrock outcrops or steep slopes due to the transportation of amendments in storm water. Tilling would be most readily implementable as it could not be used in areas of bedrock outcrops or steep slopes but is effective in flat, rocky soils.	Capital Costs: \$4.2MM O&M (NPV, 30 yr): \$2.2MM TIC: \$6.4MM	Emissions and fuel use would be lower than Alternative 3 but higher than Alternative 5. This alternative would result in the temporary destruction of existing vegetation. While the duration of the disturbance to vegetation is unknown it could last for an extended period of time.

TABLE 7-2  
**DETAILED EVALUATION OF THE REMEDIAL ALTERNATIVES - SOILS (pCu)**  
 FREEPORT-MCMORAN CHINO MINES COMPANY  
 VANADIUM, NEW MEXICO  
 SMELT TAILING SOILS USE FEASIBILITY STUDY PROPOSAL

ALTERNATIVE	DESCRIPTION	OVERALL PROTECTION	COMPLIANCE WITH ARARs	LONG-TERM EFFECTIVENESS AND PERMANENCE	REDUCTION OF TOXICITY, MOBILITY OR VOLUME	SHORT-TERM EFFECTIVENESS	IMPLEMENTABILITY	COST	GREEN REMEDIATION
5	Tilling and Monitoring	The ability of this alternative to satisfy this criterion is high for slopes that are not too steep. Tilling has been shown to be effective on the site (e.g. for rippled roads, see Appendix A). Copper present in surface soils would be filled and mixed deeper into the soil along with bringing higher pH soils to the surface. Tilling also provides the benefit of breaking up the soil to allow plant growth. Additionally, tilling would increase the rate of natural attenuation, though natural attenuation rates for pCu have not been determined. However, the net environmental benefit may be lower than alternatives without destructive remediation in areas with pCu below soil-specific PELs, as tilling temporarily destroys the entire community, including both less sensitive species that are thriving and more sensitive species that would benefit, and many species, including more sensitive species being protected, would not return for an potentially extended period of time.	There are no chemical-specific ARARs for surface soil except for the pre-FS RAC. Action-specific and location-specific ARARs would be met by remedial design. Alternative 5 may meet the threshold criterion of compliance with ARARs, as the design is achieving the desired pCu (< the PELs) is moderate with tilling but the likelihood of achieving a more diverse community with greater plant cover is high.	The long-term effectiveness of this alternative is considered moderate to high for pCu but high for tilling for improving the plant community in flat rocky soils. Tilling would destroy the entire community, including both less sensitive and more sensitive species until the species return. Because may be for an extended period of time. Because the beneficial effects of tilling in flat rocky soils (the soil type in which tilling is most effective) at pCu between 4.6 and the PEL (4.98) is uncertain, the long-term effectiveness of tilling in those areas where the beneficial effects are not expected to occur. Alternative 5 would provide a high level of long-term effectiveness and permanence only after species return. Its effectiveness in bedrock and on steep slopes is low, given tilling can not be done in bedrock and tilling equipment is difficult to use on steep slopes. Monitoring of erosion from removal areas and soil sampling would be required to ensure that recontamination did not occur.	This alternative would mix the upper soil layers with underlying soils, decreasing concentrations of copper, increasing pH, and reducing the amount of acid that would be produced. Mobility also may be reduced.	However, it may take several years for the tilling to increase the pCu values. That said, as described above, tilling at pCu at PELs specific to soil type would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species, including the more sensitive species, may not return for an extended period of time. When active remedy occurs, the whole community is affected by the remedy. While there is some gain to improve soil conditions for more sensitive species, there is no short term effectiveness because all species are simultaneously affected.	The implementability of this alternative is considered moderate. It has the same limitations as Alternative 3, soil tilling can be achieved with standard equipment and materials and with no extensive site preparation but could not be used in areas of bedrock outcrops or steep slopes.	Capital Costs: \$3.7MM O&M (NPV, 30 yr): \$2.2MM TIC: \$5.9MM	Emissions and fuel use would be lower than Alternative 3. A medium-sized dozer with a ripper could implement this alternative with relatively low emissions and less support resources. This alternative would result in destruction of existing vegetation.
6	Soil Amendments (Ferryhyrite), Tilling and Monitoring	The ability of this alternative to satisfy this criterion is possibly high similar to lime and tilling described in Alternative 4 but the effectiveness of ferryhyrite has not been tested on the site. Tilling has been shown to be effective on the site (e.g. for rippled roads, see Appendix A). Copper present in surface soils would be filled and mixed deeper into the soil along with bringing higher pH soils to the surface. Tilling also provides the benefit of breaking up the soil to allow plant growth. Additionally, tilling would increase the rate of natural attenuation, though natural attenuation rates for pCu have not been determined. However, the net environmental benefit may be lower than alternatives without destructive remediation in areas with pCu below soil-specific PELs, as tilling temporarily destroys the entire community, including both less sensitive species that are thriving and more sensitive species that would benefit, and many species, including more sensitive species being protected, would not return for an potentially extended period of time.	There are no chemical-specific ARARs for surface soil except for the pre-FS RAC, which might be met. Action-specific and location-specific ARARs would be met by remedial design. The likelihood that Alternative 6 would meet the threshold criterion of compliance with ARARs is considered moderate for ferryhyrite alone and high if combined with tilling.	The long-term effectiveness of this alternative is considered moderate to high to increase pCu values. Ferryhyrite has been shown to reduce free copper activity at similar sites. Generally, pCu values are higher at sites with an increased rate of natural attenuation resulting from tilling. However, although less mobile than lime, additional applications of ferryhyrite amendments may be required in bedrock outcrops or on steep slopes due to transportation of material from storm water flows. Tilling, implementable in flat rocky soils, would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species may not return for an extended period of time. Alternative 6 would provide a high level of long-term effectiveness and permanence only after species return. The effectiveness of tilling in bedrock and on steep slopes is low, given tilling can not be done in bedrock and tilling equipment is difficult to use on steep slopes. Monitoring of erosion from removal areas and soil sampling would be required to ensure that recontamination did not occur.	Tilling in this alternative would mix the upper soil layers with underlying soils, increasing pCu and thus reducing toxicity and amount of acid in the soils. Mobility of acid would be reduced by the addition of ferryhyrite.	The alternative would not be considered effective in the short-term. Mixing of surface soils with deeper soil layers resulting from tilling with amendments is expected to increase pH, decrease soil copper, and improve conditions for plant growth. However, it may take several years for the soil amendments to increase the pCu values and the increase may be small if copper concentrations are high. Tilling of soils with pCu at < PELs of the specific soil type would destroy the entire community, including both less sensitive and more sensitive species, and many species may not return for an extended period of time. While there is some gain to improve soil conditions for more sensitive species, there is no short term effectiveness because all species are simultaneously affected.	The implementability of soil amendments is considered high. It is an alternative that could be applied to areas with steep slopes and bedrock outcrop areas, which cannot be addressed by physical means. Additional applications of soil amendments may be required in the future in bedrock outcrops or steep slopes due to the transportation of amendments in storm water. Tilling is considered moderately implementable as it could not be used in areas of bedrock outcrops or steep slopes.	Capital Costs \$4.2MM O&M (NPV, 30 yr): \$2.2MM TIC: \$6.4MM	Emissions and fuel use would be lower than Alternative 3 but higher than Alternative 5. This alternative would result in temporary destruction of existing vegetation. While disturbance to vegetation is unknown it could last for an extended period of time.

**TABLE 7-2  
DETAILED EVALUATION OF THE REMEDIAL ALTERNATIVES - SOILS (pCu)**  
FREEPORT-MCMORAN CHINO MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELT/TAILING SOILS LIQUID FEASIBILITY STUDY PROPOSAL

ALTERNATIVE	DESCRIPTION	OVERALL PROTECTION	COMPLIANCE WITH ARARs	LONG-TERM EFFECTIVENESS AND PERMANENCE	REDUCTION OF TOXICITY, MOBILITY OR VOLUME	SHORT-TERM EFFECTIVENESS	IMPLEMENTABILITY	COST	GREEN REMEDIATION
7	Soil Cover and Monitoring	Soil cover would reduce but not eliminate plant contact with the contaminants. Also, the pathways would only be closed if the soil cover remains in place. The soil cover could be breached by digging, construction, or similar activities, or by plants with deep roots. This could potentially pose a risk to the vegetation. Additionally, in soils with pCu between the selected PELs and the lower PELs identified when thresholds for rangeland wildlife habitat are averaged (the "uncertain" pCu effect range), the net environmental benefit may be lower than alternatives without destructive remediation, as soil cover destroys the entire community, including both less sensitive species that are thriving and more sensitive species that would benefit. Many species, even the more sensitive species being protected, may not return for an extended period of time.	These are no chemical-specific ARARs for surface soils except for the pre-FS BAC. Action-specific and location-specific ARARs would be met by remedial design. Alternative 7 would meet the threshold criterion of compliance with ARARs.	A soil cover would require long-term operations, maintenance, and monitoring to continue to be effective. Alternative 7 is considered a moderately effective solution. Soils with pCu of < PELs would temporarily destroy the entire community, including both less sensitive and more sensitive species, and many species might not return for an extended period of time. Alternative 7 would provide a high level of long-term effectiveness and permanence only after species return.	This alternative would not meet the statutory preference for reducing the toxicity, mobility, or volume because no treatment would be implemented; only contaminated soil would be covered.	This alternative is not considered to be effective in the short-term. It would address all of the pCu areas < PELs; however, soil cover over acres below the PELs would destroy the entire community, including both less sensitive and more sensitive species, even after seedling for an extended period of time. When active remedy occurs in areas of < PEL of the specific soil type, the whole community is affected by the remedy. While there is some gain to improve soil conditions for more sensitive species, there is no short term effectiveness because all species are simultaneously affected. Vegetation in borrow areas would be adversely impacted in the short-term, too.	This alternative is considered moderately implementable. Unlike excavation, placing a soil cover could be implemented in areas that would be difficult to access, and might erode off bedrock, also. Covering bedrock on 47 acres would require large amounts of soil and large borrow areas. Of the 377 acres addressed for pCu exposure, 14 acres are on steep slopes. More maintenance is required with this option than for other alternatives.	Capital Costs: \$11.0MM O&M (NPV 30yr): \$2.2MM TIC: \$13.2MM	This alternative would result in similar emissions and fuel use when compared to alternatives 3. The area to be remediated would result in temporary destruction of existing vegetation. While the disturbance to vegetation is unknown it could last for an extended period of time.
8	Surface Soil Controls Phytostabilization	The overall protection of this alternative would be low. Contaminant reduction through limiting transport would take several years to achieve, and it would not reduce toxicity to the site vegetation it is designed to protect. Additionally, the net environmental benefit may be lower than alternatives without destructive remediation in areas in the uncertain pCu range, as the phytostabilization process destroys and replaces existing plants, including both less sensitive species that are thriving and more sensitive species that would benefit. Many desired species, even the more sensitive species being protected, would not return for decades.	These are no chemical-specific ARARs for surface soils except for the pre-FS BAC. Action-specific and location-specific ARARs would be met by remedial design. Alternative 8 would not meet the threshold criterion of compliance with ARARs because it is unlikely to meet the pre-FS RAC.	The long-term effectiveness for this alternative is considered to be very low compared to the other alternatives. Phytostabilization has the potential to be effective at stabilizing the soils and reducing transport of impacted soils. However, this technology used alone would not be effective at reducing site contaminant levels and/or exposure to the plants the remediation is supposed to protect.	This alternative would not meet the statutory preference for reducing the toxicity, mobility, or volume because no treatment would be implemented. This technology used alone would not be effective at reducing site contaminant levels and/or exposure to site receptors.	This alternative would be considered not to be effective in the short-term because of the length of time required for the native vegetation, destroyed during the remediation, to become established above. Phytostabilization would destroy many species in the community and many species may not return for an extended period of time. When this active remedy occurs at < PELs, species insensitive and sensitive to pCu are affected by the phytostabilization. While there is some gain to improve soil stability, there is no short-term effectiveness because the community is disrupted by the planting and remedial activities.	Revegetating soils with species that are targeted for soil stabilization is generally considered implementable at the Site. There may be certain areas of the site that may not support the phytostabilization species (due to slopes, percent soil coverage, and soil conditions). Plant species and potential locations would be determined during the remedial design process.	Capital Costs: \$4.1MM O&M (NPV 30 yr): \$2.5MM TIC: \$6.6MM	This alternative would result in lower emissions and fuel than all but Alternatives 1 and 2.

TABLE 7-4  
DETAILED EVALUATION OF THE REMEDIAL ALTERNATIVES - SURFACE WATER DRAINAGES

FREEMONT-MICROMAN CHINO MINES COMPANY  
SMELTING/TAILING SOILS USE FEASIBILITY STUDY PROPOSAL

ALTERNATIVE	DESCRIPTION	OVERALL PROTECTION	COMPLIANCE WITH ARARS	LONG-TERM EFFECTIVENESS AND PERMANENCE	REDUCTION OF TOXICITY, MOBILITY OR VOLUME	SHORT-TERM EFFECTIVENESS	IMPLEMENTABILITY	COST	GREEN REMEDIATION
1	No Action	Does not provide any additional form of overall protection of human health and the environment beyond current conditions. However, current exposure for all but the D3 drainage on average likely is low and protective overall, because only a small percentage of the samples in the drainages had an HQ > 1. Additionally, impacts to the environment are likely small because pool habitat is limited in these areas. However, this alternative does not address the uncertainty of the current conditions, and therefore does not meet the RAOs.	Alternative 1 is not compliant with ARARS because it is no action and there are some exceedances of copper based on NMAC §20.6.4.808 - 809 that need to be addressed with monitoring or other actions. Copper is the focus because it is main driver of risk, having significantly higher HQs than cadmium and lead, the other COCs for the drainages.	Not effective in the long-term or permanent because it does not address the exceedances for copper. However, natural attenuation may be happening, but rates, this alternative is not protective. This alternative has the benefit of preserving the existing upland vegetation and aquatic ecosystem, both of which would take decades to recover following the implementation of an active remedial alternative.	Alternative 1 does not reduce toxicity, mobility or volume because it does not involve any active remedy.	Not effective in the short-term because no action to address exceedances.	is implementable at the site, as no action is required.	There is no costs associated with this alternative.	No additional resources would be required to implement this alternative.
2	Monitoring	Does not provide any form of overall protection of human health and the environment. Does not meet the RAOs. However, current exposure for all but the D3 drainage on average is low and protective overall, because only a small percentage of the samples in the drainages had an HQ > 1. Additionally, impacts to the environment are likely small because pool habitat is limited in these areas. Monitoring may be protective because it provides more data to evaluate need for remediation and may trigger a remedial alternative over time.	The Monitoring alternative is not immediately compliant with ARARS or the pre-FS RAC for some point locations, but may be meeting it for drainages on average. Monitoring would allow for adaptability and may trigger remedial activities over time. In this case, the likelihood of compliance with ARARS is dependent on whether pre-FS RAC are met when evaluated across the drainages currently and dependent on the management of upstream sources, suspended sediment and dissolved metals concentrations, as well as overall water quality and attenuation rates.	The effectiveness of this remedial alternative in the long-term is limited to reduce concentrations to below aquatic life standards because it depends on natural attenuation, which appears to be occurring or may have already occurred for all but Drainage D3. However, it may be effective enough to address long-term risks if upstream sources of contaminants are addressed. Additionally, this alternative affords the opportunity to continue data collection and to take additional action if necessary.	Alternative 2 does not reduce toxicity, mobility or volume because it does not involve any active remedy.	Not effective in the short-term because no action to address exceedances. However, current conditions may be protective for all but Drainage D3. Monitoring in the short term of all the retained drainages may trigger an action if conditions warrant it.	This remedy is implementable at the site with limited additional effort. Long-term monitoring would be required to show improvements in water quality in the surface water drainages.	Capital Costs: \$20K O&M (NPV, 30 yr): \$293K TIC: \$313K	Emissions and fuel use would be limited to those associated with light vehicles and sample shipping/analysis. This alternative would save existing vegetation from destruction.
3	Excavation and Monitoring	The ability of this alternative to meet the overall protection of water quality is high to moderate. The overall protection is based on the removal of historical sediments, but also a reduction of potential upstream contributors. This alternative would destroy the existing upland and riparian vegetation and aquatic ecosystem initially, and the vegetation could take decades to recover following the implementation of an active remedial alternative. The pools are limited and dry in drier years, particularly in alluvial drainages, providing limited habitat and loss of riparian vegetation benefits supports this remedy only for Drainage D3 that has high HQs.	The likelihood of compliance with the ARARS on an ongoing basis is moderate to high with this alternative. This is the potential for reduced long-term effectiveness of this alternative if upstream contributions continue. Additional excavation may be required at a future time.	The effectiveness of this remedial alternative in the long-term is high to moderate to reduce concentrations to aquatic life standards. In recent efforts to re-sample, areas of concern were dry so long-term effectiveness at protecting aquatic biota could be low if frequently dry. Once excavation of historical sediments is complete, monitoring will be required to document that water quality criteria have been met. Long-term effectiveness of this alternative may be dependent on addressing upstream contributors (suspended sediments and storm water).	This alternative would not meet the statutory preference for the treatment. Alternative 3 does not reduce the toxicity, mobility or volume because no treatment would be implemented.	This alternative is considered moderately effective in the short-term. Surface water quality may take some time to improve. Normal risks associated with nuisance dust, excavation hazards, noise and traffic.	In specific surface water drainages this alternative will potentially be difficult to implement compared to other alternatives. Specifically in the Bolton Spring and Ash Spring corridor the excavation of historical sediments may do more harm than good to the ecological habitat. Additional permitting and coordination would be required compared to other alternatives. For other drainages the excavation of historical sediments is considered implementable. Some of the drainages may not be accessible to large equipment, due to steep slopes or tight access. However, overall this alternative is implementable at the site. No special equipment is required to complete this alternative. Equipment and materials to complete the job are readily available.	Capital Costs: \$569K O&M (NPV, 30 yr): \$368K TIC: \$958K	Additional emissions and fuel use would be required to excavate the drainages. The constrained environment would also reduce efficiency. This alternative would result in destruction of existing vegetation.

TABLE 7-4  
DETAILED EVALUATION OF THE REMEDIAL ALTERNATIVES - SURFACE WATER DRAINAGES

FREEMONT MICHIGAN CHINO MINES COMPANY  
SMELTER/TAILING SOILS III FEASIBILITY STUDY PROPOSAL

ALTERNATIVE	DESCRIPTION	OVERALL PROTECTION	COMPLIANCE WITH ARARS	LONG-TERM EFFECTIVENESS AND PERMANENCE	REDUCTION OF TOXICITY, MOBILITY OR VOLUME	SHORT-TERM EFFECTIVENESS	IMPLEMENTABILITY	COST	GREEN REMEDIATION
4	Instream Removal of Suspended Sediments and Monitoring	The ability of this alternative to meet the overall protection of water quality is moderate to high. The overall protection is based on the installation of limestone sedimentation basins within the drainage channels, thereby decreasing the bioavailability of copper in surface water. Long-term monitoring is suggested for the implementation of this alternative.	The likelihood of compliance with the ARARs on an ongoing basis is high, and potentially higher than Alternative 3. There is the potential for reduced long-term effectiveness of this alternative if upstream contributions continue. Additional excavation should be anticipated within the sedimentation basins in the future. The addition of limestone to treat bioavailability of copper increases the potential compliance with ARARs.	The effectiveness of this remedial alternative in the long-term is high to moderate to reduce concentrations to aquatic life standards. Once excavation of historical sediments with the installation is complete and sedimentation basins are installed to collect suspended sediment monitoring will be required to document that water quality criteria have been met. Overall long-term effectiveness will also depend on any additional potential upstream sources of contaminants. Chemical treatment of the surface would increase the potential effectiveness of this alternative compared to Alternative 3.	This alternative addresses toxicity by reducing the bioavailability of copper due to the presence of limestone in the sedimentation basins. It also reduces the mobility of the contaminants by removing them from contact with surface water. The volume of material will be reduced with this alternative and it will also prevent potential contamination of surface water due to the presence of historical sediments.	This alternative is considered moderately effective in the short-term. Surface water quality may take some time to improve, and building a road into these areas to install the sediment basins will destroy vegetation in the short-term. Normal risks will occur associated with nuisance dust, excavation hazards, noise, and traffic.	In specific surface water drainages this alternative will be potentially difficult to implement compared to other alternatives. Specifically, in the Bolton Spring and Ash Spring corridor, the placement of lime may do more harm than good to the ecological habitat. Some of the drainages may not be accessible to large equipment, due to steep slopes or tight access. These factors make the implementability of this alternative more challenging than other alternatives. No special equipment is required to complete this alternative. Equipment and materials to complete the job are readily available.	Capital Costs: \$2.4MM O&M (NPV, 30 yr): \$388K TIC: \$2.8MM	Emissions and fuel use would be higher than Alternative 3, as additional sediment control measures and treatment are added, both of which increase emissions and natural resource use through manufacturing, material hauling, and maintenance. This alternative would result in destruction of existing vegetation.
5	Limestone Treatment and Monitoring	The ability of this alternative to meet the overall protection of water quality is moderate to high. The overall protection is based on the installation of limestone in low-lying channels, thereby decreasing the bioavailability of copper in surface water. Long-term monitoring is suggested for the implementation of this alternative.	The likelihood of compliance with the ARARs on an ongoing basis is moderate to high, as lime would be placed in the stream channels without excavation in areas where velocities are low. The limestone would decrease the bioavailability of copper and increase the potential compliance with ARARs.	The effectiveness of this remedial alternative in the long-term is moderate to reduce concentrations to aquatic life standards, as this technology allows water to pass through the limestone without the benefit of basins to increase the resident time of the water. Similar to Alternative 4, overall long-term effectiveness will also depend on any additional potential upstream sources of contaminants.	This alternative addresses toxicity by reducing the bioavailability of copper due to the presence of limestone in the drainage channels.	This alternative is considered moderately effective in the short-term. Surface water quality may take some time to improve. Normal risks associated with nuisance dust, excavation hazards, noise, and traffic.	In specific surface water drainages this alternative will be potentially difficult to implement compared to other alternatives. Specifically, in the Bolton Spring and Ash Spring corridor, the placement of lime may do more harm than good to the ecological habitat. Some of the drainages may not be accessible to large equipment, due to steep slopes or tight access. These factors make the implementability of this alternative more challenging than other alternatives. No special equipment is required to complete this alternative. Equipment and materials to complete the job are readily available.	Capital Costs: \$8.7MM O&M (NPV, 30 yr): \$931K TIC: \$9.6MM	Emissions and fuel use would be about the same as Alternative 4, as excavation would not be required but more frequent replenishment of lime may be required. This alternative would result in destruction of existing vegetation.
6	Sediment Control, Erosion Control, and Monitoring	The ability of this alternative to meet the overall protection of water quality is low to moderate. The overall protection of this alternative is based on the control of suspended sediments and potentially dissolved concentrations in storm water. This alternative does not address historical sediments. The addition of limestone as part of the BMPs to treat dissolved/bioavailable copper would increase the potential compliance with ARARs.	The ability of this alternative to comply with this criterion is low to moderate.	The long-term effectiveness of this alternative is low to moderate. A significant amount of maintenance and monitoring will be required to implement a large enough sediment and erosion control program across all of the drainages. Potential excavation and maintenance of the BMPs would be required on a regular basis to keep this alternative effective.	This alternative (without limestone treatment) would not meet the statutory preference for treatment. Alternative 6 does not reduce the toxicity, mobility, or volume because no treatment would be implemented. If limestone or other material is used, toxicity and volume of the constituents may be reduced. Mobility of the constituents may be reduced.	This alternative would be low to moderately effective in the short-term. This alternative does not address the presence of historical sediments in the drainages. This alternative would take time for water quality to improve.	In specific surface water drainages, this alternative will be potentially difficult to implement compared to other alternatives. Some of the drainages may not be accessible to large equipment, due to steep slopes or tight access. Equipment and materials to complete the job are readily available.	Capital Costs: \$3.6MM O&M (NPV, 30 yr): \$560K TIC: \$4.2MM	This alternative would result in the highest emissions and fuel use of the five alternatives. Although excavation is removed as an activity, erosion control is added which comes with considerably more material use and maintenance. This alternative would result in destruction of existing vegetation.

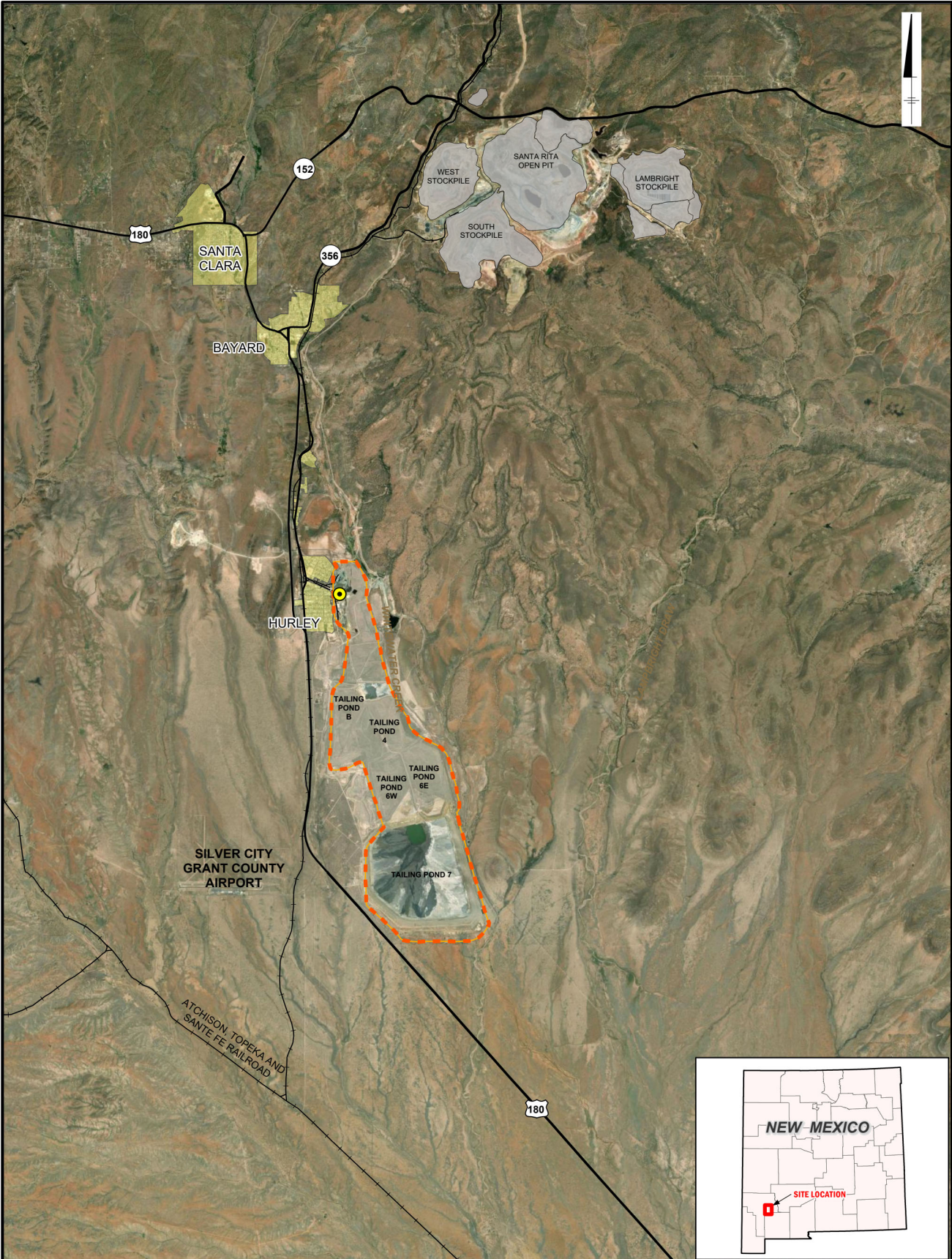
TABLE 7-5  
DETAILED EVALUATION OF THE REMEDIAL ALTERNATIVES - SURFACE WATER STOCK TANKS

FREEMPORT-MICHORANI COPPER MINES COMPANY  
VANADIUM, NEW MEXICO  
SMELTER/TAILING SOILS USE FEASIBILITY STUDY PROPOSAL

ALTERNATIVE	DESCRIPTION	OVERALL PROTECTION	COMPLIANCE WITH ARARS	LONG-TERM EFFECTIVENESS AND PERMANENCE	REDUCTION OF TOXICITY, MOBILITY OR VOLUME	SHORT-TERM EFFECTIVENESS	IMPLEMENTABILITY	COST	GREEN REMEDIATION
1	No Action	Does not provide any form of overall protection of human health and the environment. Does not meet the RAOs. However, current exposure likely is low and oscillated between 1 and 2 since 2007 and improvements were made in 2023 to provide for stable stormwater conveyance to the tank inlet. Because of these improvements, it is expected that water quality will continue to improve in stock tank 06 with time and may eventually meet the pre-FS RAC and RAOs.	Alternative 1 is not compliant with ARARs because it is no action and there are some exceedances of copper that need to be addressed with monitoring or other actions.	Alternative 1 is not effective in the long-term or permanent because it is no action and there are some exceedances for copper. However, natural attenuation may be happening, particularly following the 2023 rangeland improvements, but without monitoring to ascertain the rates, this alternative is not protective.	Alternative 1 does not reduce toxicity, mobility or volume because it is the no action alternative and does not provide treatment.	Alternative 1 is not effective in the short-term because it is no action and there are some exceedances for copper.	Is implementable at the site, as no action is required.	There is no costs associated with this alternative.	No additional resources would be required to implement this alternative.
2	Monitoring	Does not provide any form of overall protection of human health and the environment. Does not meet the RAOs. However, current exposure likely is low and oscillated between 1 and 2 since 2007 and improvements were made in 2023 to provide for stable stormwater conveyance to the tank inlet. Because of these improvements, it is expected that water quality will continue to improve in stock tank 06 with time and may eventually meet the pre-FS RAC and RAOs. Monitoring may trigger remedial activities over time.	This alternative is not immediately compliant with ARARs and there are some exceedances of copper. However, monitoring allows for adaptability and may trigger remedial activities over time. In this case, the likelihood of compliance with ARARs is dependent on the management of upstream sources, suspended sediment and dissolve metals concentrations, as well as overall water quality.	This alternative may be effective in the long-term for reducing concentrations to below aquatic life standards with natural attenuation after the rangeland improvements. This alternative may be effective enough to address long-term risks because upstream sources of contaminants are concurrently addressed with the 2023 rangeland improvements. Additionally, monitoring affords the opportunity to continue data collection and to take additional action if necessary.	This alternative does not reduce toxicity, mobility or volume because it is the no action alternative and does not provide treatment.	This alternative may be effective in the short term, despite some exceedances for copper. The rangeland improvements conducted in 2023 may have already improved water quality in the short term, but the monitoring program is needed to establish if this occurred.	This remedy is implementable at the site with limited additional effort.	Capital Costs: \$14K O&M (NPV, 30 yr): \$243K TIC: \$258K	Emissions and fuel use would be limited to those associated with light vehicles and sample shipping/analysis.
3	Excavation and Monitoring	Removal of historical sediments from stock tanks would address a major contributor to aquatic risk; however, there is still the potential for long-term risk to aquatic receptors due to upstream contribution from either suspended sediments or dissolve metals concentrations in surface water. Improvements were made in 2023 to provide for stable stormwater conveyance to the tank inlet. Because of these improvements, it is expected that water quality will continue to improve with time with or without active remediation.	The likelihood of compliance with the ARARs on an ongoing basis is moderate to high with this alternative. There is the potential for reduced long-term effectiveness of this alternative if upstream contributions continue. Additional excavation may be required at a future time.	The effectiveness of this remedial alternative in the long-term is high to moderate to reduce concentrations to below aquatic life standards. Once excavation of historical sediments is complete, monitoring will be required to document that water quality criteria have been met.	This alternative would not meet the statutory preference for treatment. Alternative 6 does not reduce the toxicity, mobility, or volume because no treatment would be implemented.	This alternative is considered effective in the short-term. Normal risks associated with nuisance dust, excavation hazards, noise, and traffic. Coordination may need to be made to provide alternative water sources to tanks owners during construction.	Based on the location and accessibility of stock tank 6, this alternative is implementable at the site. Equipment and materials to complete the job are readily available.	Capital Costs: \$80K O&M (NPV, 30 yr): \$245K TIC: \$325K	Additional emissions and fuel use would be required to excavate the material from the stock tanks and haul it to the operation area. The material would be reused and recycled.

# Figures

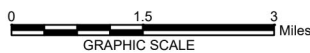




**LEGEND:**

- Location of former Hurley Smelter
- Railroad
- Major Roads
- Stockpiles
- Local City
- Operational Boundary of Smelter and Tailings Ponds

Service Layer Credits: ESRI World Imagery:  
 Earthstar Geographics

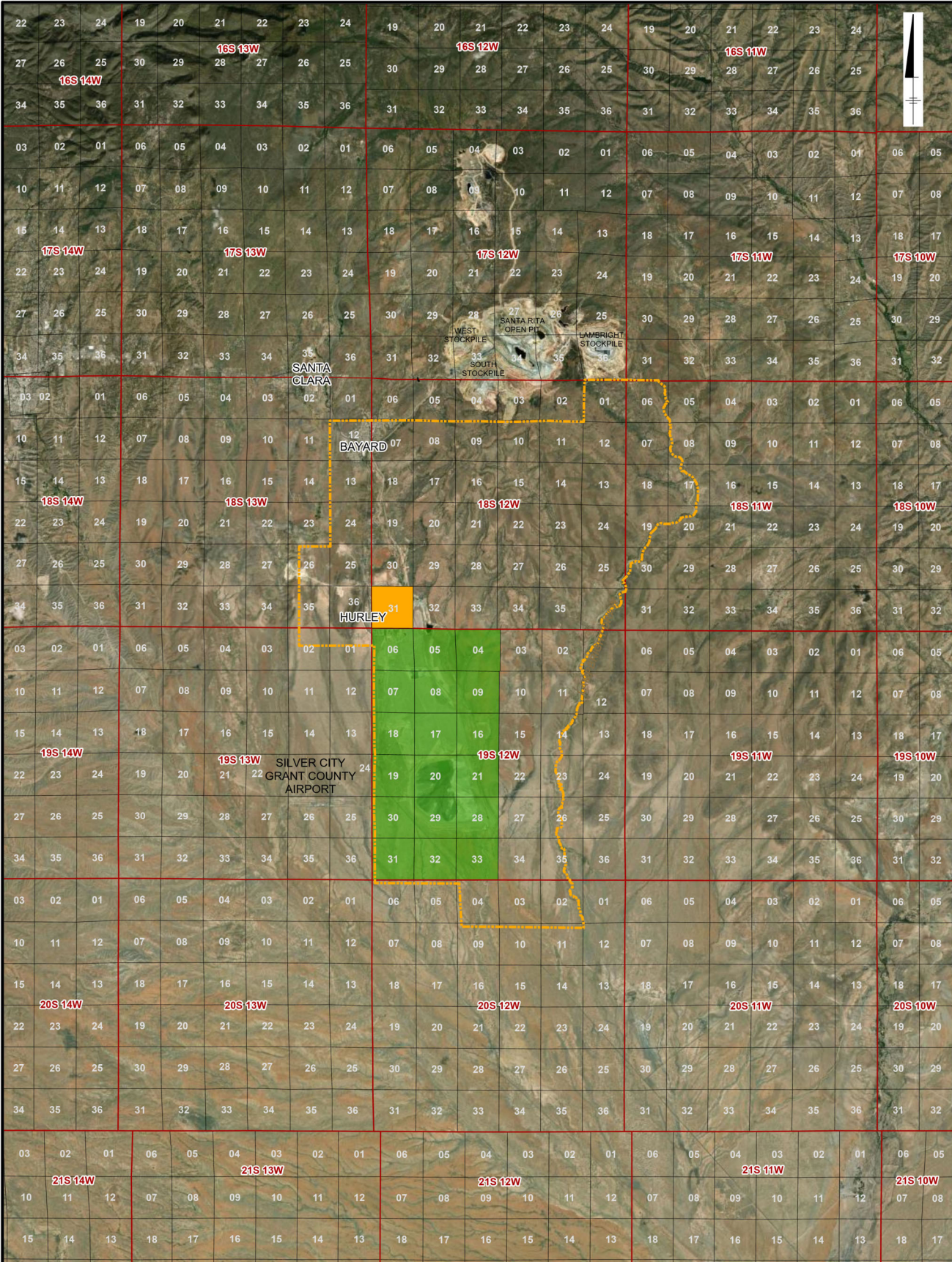


FREEPORT-MCMORAN CHINO MINES COMPANY  
 VANADIUM, NEW MEXICO  
**SMELTER/TAILINGS SOILS IU FS**

**SITE OVERVIEW**

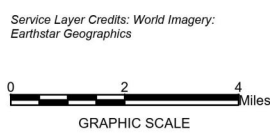


FIGURE  
**1-1**



**LEGEND:**

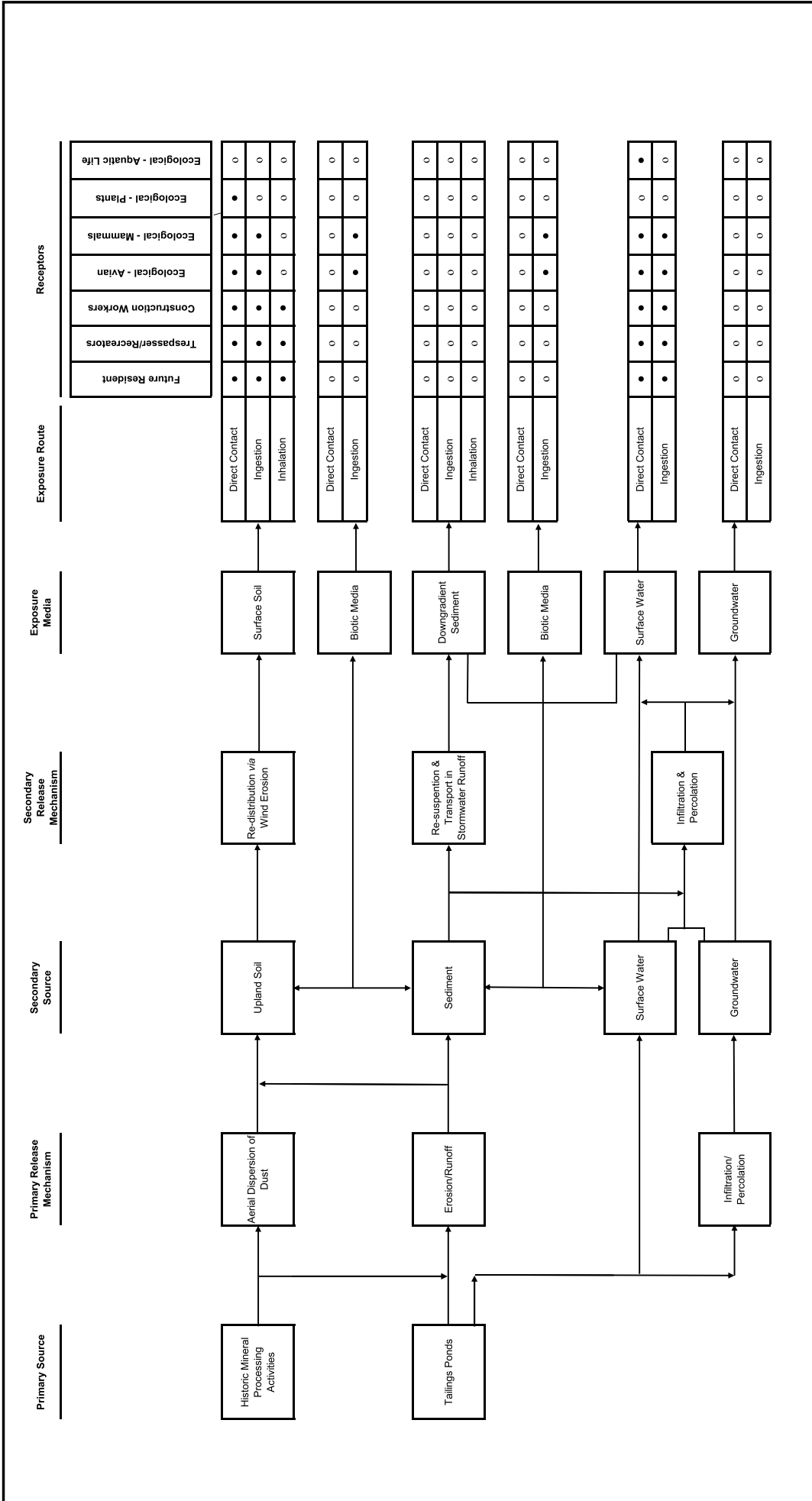
- STSIU Study Boundary
- Smelter Investigation Unit Section
- Tailing Impacted Soils Investigation Unit Sections
- Township-Range Boundaries
- Section Boundaries



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STSIU STUDY BOUNDARY  
 AND INVESTIGATION UNITS

FIGURE  
**2-1**



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CONCEPTUAL SITE MODEL

● Potentially complete exposure pathway  
 ○ Incomplete, or potentially complete but considered insignificant exposure pathway

**LEGEND & NOTES**

- EXISTING TAILINGS
- GRAVEL ROAD
- PAVED ROAD
- BUILDING
- RECLAMATION AREAS
- BORROW
- COVER
- VEGETATION

REVISIONS SUPPLIED BY CHINO MINE

DATE: 01/20/2017 TOPOGRAPHY

SCALE: 1:1000

CONTAINER INTERNAL TO PLAN

CONTOUR INTERVAL: 5 FEET

GRID COORDINATE SYSTEM: UTM

**REVISIONS**

#	DESCRIPTION	DATE
1	ADDED BORROW AREAS	01/20/2017
2	ADDED COVER AREAS	01/20/2017
3	ADDED VEGETATION AREAS	01/20/2017
4	ADDED BORROW AREAS	01/20/2017
5	ADDED COVER AREAS	01/20/2017
6	ADDED VEGETATION AREAS	01/20/2017

DATE:	06/26/24
PROJECT:	2009-03-001
TASK NUMBER:	02
DRAWN BY:	SW
CHECKED BY:	SW
DESIGNED BY:	SW
SCHEDULED BY:	JC

CHINO MINE RECLAMATION

**SMA COVER PLAN**

SHEET NUMBER: 23

REVISION NUMBER:

**TELESTO**

11100 S. UNIVERSITY AVENUE

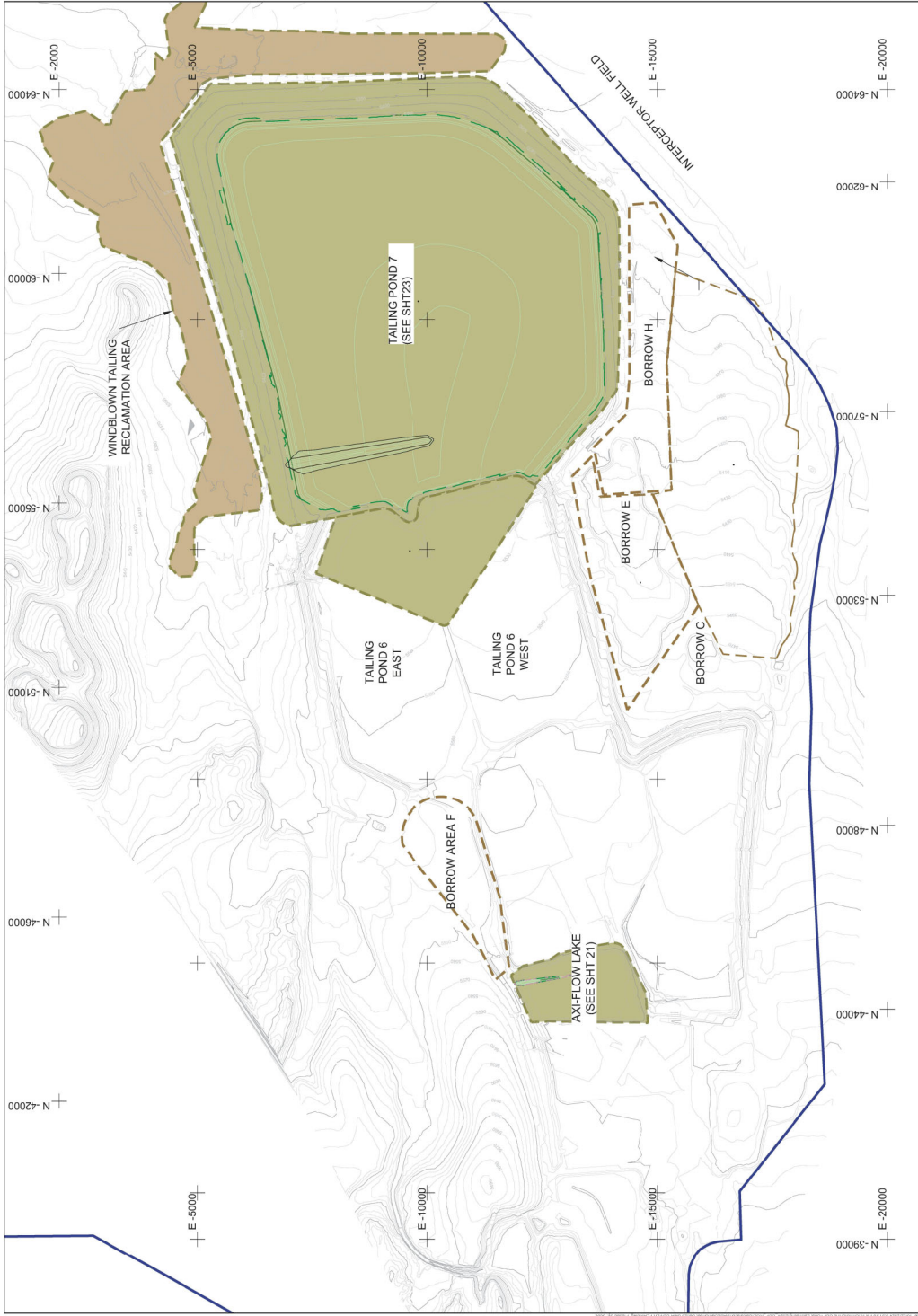
PHOENIX, AZ 85086

TELEPHONE: 480.461.4400

FAX: 480.461.4401

WEBSITE: [www.telesto.com](http://www.telesto.com)

PROJECT: FREEPORT-MCMORAN



NOTE:  
 1. THE BROWN AREA IS RECENT WINDBLOWN TAILING AREA AFFECTED BY TP-7 AND EXCLUDED FROM THIS FEASIBILITY STUDY

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 SMELTER/TAILINGS SOILS IU FS



**CURRENT EXTENT OF WINDBLOWN TAILINGS ASSOCIATED WITH TP-7 AND INCLUDED UNDER THE CLOSURE-CLOSEOUT PLAN FOR TP-7**

**ARCADIS**

FIGURE  
**3-2a**

**NOTES:**

AERIAL IMAGERY: CHINO MINES  
 SATELLITE SURVEY, APRIL 21, 2024

 WILDLIFE  
 WATER MANAGEMENT



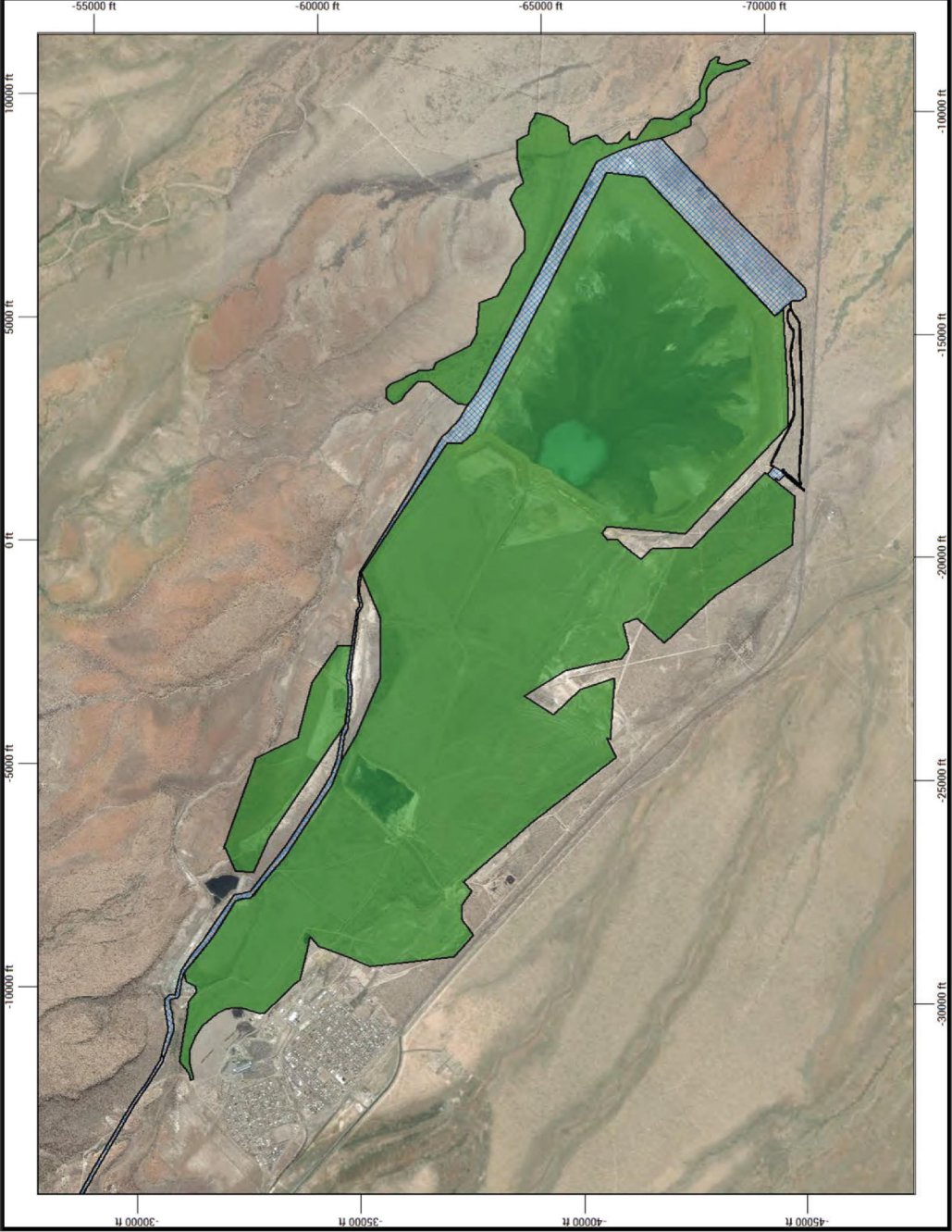


 SCALE IN FEET  
 COORDINATE SYSTEM  
 CHINO LOCAL MINE

FIGURE 12  
 SMA PMLU

PREPARED BY: **TELESTO**  
 SOLUTIONS INCORPORATED

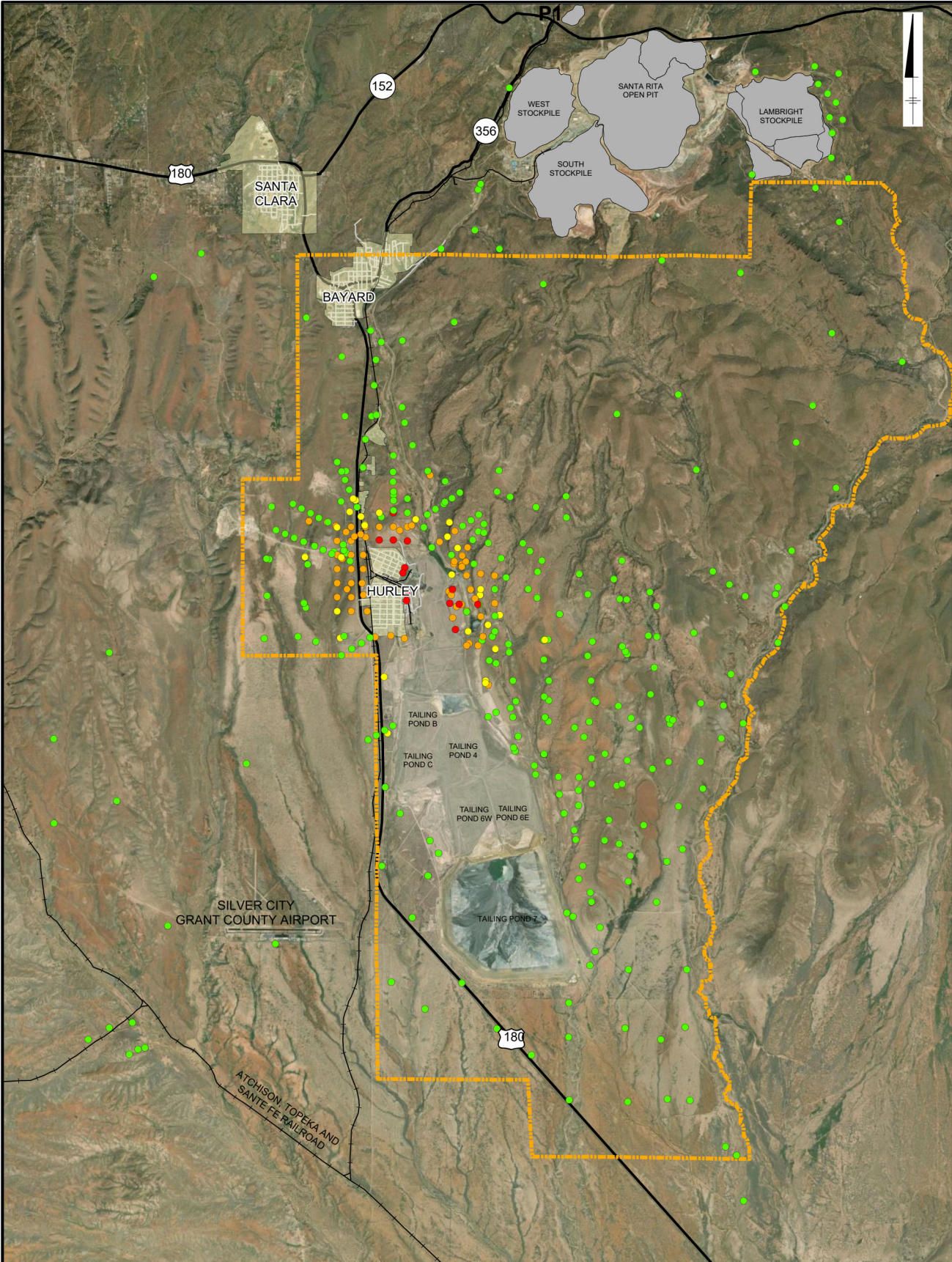
PREPARED FOR: **FREEPORT-McMORAN**



FREEPORT-MCMORAN CHINO MINES COMPANY  
 VANADIUM, NEW MEXICO  
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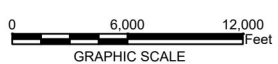
**CLOSURE CLOSEOUT PLAN'S DESIGNATION OF  
 AREAS FOR WILDLIFE AND WATER MANAGEMENT  
 THAT INCLUDES RECENT WINDBLOWN TAILING AREA**

**ARCADIS** | FIGURE 3-2b



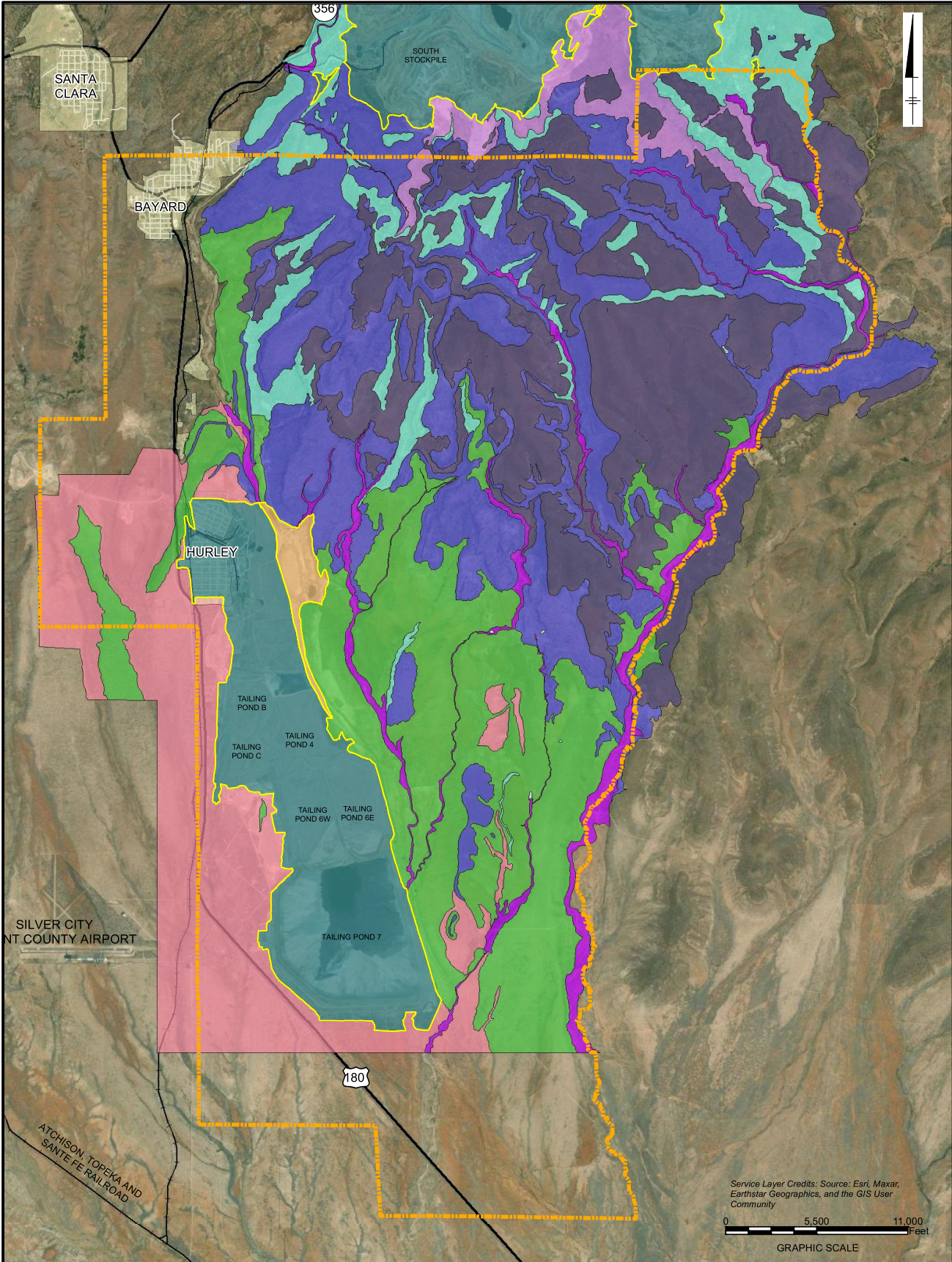
- LEGEND:**
- |   |   |
|---|---|
| <span style="color: green;">●</span> <1,100         | <span style="background-color: gray; border: 1px solid black; display: inline-block; width: 15px; height: 10px;"></span> Stockpiles   |
| <span style="color: yellow;">●</span> 1,100 - 1,600 | <span style="background-color: yellow; border: 1px solid black; display: inline-block; width: 15px; height: 10px;"></span> Local City |
| <span style="color: orange;">●</span> 1,600 - 5,000 | <span style="border-bottom: 2px dashed orange; display: inline-block; width: 20px;"></span> STSIU Boundary                            |
| <span style="color: red;">●</span> >5,000           |   |

Service Layer Credits: World Imagery: Earthstar Geographics



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 SMELTER/TAILINGS SOILS IU FS

**COPPER SAMPLE LOCATIONS AND CONCENTRATIONS IN SOIL FOR PRE-IRA SAMPLES**

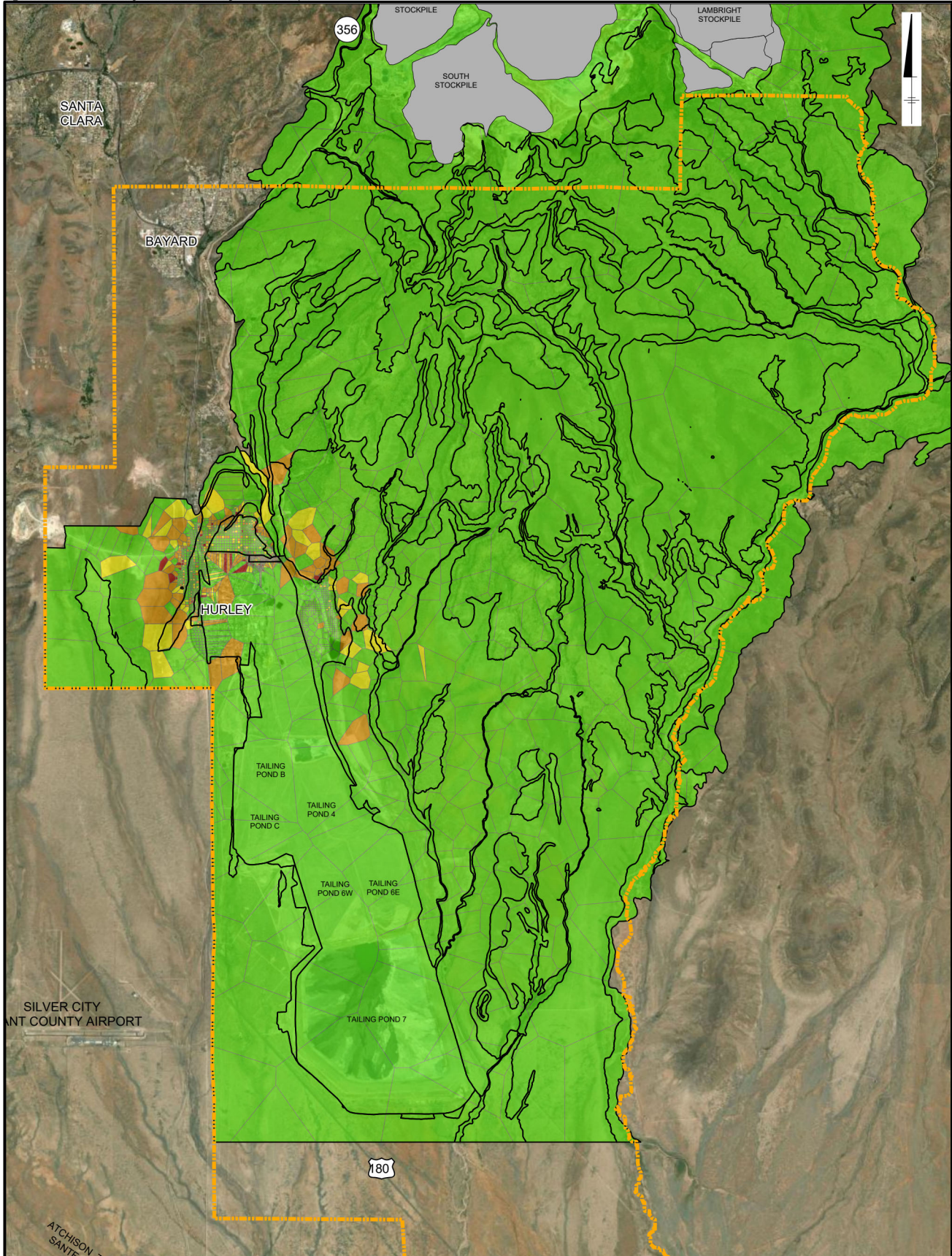


**LEGEND:**

- STSIU Boundary
- Mine Facilities Alliances**
- Mine Facilities/Other
- Mine Facilities/Urban
- Vegetation Alliances**
- Alligator Juniper-Oak Woodland Alliance
- Alligator Juniper-Oak/Grama Woodland Alliance
- Fluvial Forest and Shrubland Alliance
- Mesquite/Mixed Grama Shrubland Alliance
- Mixed-Grama Herbaceous Alliance
- Mountain Mahogany Shrubland Alliance
- Not Classified
- Ponderosa Pine-Oak Forest Alliance

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**VEGETATION ALLIANCES**



**LEGEND:**

- |   |   |
|---|---|
| <p><b>Copper Concentration (mg/kg)</b></p> <ul style="list-style-type: none"> <li><span style="display: inline-block; width: 15px; height: 10px; background-color: #90EE90; border: 1px solid black; margin-right: 5px;"></span> &lt; 1,100</li> <li><span style="display: inline-block; width: 15px; height: 10px; background-color: #FFFF00; border: 1px solid black; margin-right: 5px;"></span> 1,100 - 1,600</li> <li><span style="display: inline-block; width: 15px; height: 10px; background-color: #FFA500; border: 1px solid black; margin-right: 5px;"></span> 1,600 - 3,000</li> <li><span style="display: inline-block; width: 15px; height: 10px; background-color: #C00000; border: 1px solid black; margin-right: 5px;"></span> &gt; 3,000</li> </ul> | <ul style="list-style-type: none"> <li><span style="display: inline-block; width: 15px; height: 10px; border-bottom: 2px dashed orange; margin-right: 5px;"></span> Vegetation Alliance Polygons</li> <li><span style="display: inline-block; width: 15px; height: 10px; border-bottom: 2px dashed orange; margin-right: 5px;"></span> STSIU Boundary</li> <li><span style="display: inline-block; width: 15px; height: 10px; background-color: #A9A9A9; border: 1px solid black; margin-right: 5px;"></span> Stockpiles</li> </ul> |
|---|---|

Service Layer Credits: World Imagery: Earthstar  
 Geographics

0 5,000 10,000  
 Feet

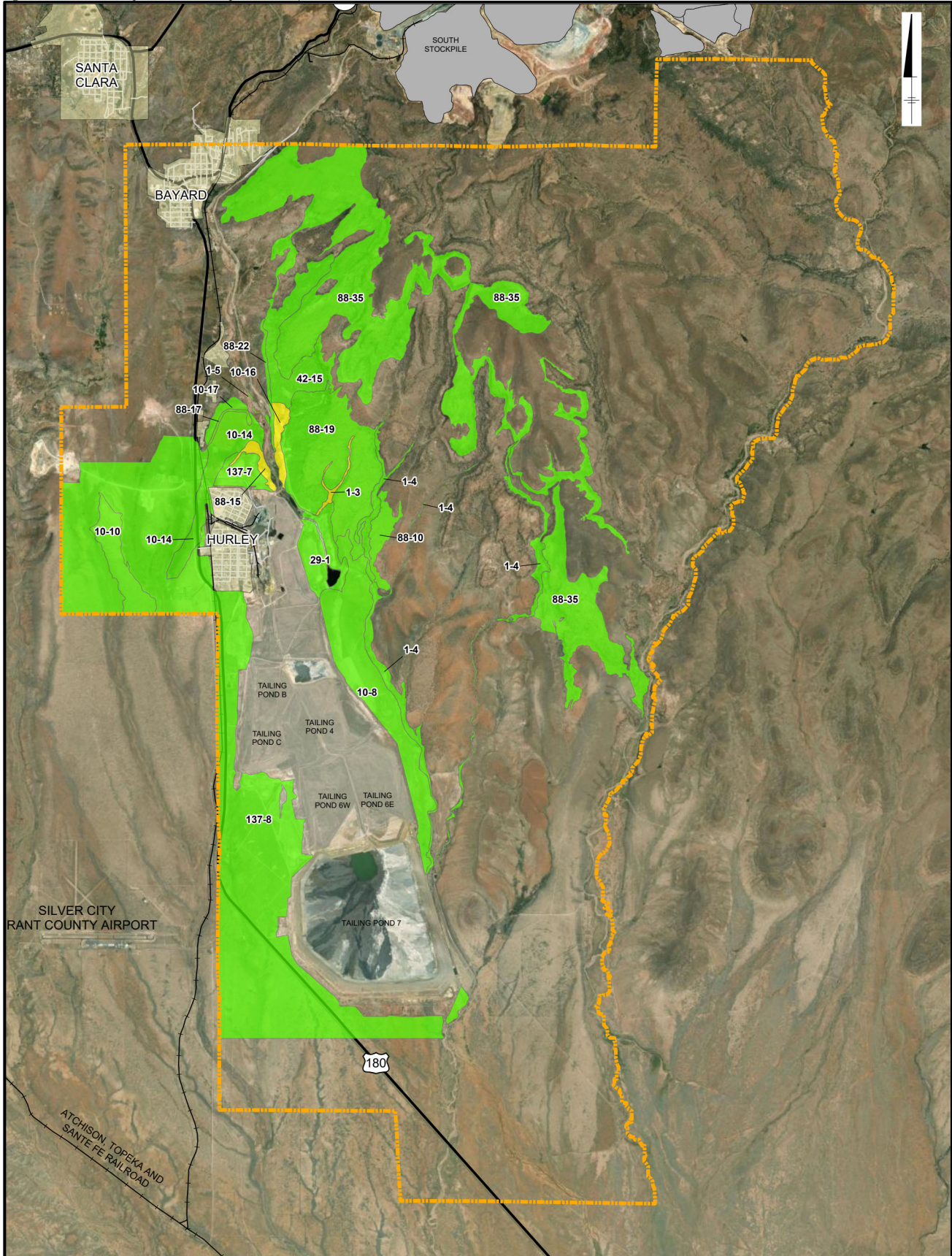
GRAPHIC SCALE

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 VANADIUM, NEW MEXICO  
 SMELTER/TAILINGS SOILS IU FS

**COPPER THIESSENS WITH  
 VEGETATION BOUNDARIES**

**ARCADIS** | **FIGURE 3-4**

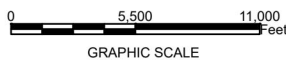




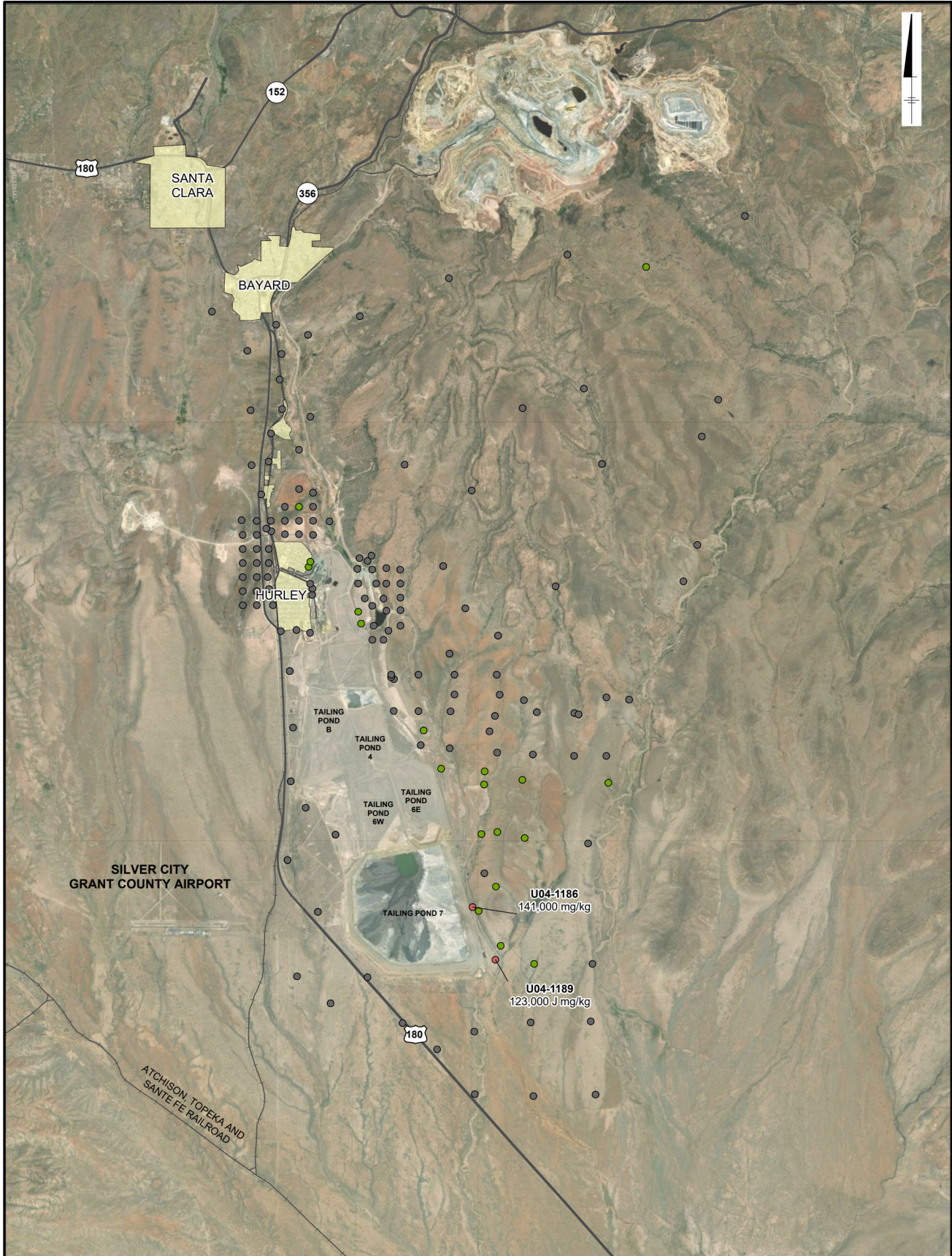
**LEGEND:**

- |  |                |
|--|----------------|
| <b>95 UCL Copper Concentration (mg/kg)</b> | STSIU Boundary |
| <1,100                                     | Local City     |
| 1,100 - 1,600                              | Stockpiles     |

*Note:*  
 1. Unshaded areas represent polygons with a maximum copper concentration <1,100 mg/kg and were not evaluated for UCLs; see Figure 3-3 for full set of alliance polygons.  
 2. Service Layer Credits: World Imagery: Earthstar Geographics



FREEPORT-MCMORAN CHINO MINES COMPANY VANADIUM, NEW MEXICO SMELTER/TAILINGS SOILS IU FS	
<b>95 UCL AREA-WEIGHTED                  AVERAGE COPPER CONCENTRATIONS                  ACROSS VEGETATION ALLIANCES</b>	
	FIGURE <b>3-5</b>



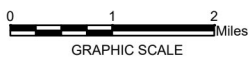
**LEGEND:**

**Iron Concentration (mg/kg)**

- <40,000
- 40,000 - 70,000
- 70,000 - 100,000
- >100,000

- Local Cities
- Railroad
- Major Roads

Service Layer Credits: World Imagery:  
 Earthstar Geographics

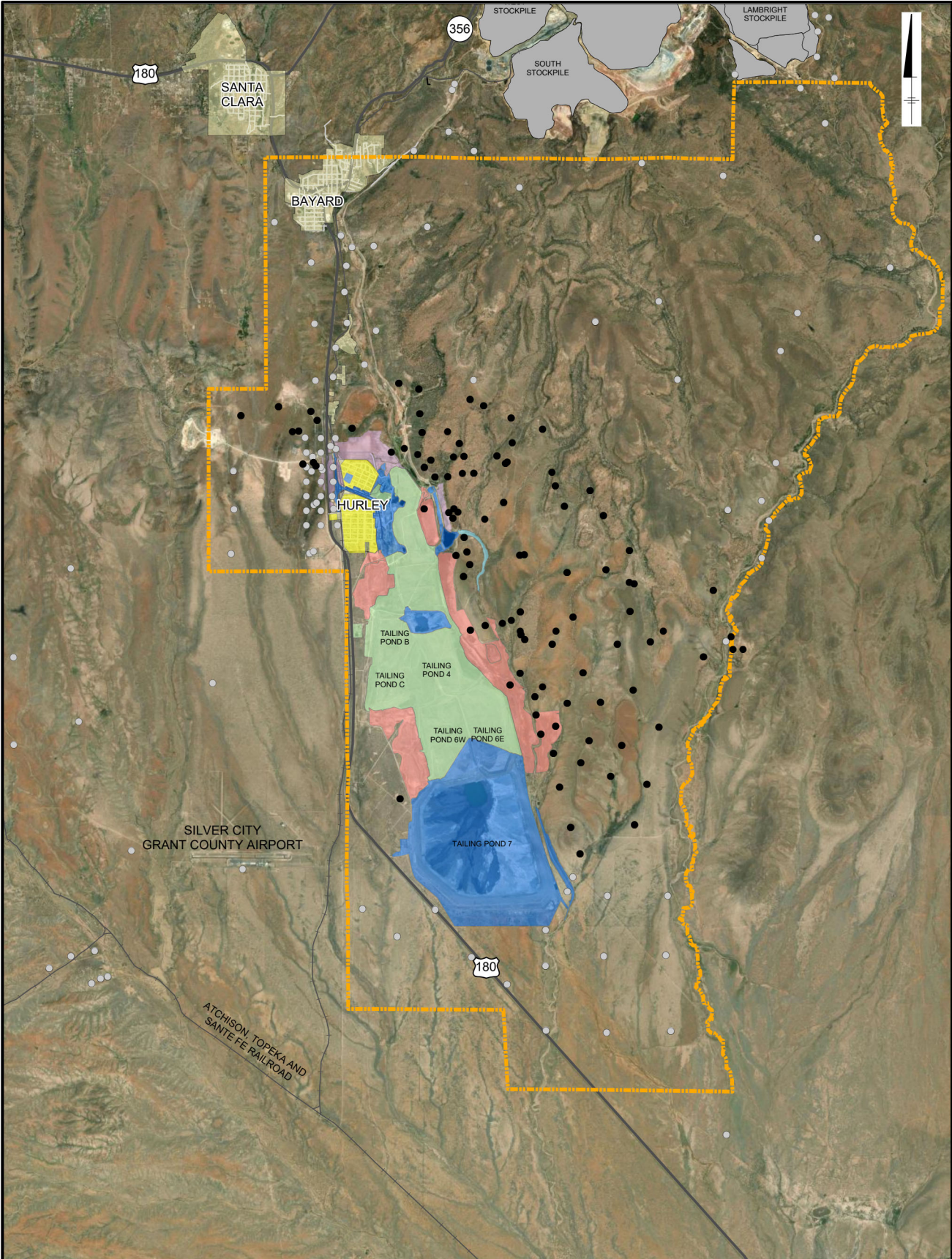


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 SMELTER/TAILINGS SOILS IU FS

**STSIU SAMPLE LOCATIONS AND  
 IRON CONCENTRATIONS  
 IN SOILS**



FIGURE  
**3-6**



**LEGEND:**

- Pre-White Rain Sample Location
- Post-White Rain Sample Location
- STSIU Boundary
- Stockpiles
- Local City
- Borrow Pit
- CCP
- Diversion Project
- Hurley Soils IU
- IRAs
- Ops Reclamation

Note:  
 1. Rangeland condition was determined using the observed apparent trend (OAT).

Service Layer Credits: World Imagery:  
 Earthstar Geographics

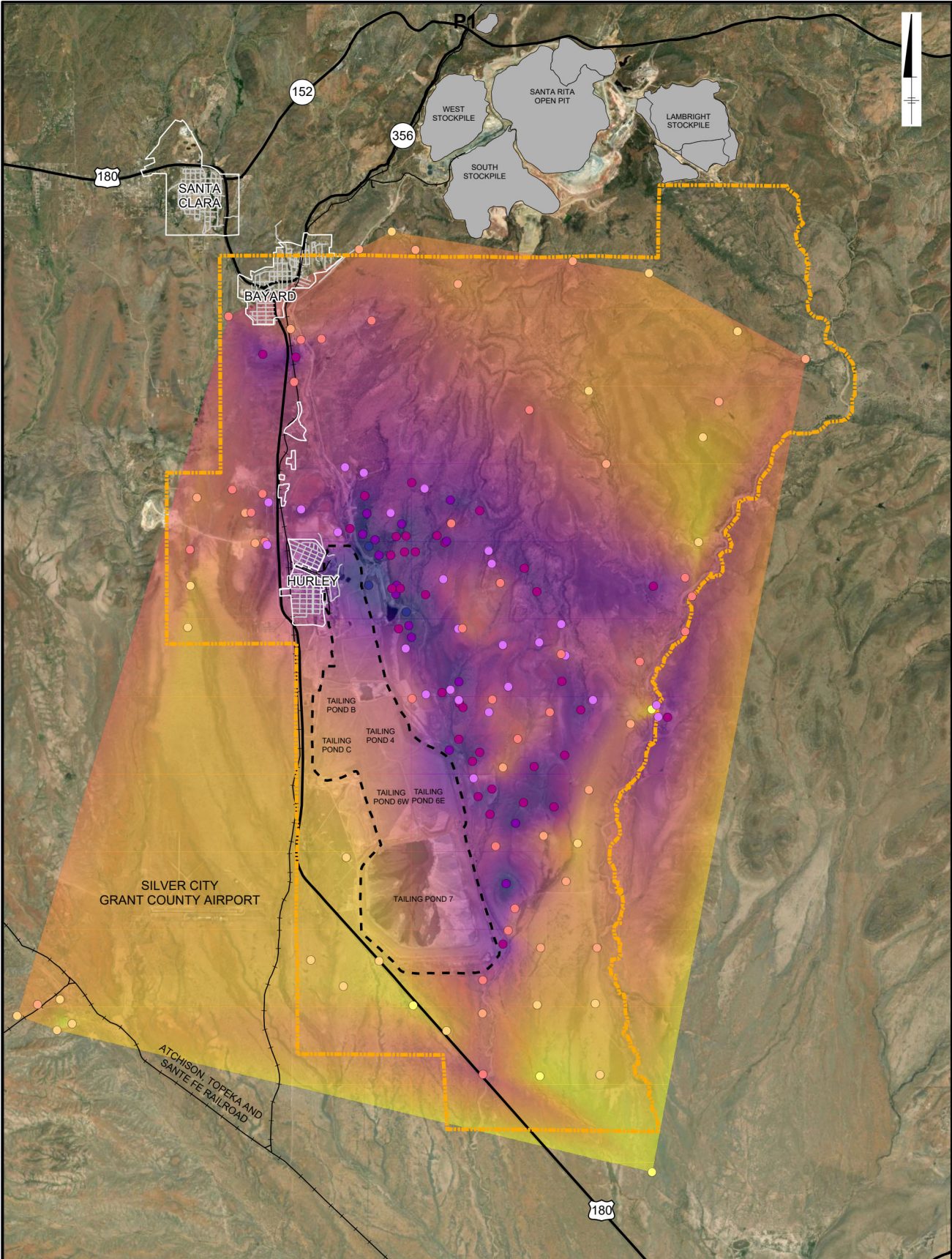


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 VANADIUM, NEW MEXICO  
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**PRE- AND POST-WHITE RAIN  
 pCu SAMPLE LOCATIONS**



FIGURE  
**3-7**

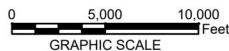


**LEGEND:**

- pCu Samples**
- >10
  - 9 - 10
  - 8 - 9
  - 7 - 8
  - 6 - 7
  - 5 - 6
  - 4 - 5
  - 3 - 4
  - <3

**Natural Neighbor Interpolation**

- 10.1
- 2.7
- Stockpiles
- STSIU Boundary
- Smelter Tailings Boundary



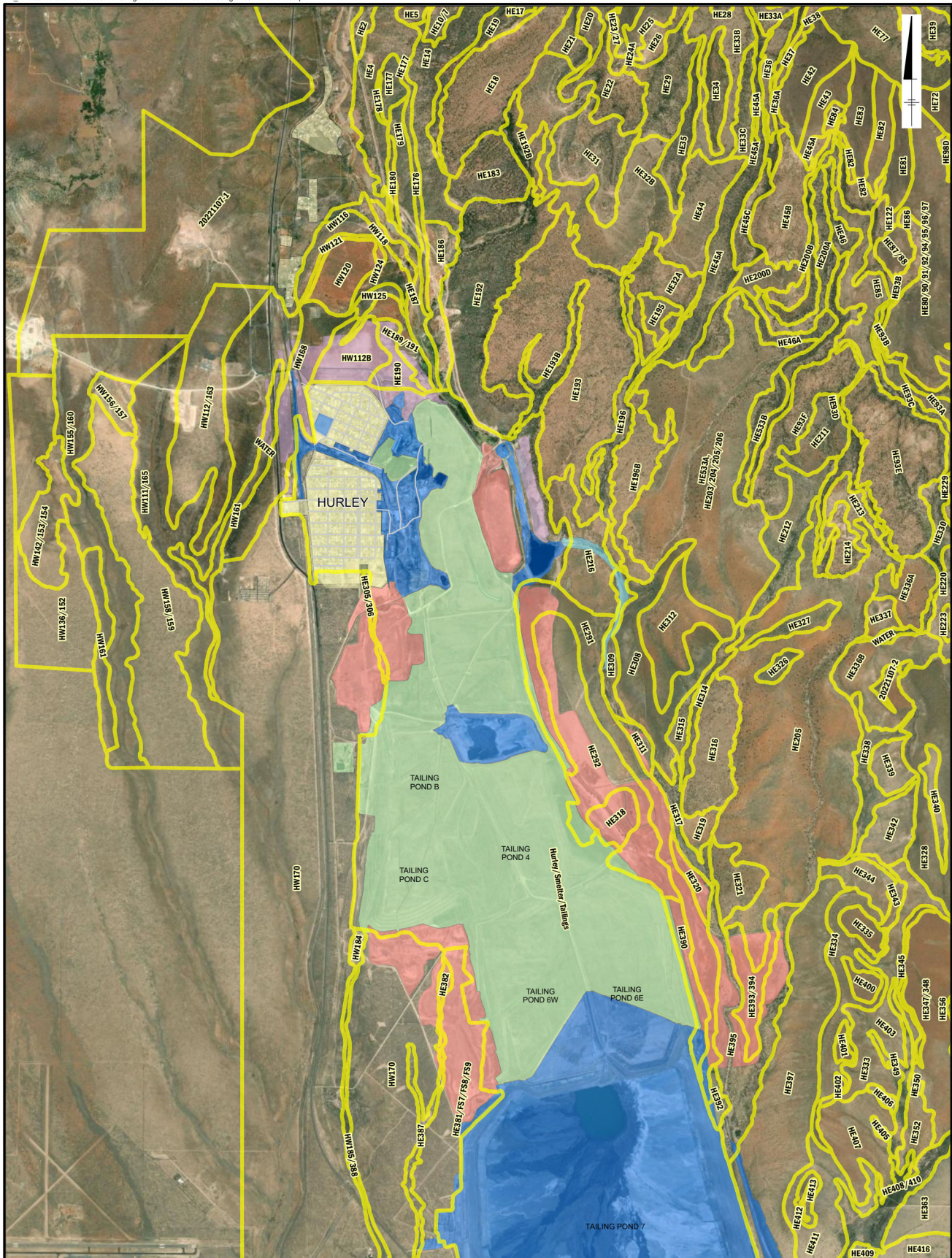
Service Layer Credits: World Imagery: Earthstar Geographics

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 VANADIUM, NEW MEXICO  
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**pCu SAMPLE RESULTS AND  
 NATURAL NEIGHBOR INTERPOLATION**

**ARCADIS**

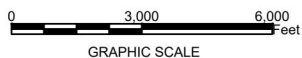
FIGURE  
**3-8**



**LEGEND:**

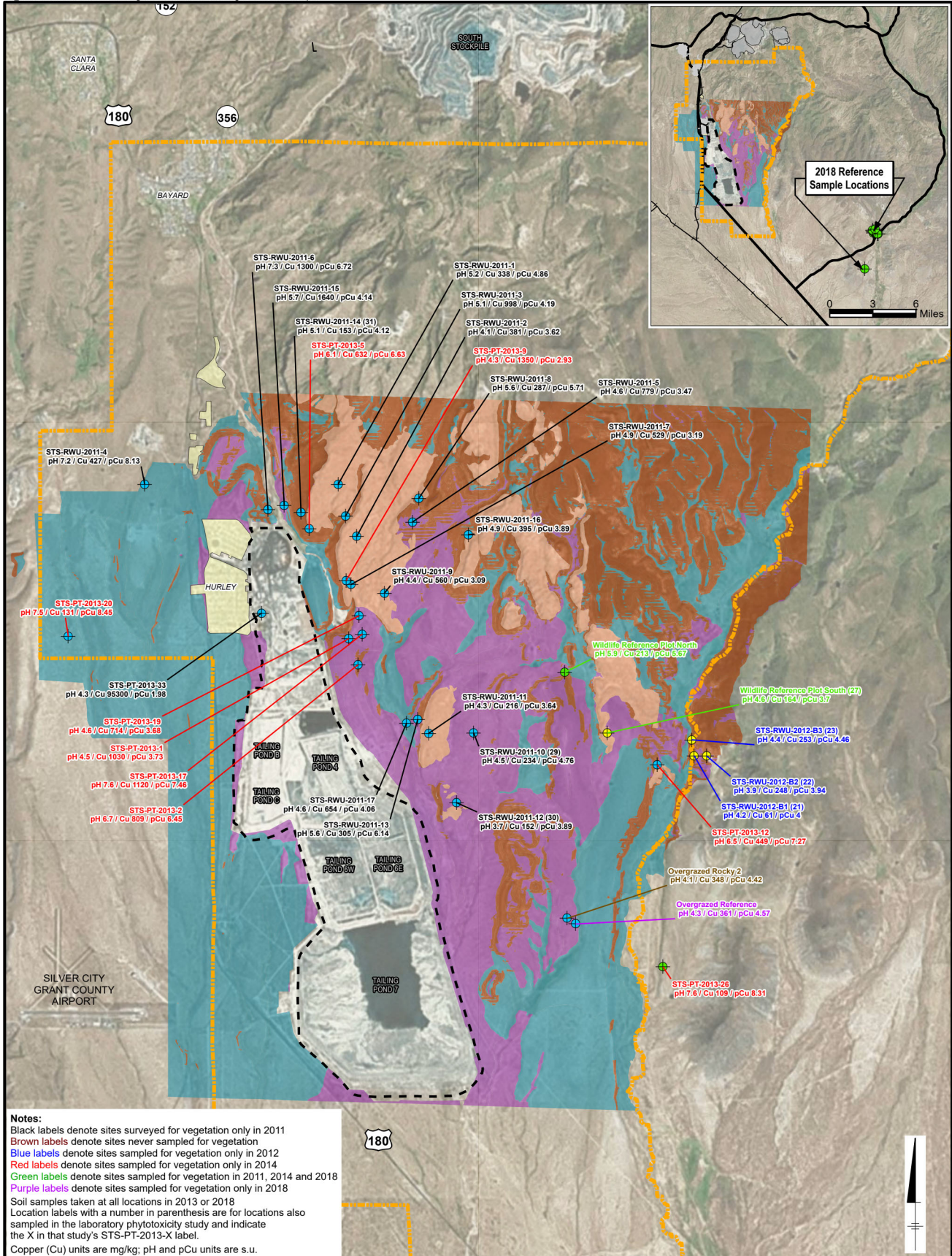
- |  |                   |
|--|-------------------|
| Rangeland Polygons (Exposure Unit for pCu) | Borrow Pit        |
| Town Roads                                 | CCP               |
| Stockpiles                                 | Diversion Project |
| Local City                                 | Hurley Soils IU   |
|  | IRAs              |
|  | Operations        |

Note:  
 1. Service Layer Credits: World Imagery: Maxar.  
 2. See abbreviations in report.



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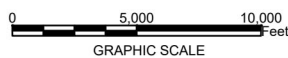
**RANGELAND POLYGONS**



**Notes:**  
 Black labels denote sites surveyed for vegetation only in 2011  
 Brown labels denote sites never sampled for vegetation  
 Blue labels denote sites sampled for vegetation only in 2012  
 Red labels denote sites sampled for vegetation only in 2014  
 Green labels denote sites sampled for vegetation in 2011, 2014 and 2018  
 Purple labels denote sites sampled for vegetation only in 2018  
 Soil samples taken at all locations in 2013 or 2018  
 Location labels with a number in parenthesis are for locations also sampled in the laboratory phytotoxicity study and indicate the X in that study's STS-PT-2013-X label.  
 Copper (Cu) units are mg/kg; pH and pCu units are s.u.

- LEGEND:**
- Site vegetation survey and soil sampling location
  - De minimus vegetation survey and soil sampling location
  - Reference vegetation survey and soil sampling location
  - STSIU Boundary
  - Smelter Tailings Boundary
- Soil Category**
- Flat Granular Soil
  - Flat Rocky Soil
  - Slope > 13%
  - Bedrock

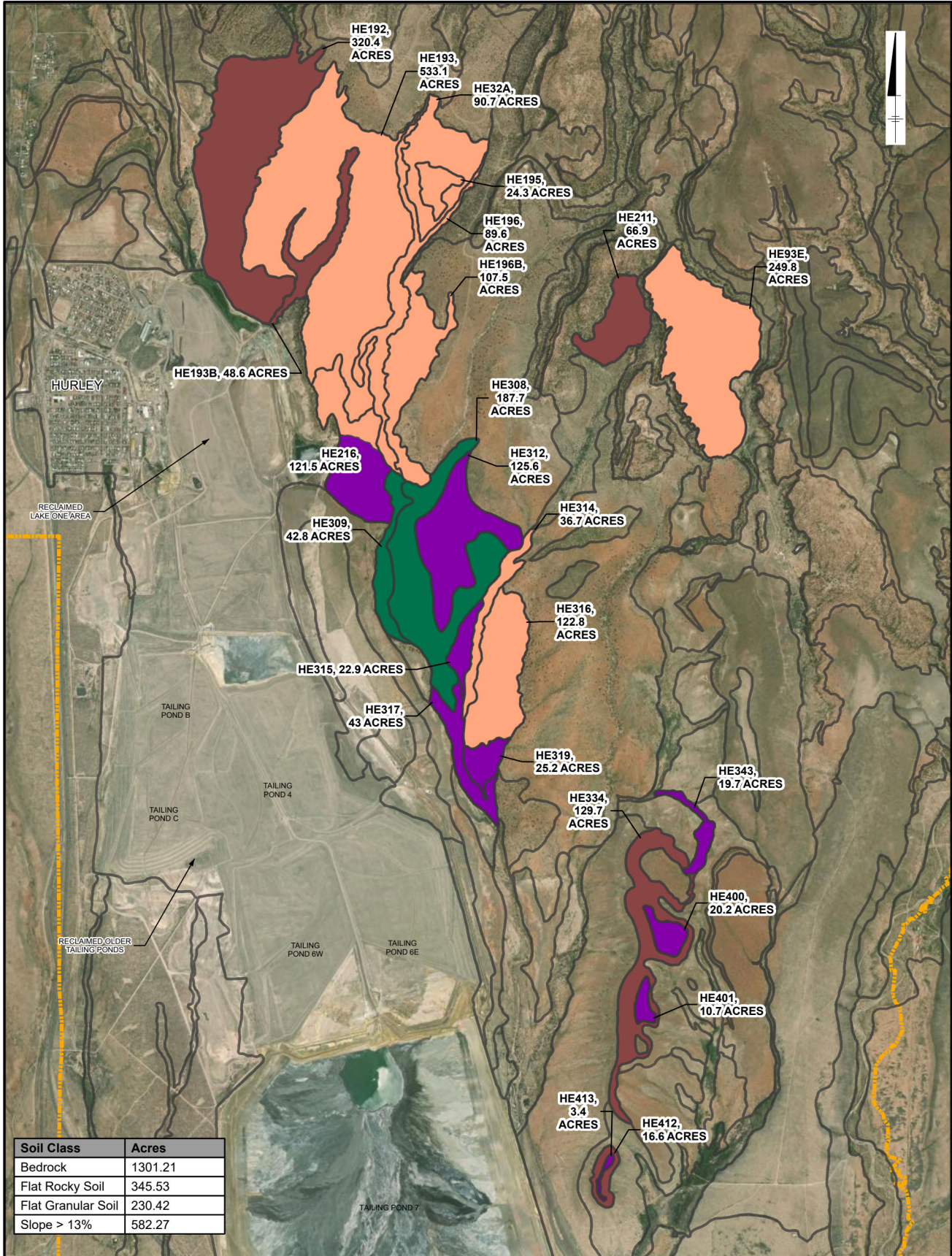
Service Layer Credits: World Imagery (Clarity). Source: Esri, Maxar, Earthstar Geographics, IGN, and the GIS User Community



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**VEGETATION SAMPLING LOCATIONS ON SOIL CATEGORY MAP**

**ARCADIS** | **FIGURE 3-10**



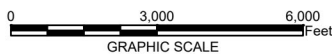
Soil Class	Acres
Bedrock	1301.21
Flat Rocky Soil	345.53
Flat Granular Soil	230.42
Slope > 13%	582.27

**LEGEND:**

**Soil Class**

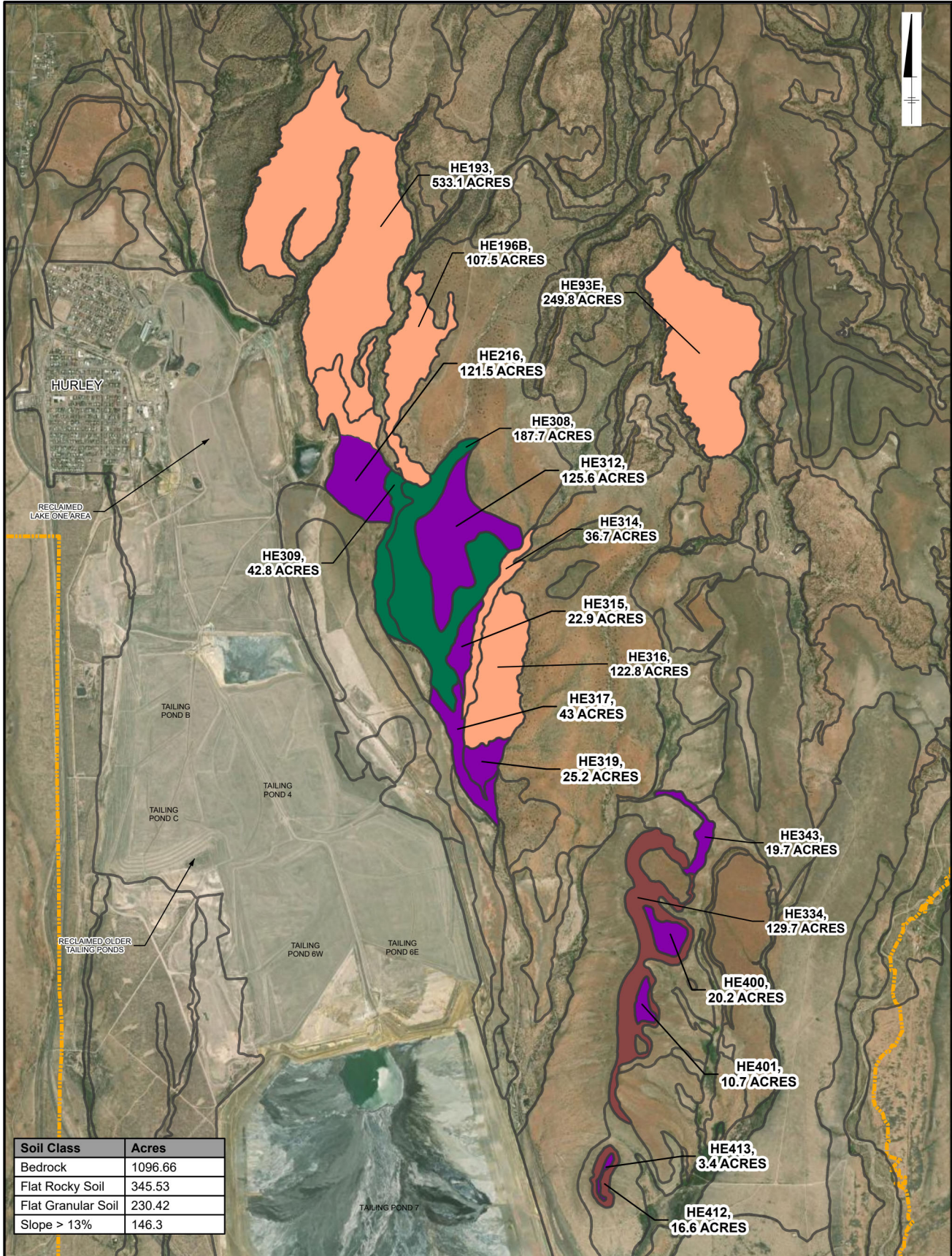
- Bedrock < 5.0 pCu & Cu > 327
- Flat Rocky Soil < 5.0 pCu & Cu > 327
- Slope greater than 13% < 5.0 pCu & Cu > 327
- Flat Granular Soil < 5.0 pCu & Cu > 327
- STSIU Boundary

Notes:  
 1. Service Layer Credits: World Imagery: Maxar  
 2. Copper threshold is greater than 327 mg/kg regardless of soil category.



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**RANGELAND POLYGONS WITH pCu  
 PRE-FS RAC BASED ON DATA POST-IRA**



Soil Class	Acres
Bedrock	1096.66
Flat Rocky Soil	345.53
Flat Granular Soil	230.42
Slope > 13%	146.3

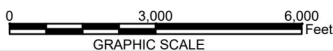
**LEGEND:**

**Soil Class**

- Bedrock < 5.0 pCu & Cu > 327
- Flat Rocky Soil < 5.0 pCu & Cu > 327
- Slope greater than 13% < 5.0 pCu & Cu > 327
- Flat Granular Soil < 5.0 pCu & Cu > 327
- STSIU Boundary

**Notes:**

1. Service Layer Credits: World Imagery; Earthstar Geographics
2. Areas unacceptable have an unacceptable rangeland condition (unacceptable OAT score) or unacceptable wildlife habitat (e.g., unacceptable richness or unacceptable cover is considered unacceptable wildlife habitat).
3. Copper threshold is greater than 327 mg/kg regardless of soil category.



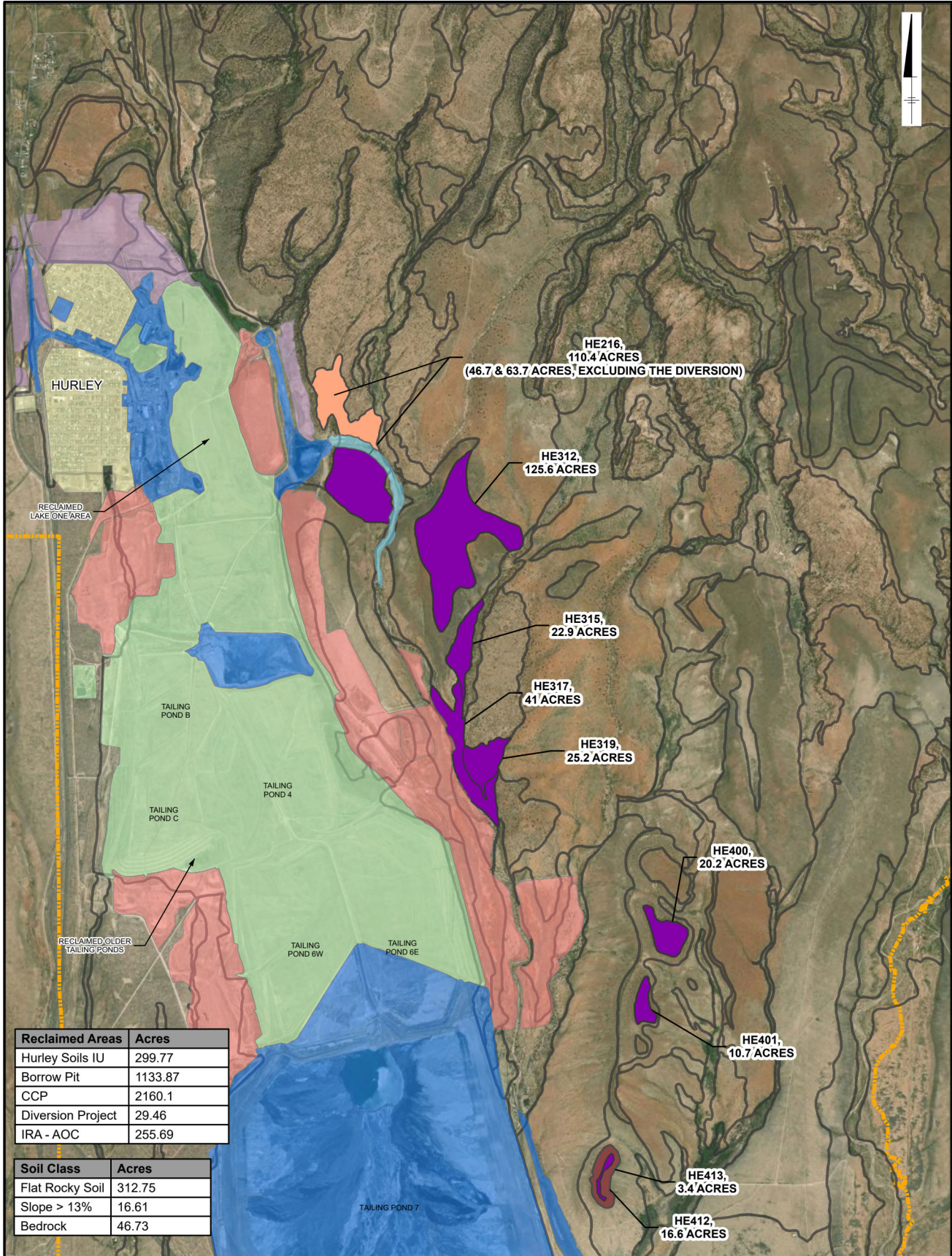
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 SMELTER/TAILINGS SOILS IU FS

**RANGELAND POLYGONS WITH pCu  
 PRE-FS RAC ON DATA POST-IRA USING  
 UNACCEPTABLE RANGELAND OR  
 WILDLIFE HABITAT CRITERIA**



FIGURE  
**3-12**





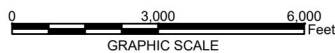
Reclaimed Areas	Acres
Hurley Soils IU	299.77
Borrow Pit	1133.87
CCP	2160.1
Diversion Project	29.46
IRA - AOC	255.69

Soil Class	Acres
Flat Rocky Soil	312.75
Slope > 13%	16.61
Bedrock	46.73

**LEGEND:**

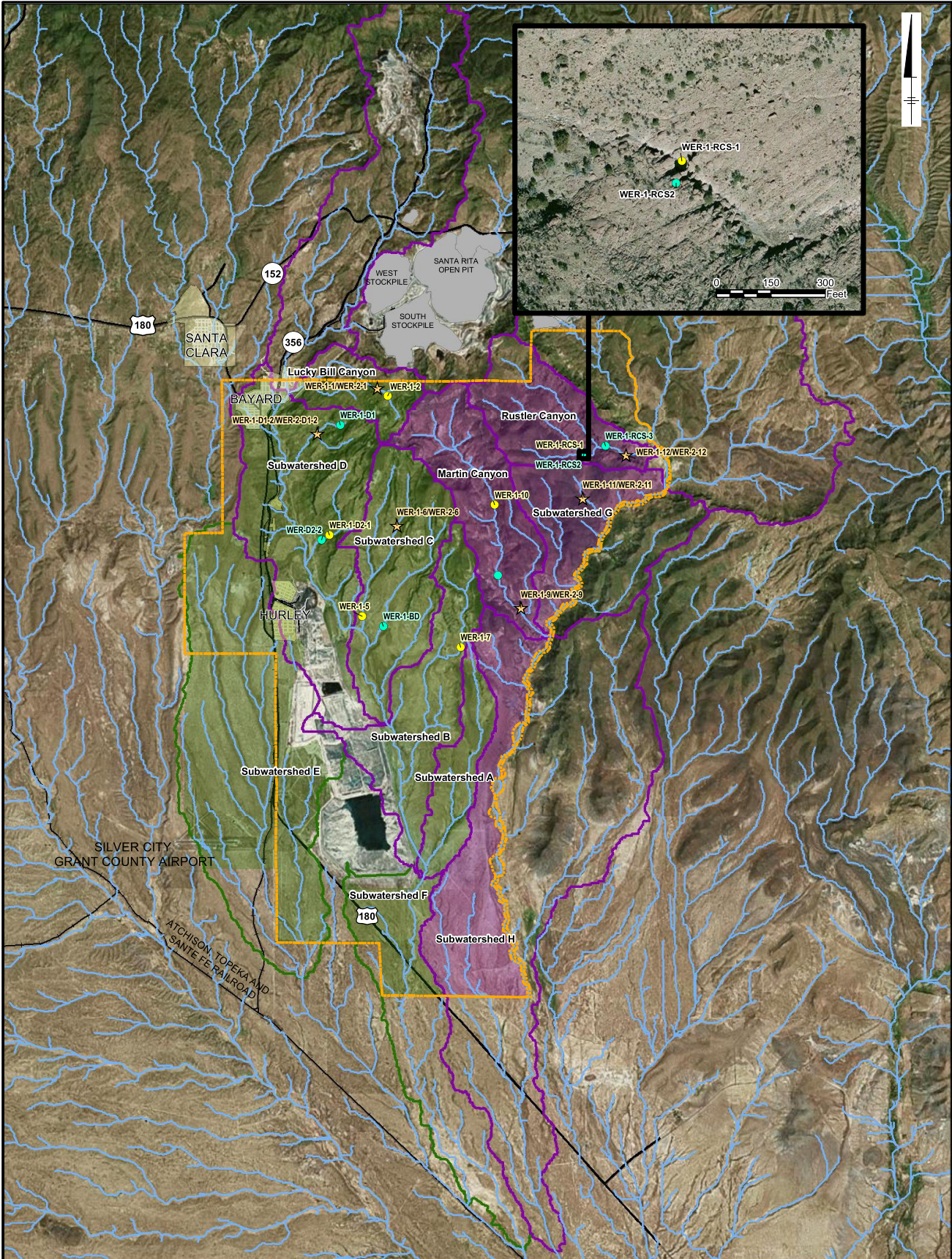
- Soil Class**
- Bedrock < 4.4 pCu & Cu > 327
  - Flat Rocky Soil < 4.98 pCu & Cu > 327
  - Slope greater than 13% < 4.11 pCu & Cu > 327
  - STSIU Boundary
  - Borrow Pit
  - CCP
  - Diversion Project
  - Hurley Soils IU
  - IRAs
  - Operations

- Notes:**
1. Service Layer Credits: World Imagery; Maxar
  2. Areas unacceptable have an unacceptable rangeland condition (unacceptable OAT score) or unacceptable wildlife habitat (e.g., unacceptable richness or unacceptable cover is considered unacceptable wildlife habitat). Copper threshold is greater than 327 mg/kg regardless of soil category.
  3. Flat Granular Soils have a PEL Criteria of 4.11, these rangeland units are no longer considered for remediation.



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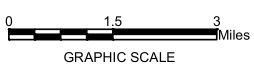
**RANGELAND POLYGONS WITH pCu  
 PRE-FS RAC ON DATA POST-IRA USING  
 UNACCEPTABLE AND PEL CRITERIA**



**LEGEND:**

- Analytical Sample Only
- Tox and Analytical Sample
- ★ Tox and Analytical Sample (Sampled Twice)
- Lampbright Subwatershed Boundaries
- Lampbright Subwatersheds within AOC
- Hanover-Whitwater Subwatershed Boundaries
- Hanover-Whitwater Subwatersheds within AOC
- Stockpiles
- STSIU Study Boundary
- Local City
- Highway
- Railroad
- Town Roads
- Drainages

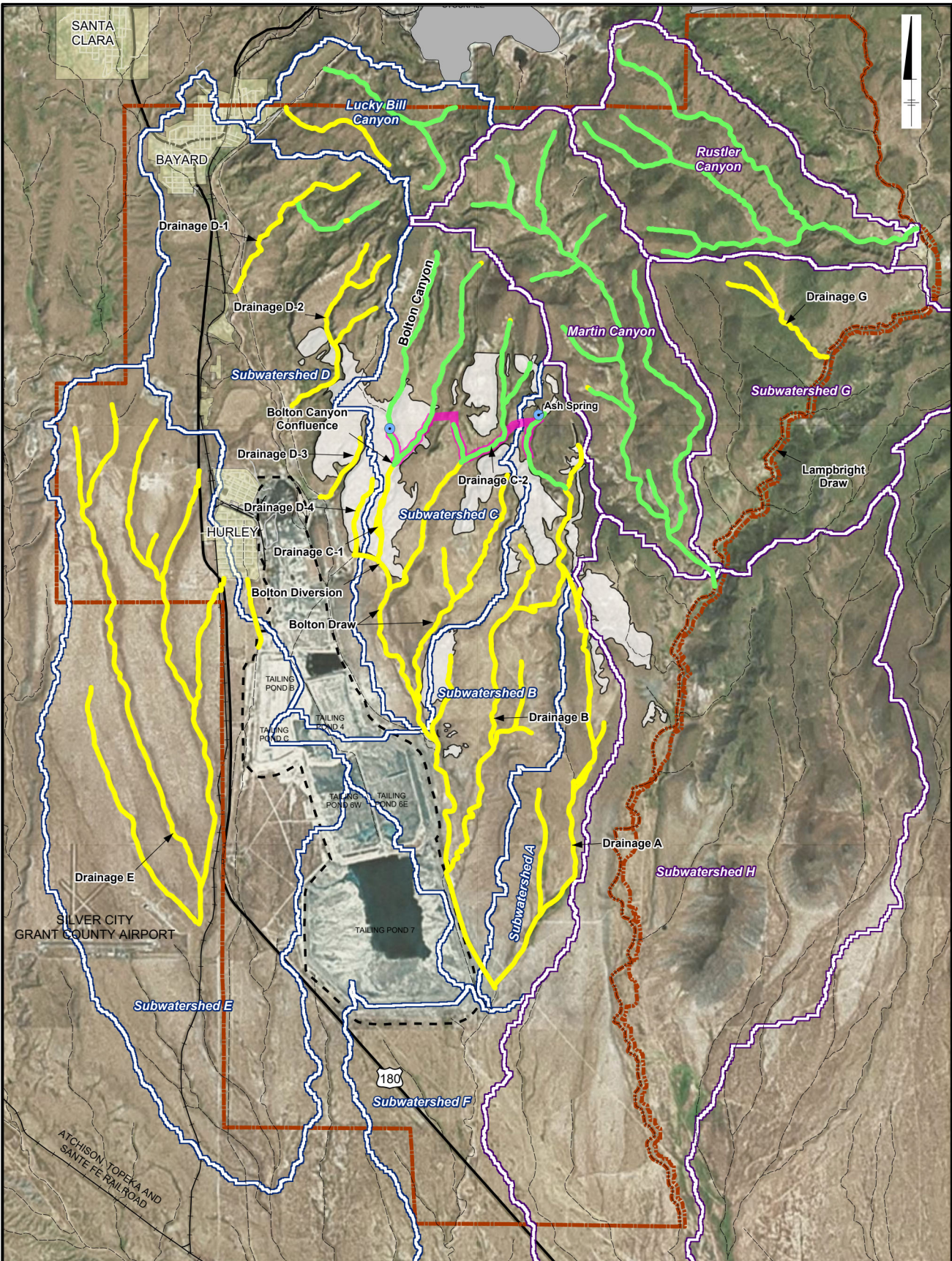
*Service Layer Credits: Source:  
 Esri, DigitalGlobe, GeoEye,  
 Earthstar Geographics,  
 CNES/Airbus DS, USDA, USGS,  
 AeroGRID, IGN, and the GIS  
 User Community*



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**SMELTER/TAILINGS SOILS IU FS**

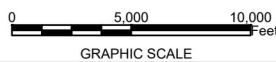
**WER SAMPLE LOCATIONS**

**ARCADIS** | **FIGURE 4-1**



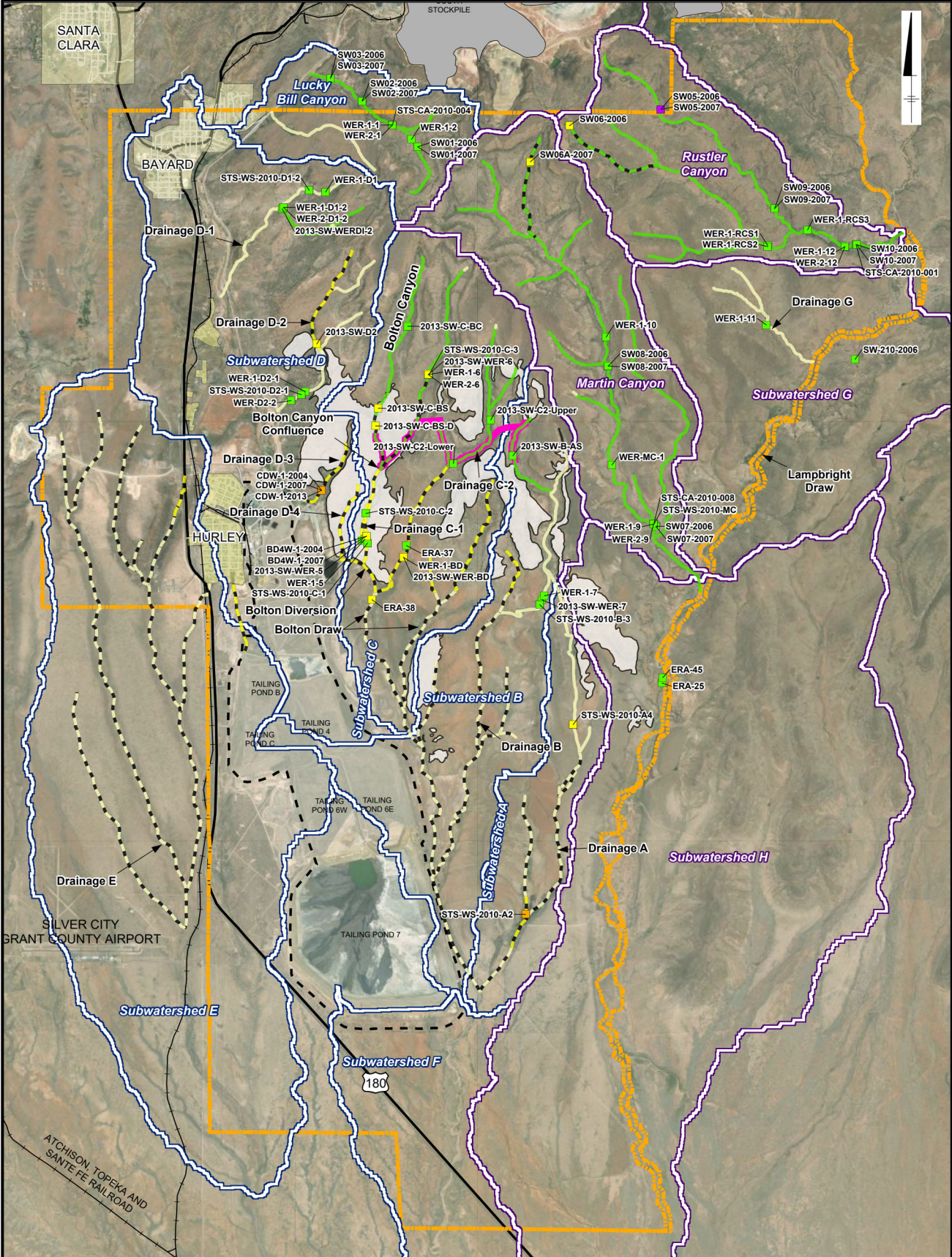
**LEGEND:**

- |  |                        |                      |
|--|------------------------|----------------------|
| Springs                                    | STSIU Boundary         | Non-Ephemeral Stream |
| Lampbright Subwatershed Boundaries         | Bedrock Extents        | Smelter Tailings     |
| Hanover-Whitewater Subwatershed Boundaries | USFWS Critical Habitat | Boundary             |
| Hanover-Whitewater Subwatershed Boundaries | Ephemeral Stream       | Local City           |
|  |                        | Stockpiles           |



FREEPORT-MCMORAN CHINO MINES COMPANY  
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**HYDROLOGIC CLASSIFICATIONS  
 OF STSIU DRAINAGES**



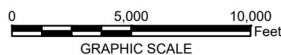
**LEGEND:**

Highest HQ Value

- < 1
- 1 - 3
- 3 - 7
- > 7
- USFWS Critical Habitat Transect
- Lampbright Subwatershed Boundaries
- Hanover-Whitewater Subwatershed Boundaries
- STSIU Boundary

- Non-Ephemeral/No Concerns
- Non-Ephemeral/Potential Concerns
- Ephemeral/No Concerns
- Ephemeral/Potential Concerns
- Ephemeral/Potential Concerns but Limited Potential for Pools

Notes:  
 1. Service Layer Credits: World Imagery; Earthstar Geographics  
 2. Multiple samples may have been collected at the locations, and to reference the Section 4 tables for results by date.

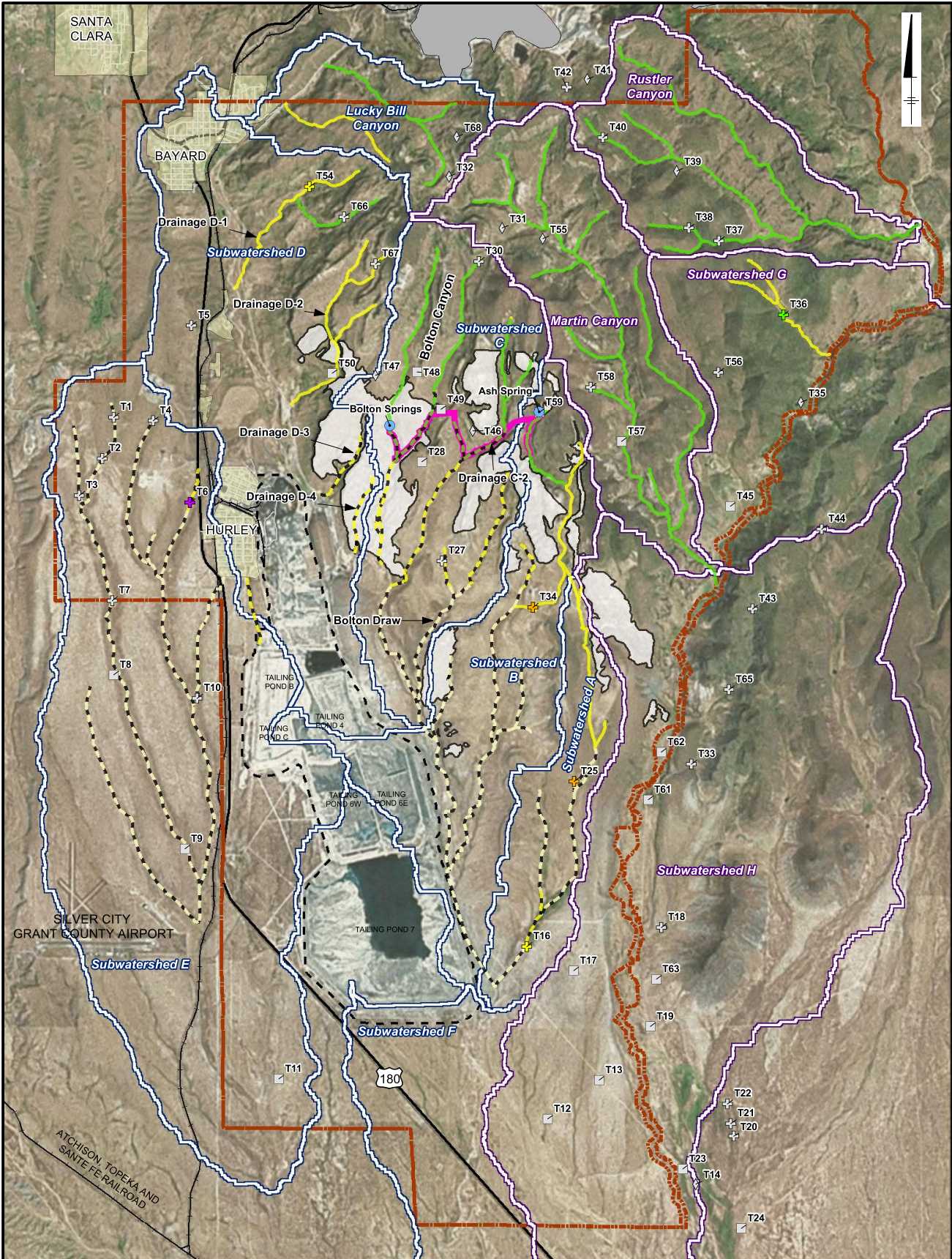


FREEPORT-MCMORAN CHINO MINES COMPANY  
 VANADIUM, NEW MEXICO  
 SMELTER/TAILINGS SOILS IU FS

**NATURE AND EXTENT OF COPPER CONTAMINATION TO SURFACE WATER DRAINAGES**

**ARCADIS**

FIGURE 4-3



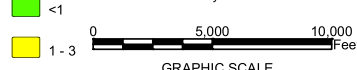
**LEGEND:**

- Springs
- USFWS Critical Habitat Transect
- STSIU Boundary
- Bedrock Extents
- Lambright Subwatershed Boundaries
- Hanover-Whitewater Subwatershed Boundaries
- Non-Ephemeral/Confident No Concerns
- Non-Ephemeral/Potential Active Remediation or Monitored Natural Attenuation
- Ephemeral/Confident No Concerns Based on Proposed WER Model
- Ephemeral/Potential Active Remediation or Monitored Natural Attenuation
- Ephemeral/Limited Potential for Pools

- + Earthen Impoundments in the Drainage
- ◇ Earthen Impoundments Outside the Drainage
- Concrete or Steel Tank

- Highest HQ Value**
- No Data
  - <1
  - 1 - 3
  - 3 - 7
  - >7

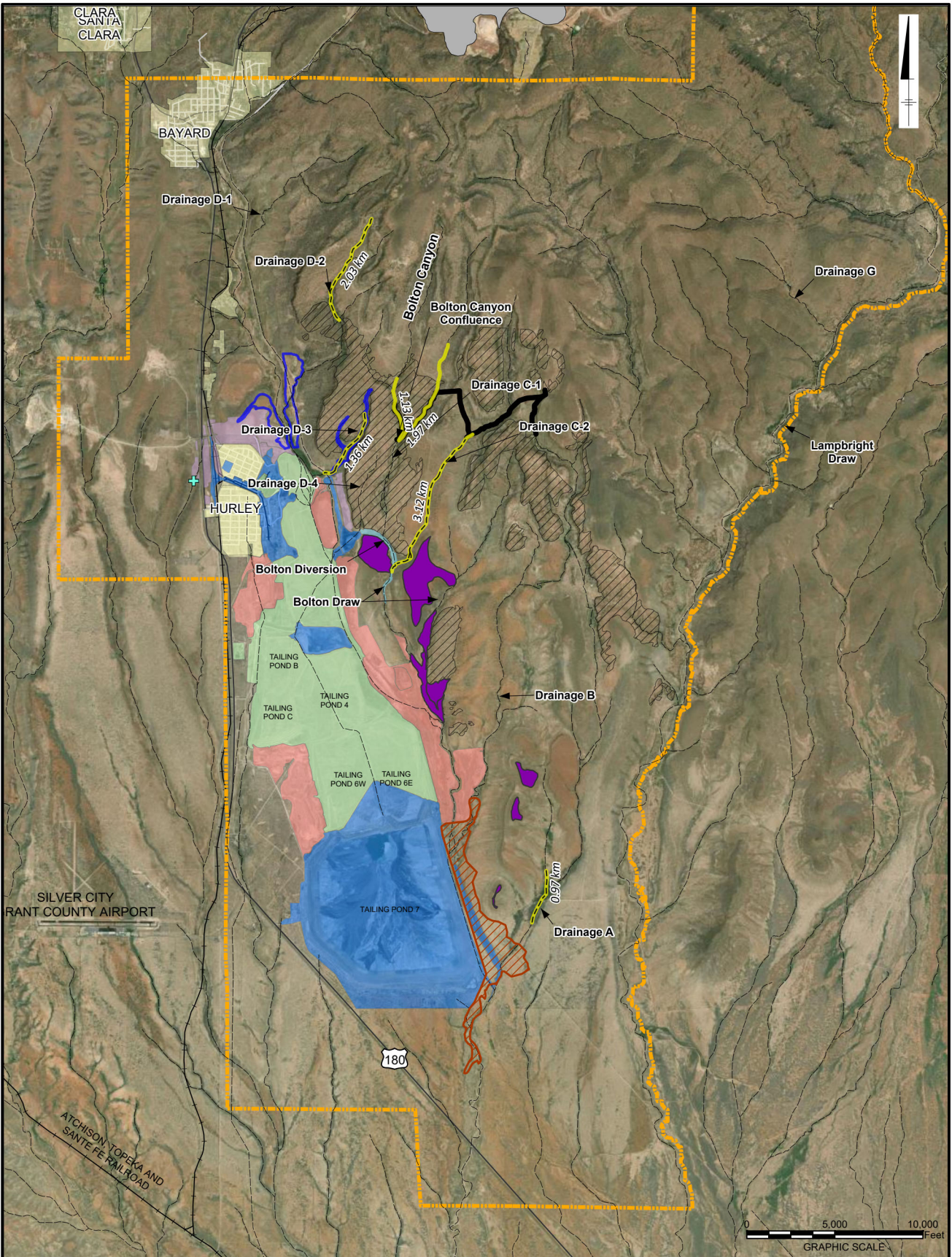
Service Layer Credits: Source: Esri, DigitalGlobe, GeoEye, Earthstar Geographics, CNES/Airbus DS, USDA, USGS, AeroGRID, IGN, and the GIS User Community



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**STSIU STOCK TANK LOCATIONS**

**ARCADIS** | **FIGURE 4-4**



**LEGEND:**

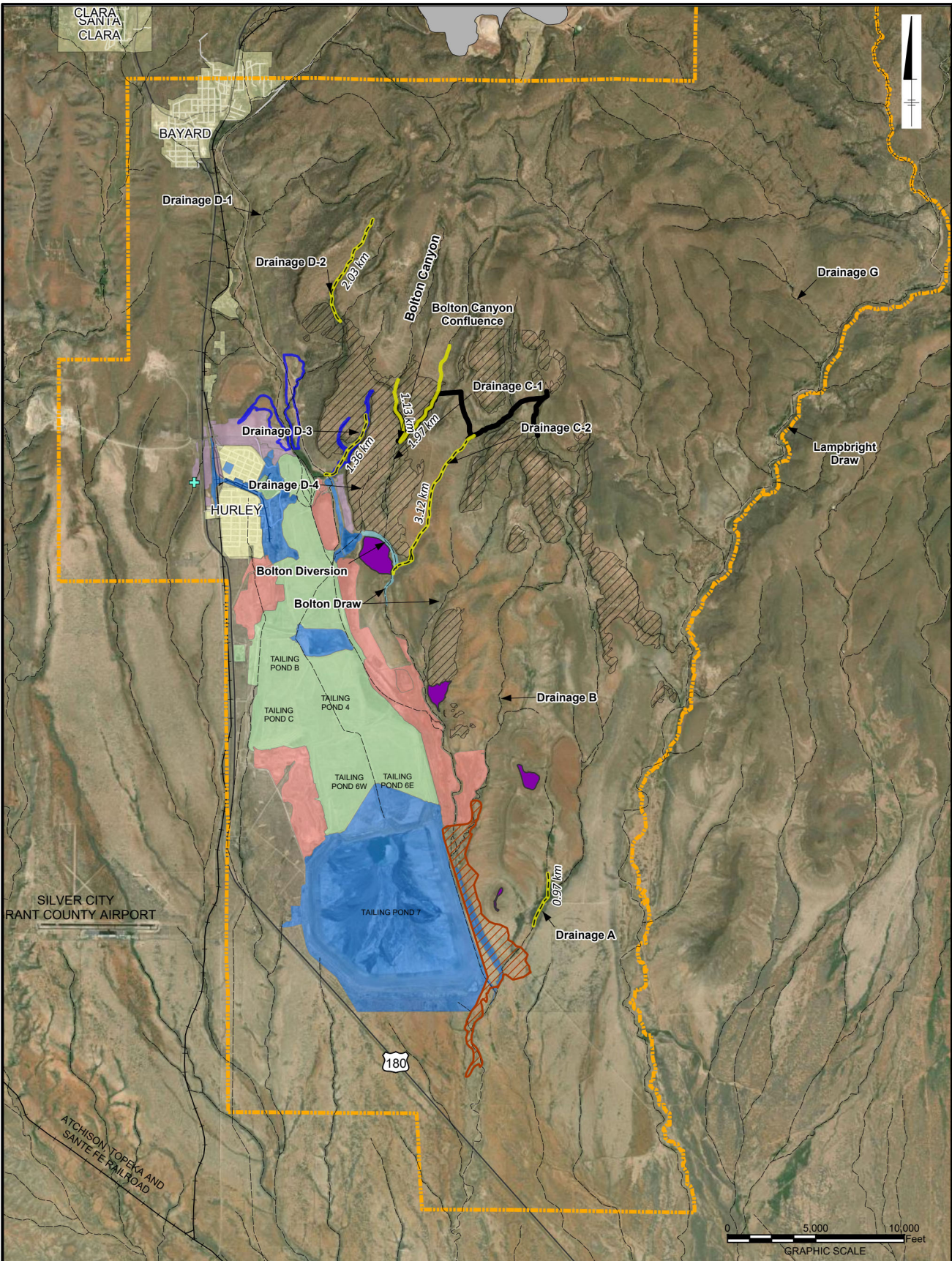
- |                                       |  |                           |            |
|---------------------------------------|--|---------------------------|------------|
| Stock Tank Proposed for Monitoring    | <b>pCu-driven Action</b>   | Avian Monitoring Boundary | IRAs       |
| <b>Stream Type/Condition</b>          | Flat Rocky Soil < 4.98 pCu & Cu > 327                            | Stockpiles                | Operations |
| Ephemeral/Proposed for Monitoring     | Bedrock Extents  | Local City                |            |
| Non-Ephemeral/Proposed for Monitoring | Recent Windblown Tailings from TP-7 Outside Operational Boundary | Borrow Pit                |            |
| USFWS Critical Habitat Transect       | STSIU Boundary   | CCP                       |            |
|                                       |  | Diversion Project         |            |
|                                       |  | Hurley Soils IU           |            |

**Notes:**

1. Service Layer Credits: World Imagery: Earthstar Geographics.
2. While this figure portrays a scenario where PEL Criteria is 4.98, the final acres are based on a PEL of 4.6 (and 140 acres are posed for monitoring).
3. Southern edge of blue polygon includes recent windblown tailings from TP-7 inside the operational boundary (see Figure 3-2a).

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**AREAS IDENTIFIED FOR  
 REMEDY OR MONITORING**



**LEGEND:**

- |                                       |  |                           |            |
|---------------------------------------|--|---------------------------|------------|
| Stock Tank Proposed for Monitoring    | <b>pCu-driven Action</b>   | Avian Monitoring Boundary | IRAs       |
| <b>Stream Type/Condition</b>          | Flat Rocky Soil < 4.6 pCu & Cu > 327                             | Stockpiles                | Operations |
| Ephemeral/Proposed for Monitoring     | Bedrock Extents  | Local City                |            |
| Non-Ephemeral/Proposed for Monitoring | Recent Windblown Tailings from TP-7 Outside Operational Boundary | Borrow Pit                |            |
| USFWS Critical Habitat Transect       | STSIU Boundary   | CCP                       |            |
|                                       |  | Diversion Project         |            |
|                                       |  | Hurley Soils IU           |            |

Notes:  
 1. Service Layer Credits: World Imagery: Earthstar Geographics.  
 2. Southern edge of blue polygon includes recent windblown tailings from TP-7 inside the operational boundary (see Figure 3-2a).

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**PREFERRED AREAS IDENTIFIED FOR  
 REMEDY OR MONITORING**